*Exxon Valdez* Oil Spill State/Federal Natural Resource Damage Assessment Final Report

Effects of Pink Salmon (*Oncorhynchus gorbuscha*) Escapement Level on Egg Retention, Preemergent Fry, and Adult Returns to the Kodiak and Chignik Management Areas Caused by the *Exxon Valdez* Oil Spill

> Fish/Shellfish Study Numbers 7B and 8B Final Report

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**Study History:** Fish/Shellfish Study Numbers 7B and 8B were initiated under a 1989 detailed study plan as Fish/Shellfish Studies 7 and 8 (Injury to Pink Salmon Spawning Areas and Pink Salmon eggs and Fry in Areas Outside of Prince William Sound). In October 1990 the study was stratified owing to disparate objectives and concerns into streams within Lower Cook Inlet (F/S Study Numbers 7A and 8A), and streams within the Kodiak-Chignik areas (F/S Study Numbers 7B and 8B). Injury assessment was continued with field studies terminating after the 1991 field season. a draft report was submitted by C.O. Swanton entitled Effects of Pink Salmon (*O. gorbuscha*) Escapement Level on Egg Retention, Preemergent Fry, and Adult Returns to the Kodiak and Chignik Management Areas Caused by the *Exxon Valdez* Oil Spill. No restoration efforts were required for pink salmon populations to either of the affected areas.

Abstract: Potential impacts of overescapement on several life history stages of pink salmon from streams located within the Kodiak and Chignik commercial salmon fishing areas were studied. The 1989 pink salmon escapement for Kodiak was 21.0 million (odd-year escapement goal 4.7 million) and for Chignik 1.4 million fish (odd-year escapement goal 0.7 million). Measurements of egg retention, fecundity, stream residence time (stream life), total available spawning habitat, and preemergent fry densities were obtained. Egg retention was found to be positively related to spawner density; observed 1990 preemergent fry densities were significantly below predicted values for 23 Kodiak and 7 Chignik streams indicating reduced spawner success for some streams. Return per spawner analyses for Kodiak showed a significant density dependent response, however no such result was found for the Chignik data. Overall no conclusive evidence of reduced production of pink salmon adults from the 1989 escapement event was found.

Key Words: Adult returns, Chignik, egg retention, Exxon Valdez, Kodiak, Oncorhynchus gorbuscha, overescapement, pink salmon, preemergent fry.

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## EXECUTIVE SUMMARY

As a result of the 1989 *Exxon Valdez* oil spill, commercial salmon fishing in and around the Kodiak and Chignik areas was severely restricted throughout the 1989 season. Consequently, pink salmon escapements for these areas greatly exceeded targeted escapement objectives. Investigations were conducted within the Kodiak and Chignik Management Areas during 1989 and 1990 to determine if negative impacts on future odd-year brood line pink salmon production occurred as a result of overescapement in 1989.

The 1989 pink salmon escapements for the Kodiak and Chignik management areas were estimated to be 21.0 and 1.4 million fish, respectively; odd-year escapement objectives are 4.7 million (Kodiak) and 0.7 million (Chignik). Egg retention of spawners was found to be positively related to spawner density. Observed 1990 preemergent fry densities were significantly below those predicted from linear regression models in 23 Kodiak and 7 Chignik index streams; depressed fry production was found in 18 Kodiak and Chignik streams, collectively. Conversely, there were also Kodiak and Chignik preemergent fry index streams that had observed 1990 fry density values which were greater than average, suggesting no density dependent effect for these streams pink salmon populations. Return per spawner analyses for Kodiak resulted in a significant density dependent parameter estimate, while for Chignik no such result was found. A possible mechanism for the density dependence found within the return per spawner analysis is the effect of spawner density on egg retention with a subsequent reduction in preemergent fry yield for some Kodiak area streams.

For Kodiak, the 1991 pink salmon return was 22.0 million; predicted returns using midpoint and desired odd-year escapement goals of 4.7 and 7.0 million were 11.9 (range 3.6 to 24.4 million) and 15.7 (range 5.8 to 29.5 million) million, respectively. Had Kodiak area pink salmon escapement midpoint or desired goals been met the 1991 return could easily have been higher or lower than the actual return of 22.0 million. The Chignik area experienced a 1991 return of 1.9 million; the predicted return using the escapement goal (0.7 million) was 1.3 million.

Even though spawner density had an affect on egg retention and preemergent fry, we do not believe that a measurable impact occurred on either Kodiak or Chignik area pink salmon returns from the 1989 escapement event.

## INTRODUCTION

During 1989, numerous salmon harvest opportunities within the Kodiak (KMA) and Chignik Management Areas (CMA) were foregone owing to the *Exxon Valdez* Oil Spill. Pink salmon (*Oncorhynchus gorbuscha*) runs occur in 454 Kodiak and 107 Chignik streams and comprise 78% and 31% (1978-1988), respectively, of the Kodiak and Chignik annual salmon harvests (Malloy and Prokopowich 1992; Thompson and Owen 1992). The ex-vessel value of this harvest has averaged (1978-1988), in millions of dollars, 14.2 and 1.5 for the Kodiak and Chignik areas. The KMA odd-year midpoint pink salmon escapement goal is 4.67 million, whereas the estimated total escapement for 1989 was about 21.0 million (Barrett et al. 1990). No annual escapement within the last three decades (1963-1991) has approached this level. The CMA desired pink salmon escapement goal is 0.70 million (Probasco et al. 1987); estimated total escapement during 1989 was 1.4 million (Barrett 1990).

In addition to monetary losses, the unrealized harvests of pink salmon could potentially have resulted in overescapement of spawners with consequent depression in returns from the 1989 brood year. Redd superimposition resulting from high spawner densities can be an important cause of mortality in Pacific salmon (Gilbert and Rich 1927; Smirnov 1947; Morgan and Henry 1959). McNeil (1964) demonstrated that pink salmon egg mortality during spawning was directly related to the density of females on spawning beds in two Southeastern Alaska streams. McNeil (1964) and Beverton and Holt (1957) both surmised that the production of fry within spawning beds is limited by maximum fry yields owing to density dependent mortality of eggs and alevins. Heard (1978) presented compelling evidence that this occurred in a Southeastern Alaska stream that experienced an extremely high pink salmon escapement in 1967.

The KMA encompasses the entire Kodiak archipelago and that portion of the Alaska Peninsula draining into Shelikof Strait from Cape Douglas to Kilokak Rocks bordering Imuya Bay (Figure 1). The archipelago and Alaska Peninsula portions of the management area are each about 241 km in length while Shelikof Strait which separates the two, averages approximately 48 km in width. The commercial salmon fishery occurs within seven districts which enclose about 454 pink salmon spawning streams. Fishery managers employ aerial and foot survey escapement counts into 51 index streams as part of the inseason fishery management program; 43 index streams have preemergent fry data collected for generating preseason run forecasts (Figure 2).

The CMA includes all coastal waters and inland drainages of the northwest Gulf of Alaska extending from Kilokak Rocks bordering Imuya Bay to Kupreanof Point (Figure 3). There are five commercial salmon fishing districts which contain 107 pink salmon spawning streams. Within the CMA are 31 aerial survey index streams, 18 of which have had preemergent fry sampling conducted (Figure 4).

We employed approaches spanning several life history stages for examining whether 1989 foregone harvests led to depression in spawning success, fry yields, and the subsequent returns of pink salmon. Egg retention, fecundity, stream residence time (stream life) and estimated total escapement, total available spawning habitat, and preemergent fry data were collected during 1989 and 1990. Goals of this study were:



Figure 1. Map of the Kodiak Management Area and associated commercial salmon fishing districts.



Figure 2. Approximate locations of Kodiak Management Area aerial escapement survey, preemergent fry, egg retention, stream life, and fecundity sampling streams.



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Figure 3. Map of the Chignik Management Area showing commercial salmon fishing districts.



Figure 4. Chignik Management Area aerial escapement survey, preemergent fry, egg retention, and fecundity sampling index streams.

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- estimate the numbers of spawning pink salmon by "index" stream for systems outside Prince William Sound where historical fry density data exist. These include 43+ streams in the Kodiak Island/Shelikof Strait mainland area and 18 streams in the Chignik area;
- (2) produce a catalog of aerial photographs and detailed maps of pink salmon spawner distribution for index streams in the Kodiak and Chignik areas;
- (3) determine abundance of pink salmon eggs and preemergent fry. Inclusive of deriving a fecundity-length relationship for selected odd-year Kodiak and Chignik populations; estimating egg retention for populations utilizing selected preemergent fry index streams; and determine total available spawning habitat for these streams;
- (4) estimate or derive indicies of overwinter mortality (egg to preemergent fry) of pink salmon eggs;
- (5) determine reductions, if any, in pink salmon returns from the 1989 escapement event;
- (6) identify potential alternative methods and strategies for restoration of lost use, populations, or habitat where injury is identified.

#### **OBJECTIVES**

- estimate the numbers of spawning pink salmon by "index" stream for systems outside Prince William Sound where historical fry density data exist. These include 43+ streams in the Kodiak Island/Shelikof Strait mainland area and 18 streams in the Chignik area;
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- (6) identify potential alternative methods and strategies for restoration of lost use, populations, or habitat where injury is identified.

#### METHODS

#### Stream Life

Stream life for this study is defined as the time span from when an individual pink salmon enters freshwater (counted through a weir or is tagged) until the individual dies (Bocking et al. 1988). It is represented as an average over all fish within a population. During 1989, fish counting weirs were installed on the Akalura, E. Paramanof, and Litnik systems to obtain daily and total escapement counts by species (Figure 2). On a minimum of two and maximum of three day rotational basis, stream foot surveys were conducted and live and dead pink salmon enumerated. Pillar Creek located about 12 km distant from the city of Kodiak, was foot surveyed only. Each survey performed had fish visibility conditions rated as affected by turbidity, water level, and cloud cover.

During 1990, weir stations were operated on the E. Paramanof, Litnik, Pink, Barling, Saltery, and Akalura streams. Daily, upstream migrating salmonids were identified and enumerated by species. Extending through about four weeks of each population's migration, color coded (unique color for each week) 30.5 cm long floy tags were affixed about 2.5 cm below the base of the dorsal fin to about 150 upstream migrants/population/week. Fish used for tagging were captured in a 2.7 m x 5.0 m trap constructed of aluminum weir panels attached to the upstream side of the weir. After tagging, fish were released upstream of the weir; tag color and number of tagged

fish released by date and location were recorded (Table 1). Foot surveys were conducted on a two to three day rotational basis beginning usually a day after tagging, with live and dead tagged (by color code) and untagged fish counted and recorded. Pillar Creek was foot surveyed only. For 1990, foot survey methods were modified so that after being counted, carcasses were removed from the stream bed and gravel bars to prevent double counting on subsequent surveys.

#### Total Escapement and Commercial Catch Estimation

Pink salmon escapements into Kodiak and Chignik management area streams have been assessed via aerial surveys using fixed wing aircraft and observers for over 30 years. Foot surveys and weir counts are also used for total escapement estimation. Modal "peak" pink salmon counts for a given stream-year are assumed to represent some unknown fraction of the total escapement for each species. Aerial and foot survey conditions were subjectively rated by the observer from poor to excellent depending upon fish visibility within a stream (David Prokopowich, Alaska Department of Fish and Game, Kodiak, personal communication). For this study the historical odd numbered year (odd-year brood line) survey databases for the years 1963-1991 from Kodiak and Chignik were employed to estimate total area wide pink salmon escapements (refer to Data Analysis). The 1989 pink salmon estimated total escapements and catch figures for the KMA were obtained from Barrett et al. (1990) and for the CMA from Barrett (1990).

Total KMA commercial catch numbers for 1963-1985 were extracted from Manthey et al. (1986) and for the years 1987-1991 from the ADF&G Division of Commercial Fisheries fish ticket summary reports. For annual KMA total commercial catch, Kitoi Bay hatchery produced pink salmon were not included. CMA catch numbers 1963-1989 were obtained from Thompson and Owen (1992); catch figures for 1991 were provided by David Owen (personal communication, Alaska Department of Fish and Game, Kodiak).

## Total Available Spawning Habitat

Numerous aquatic habitat inventory sampling designs exist; however most are directed at estimating size of resident fish populations and impacts of land use practices (Platts et al. 1983; Frissel 1986; Murphy et al. 1987; Hankin and Reeves 1988). The use of visual classification of habitat units (Hankin and Reeves 1988) and substrate size (Shirazi and Seim 1979) are proven alternatives to designs relying upon direct measurements. According to habitat suitability index models developed by Raleigh and Nelson (1985), substrate size and water velocity have the largest control over spawning success of pink salmon, while substrate embeddedness is also thought to be influential (Platts et al. 1983). We employed visual classification of substrate size, flow velocity, and stream depth coupled with direct stream width measures for estimating total available pink salmon spawning habitat. The paucity of data for Kodiak and Chignik streams regarding habitat unit types (i.e. pools, riffles, and glide areas) precluded using a stratified systematic sampling design.

	System														
		Barling			Litn	ik		Akalu	ra		Salter	Y		E. Parama	anof
Tagging Event	N	Date	Tag Color	N	Date	Tag Color	N	Date	Tag Color	N	Date	Tag Color	N	Date	Tag Color
1	150	8/6-7	Orange	150	8/7	Orange	134	8/27	Orange	150	8/7-8	Orange	150	7/28	Orange
2	150	8/15	Blue	150	8/18	Blue	140	9/3-5	Blue	151	8/14-15	Yellow	150	8/9-10	Yellow
3	150	8/22-23	Yellow							150	8/22-23	Blue	150	8/17-18	Blue
4							-			150	8/29-30	Pink	150	8/23-24	Pink

Table 1. Tagging dates and numbers of tagged fish released for estimating pink salmon stream life, 1990.

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Streams utilized by pink salmon for spawning within the KMA and CMA are generally small (less than 10 km in length) second and third order streams with low base flows typically occurring in July-August. Spawning habitat data from 45 KMA and 14 CMA index streams were collected during 1989-1990. Utilizing enlargements of United States Geological Survey (USGS) 1:250,000 topographic maps, individual stream maps were constructed on which fishery managers demarcated observed reaches of historical pink salmon spawner distribution for each stream. Total stream length (km) was measured from maps with a calibrated map wheel. Braided channels, mainstem tributaries, and intertidal areas were included in length measures where applicable. A two stage systematic sampling design incorporating each stream's large scale linear trends in gravel size and flow regime was employed to select a maximum of 60 onemeter strip transects from each stream. Total stream length was divided into 300 m primary units and assigned a number (001 to N). Five primary units were randomly selected employing a random number table, and within each primary unit, a systematic sample of 12 one-meter subunits, spaced 25 m apart were chosen for measurement.

Location of selected primary units was accomplished using maps and helicopter instrumentation. Stream width was measured every 25 m providing 12 stream width measurements per selected primary unit. Measures of stream width were to the nearest 2.5 cm from bank to bank where water depth was greater than 15 cm using a hip chain; islands were excluded. Designation of percent spawning habitat (recorded to nearest 10%) was visually estimated as the overall area of a one meter strip transect with designations founded upon ranges of channel substrate size, stream depth, and velocity (Table 2). Gravel embeddedness was evaluated based upon whether extensive force was needed to loosen substrate materials. Suitable substrate size ranges of 0.6 to 13.7 cm; stream depth greater than 15 cm; and a water velocity range of 0.3 to 0.9 m/s were used in this study component. Habitat was deemed unsuitable for spawning if visual determinations were outside of these ranges (Andrew and Geen 1960; Chambers 1956; Dvinin 1952; Krueger 1981; Neave 1966; Raleigh and Nelson 1985; Wilson et al. 1981).

## Fecundity and Egg Retention

Fecundity data (number of eggs per individual) were collected from three KMA streams (Akalura, E. Paramanof, and Litnik) and one CMA (Lake Bay) system in 1989. Sampling was conducted at fish counting weirs for Kodiak streams and with a beach seine measuring 50 m x 2.5 m with 110 mm stretch mesh for the Lake Bay system. Each fish was measured for length (mid-eye to fork-of-tail) to the nearest five mm, and egg skeins extracted. Afterward, each skein was immersed in boiling water and lightly teased to separate individual ovaries, and direct counts of eggs recorded. A total of 301 individuals, 200 from Kodiak, and 101 from Chignik were sampled.

Egg retention data from 34 spawning populations (23 KMA and 11 CMA) were collected during 1989 with a maximum of 150 postspawning females sampled from each population (run). Sampling was structured systematically by selecting every third carcass encountered where concentrations were visually estimated to be less than 1,000 and every fifth carcass for concentrations greater than 5,000. This approach was conceived because of the clumped distribution of carcasses within each stream. All samples were collected after the peak spawning

Spawning Habitat		Repor	ted Ra	nge	0			
Parameter	Minimum	1	Ma	ximum	Source			
Water Velocity	0.4 0.45 0.19 0.10 0.33	to to to to	0.8 0.73 0.66 1.32 0.85	m/s m/s m/s m/s m/s	Dvinin (1952) Hourston and Mackinnon (1957) Wilson et al. (1981) Graybill (1979) Andrew and Geen (1960)			
Average	0.29	to	0.87	m/s				
Optimum	0.30	to	0.90	m/s	Raleigh and Nelson (1985)			
Substrate Partical Size	0.04 0.3 0.3 2.0	to to to to	25 cm 10 cm 10 cm 10 cm	(dia.) (dia.) (dia.) (dia.)	Wilson et al. (1981) Andrew and Geen (1960) Chambers (1956) Lucas (1960)			
Average	0.66	to	13.75	cm (dia.)				
Optimum	Un	known	1					
Water Depth	0.2 m 0.2 0.28 0.37	to to to	7 m 0.7 0.6	8 m 3 m	Dvinin (1952) Chambers (1956) Graybill (1979) Wilson et al. (1981)			
Average	0.26	to	2.8	m				
	≥.15	m			Raleigh and Nelson (1985)			

 Table 2.
 Parameter values employed for estimating pink salmon total available spawning habitat, Kodiak and Chignik Management Area streams.

period from gravel bars and banks of each stream. The body cavity of each carcass was inspected and direct counts of retained eggs and length (mid-eye to fork-of-tail) recorded to the nearest 1 mm. Carcasses exhibiting signs of predation and those deemed unspawned were noted.

## Preemergent Fry Sampling

Preemergent sac fry sampling of KMA index streams has been conducted on an annual or everyother year basis, although sampling did not actually occur in all scheduled years for most streams. The CMA preemergent fry sampling program was discontinued after 1983, however a database of at least eight odd-years were available for analysis. Sampling took place in late February through mid-April, after the period of hatching and early fry development but prior to any significant emergence and emigration having occurred. Numbers of sampling sites (spawning riffles) per stream varies from 2 to 15, and was directly proportional to escapement and stream size. Sampling site selection was based on spawner distribution and habitat usage recorded from aerial surveys; the same riffles were sampled each year. Normally, 10 samples (digs) were taken in an "X"-shaped configuration at a randomly selected site in each riffle. Digs were made with a cylindrical frame benthic sampler that captures material forced out of the substrate by a gas powered hydraulic pump and washed downstream into a five foot long net with attached codend. An aluminum mesh covered the front half of the frame to exclude material washed from upstream, outside of the dig area. The substrate area sampled was 0.18  $m^2$ , and digs were made to a depth of approximately 15 to 47 cm for 1 to 3 minutes. Live and dead fry and eggs were counted separately, and stream temperature, predator presence, stage of fry development, number of egg fragments, and evidence of stream bed scouring and shifts were noted.

## DATA ANALYSIS

### Stream Life

Estimates of average pink salmon stream life during 1989 for the Litnik and E. Paramanof populations were derived from foot survey and cumulative weir count data, and an area-underthe spawner abundance curve (AUC) method (Johnson and Barrett 1988). Within the computer program employed for the AUC model, a stream life value was estimated iteratively until the resulting escapement estimate based on survey counts (converted to spawner days) converged on the total cumulative weir count, or "true" escapement. A spawner day represents each day that a fish is alive within the surveyed stream reach.

For foot survey data collected during 1990, several approaches were used to estimate average stream-life: (1) total number of spawner days derived from foot survey counts (live fish counts) divided by total escapement counted through the weir; (2) total spawner days obtained from foot surveys (live fish counts) divided by total carcass count; (3) iterated AUC (Johnson and Barrett 1988) method employing tagged fish data (live fish counts) where each tagged fish observed is

treated as representing a spawner day; (4) iterated AUC method with tagging data adjusted for estimated proportion of tagged fish detectable; (5) iterated AUC method employing tagged carcass counts and total number of carcasses of a color code.

Adjustment of tagging data was founded upon the assumption that not all live tagged fish within the surveyed stream reach are visible to the observer (Perrin and Irvine 1990). We further assumed that: (1) the proportion of live tagged fish from the j<sup>th</sup> release that were detectable was independent of survey date and release date; and (2) all fish from the initial release of a tag color remained alive for at least one day after their release. Let  $N_{i,j,k}$  denote the number of tagged fish from the j<sup>th</sup> release for the k<sup>th</sup> stream that remain alive i days after release, noting that  $N_{0,j,k}$  is the number of tagged fish released. Also, let  $Ct_{i,j,k}$  equal the survey count of live fish from the j<sup>th</sup> release in the k<sup>th</sup> stream i days after release. From assumptions (1) and (2) above, the estimated proportion of live tagged fish from the j<sup>th</sup> release detectable in stream k (P<sub>jk</sub>) was estimated by:

$$\hat{P}_{jk} = \frac{Ct_{1,j,k}}{N_{0,j,k}} \,. \tag{1}$$

Then, we estimated  $N_{i,j,k}$  for i greater than one by:

$$\hat{N}_{i,j,k} = \frac{Ct_{i,j,k}}{\hat{P}_{j,k}}.$$
(2)

#### Estimated Total Escapement

Historically within the KMA and CMA "peak" aerial and foot survey counts have been used to index pink salmon escapements. Owing to a number of factors peak counts represent only a fraction of the estimated total escapement (Cousens et al. 1982). Total pink salmon escapements for Kodiak and Chignik streams during 1989 were estimated using aerial and foot survey counts, an AUC model (Johnson and Barrett 1988) and a 15 d stream life value (Barrett et al. 1990; Barrett 1990). During 1989 for both areas more streams were surveyed and with greater frequency than any year on record. Performance of the AUC model improves (increased accuracy of escapement estimates) as survey frequency increases.

For the KMA odd-year 1963-1991 stream survey databases, both the frequency and number of streams surveyed were substantially less than in 1989, therefore the methods employed during 1989 were not used. However, 1989 estimated total escapements were employed for expanding odd-year (1963-1991) peak counts.

Estimating total escapements by year (ETE<sub>i</sub>) incorporated the sum of peak aerial and foot survey

counts for n streams in year j (PC<sub>j</sub>), index stream peak counts (PC<sub>j</sub><sup>\*</sup>), sum of all weir counts (WC<sub>j</sub>; assumed to be total escapement); and expansion factors PCXF<sub>89</sub> (expands PC<sub>j</sub>'s) and EDXF (which allows for estimating pink salmon escapements into streams not surveyed in year j) where;

$$PC_{.j} = \sum_{i=1}^{n_j} PC_{ij},$$
 (3)

$$PCXF_{89} = \frac{ETE_{89}}{PC_{89}},$$
 (4)

(5)

and

The KMA escapement distribution expansion factor (EDXF) used aerial and foot stream survey coverage data from odd-years 1969-1987. The total pink salmon escapement for a given year within an area was estimated by

 $EDXF = \frac{\sum_{j=0}^{9} PC_{1969+2j}}{\sum_{j=0}^{9} PC_{1969+2j}}.$ 

$$ETE_{j} = PC_{j} * PCXF_{89} * EDXF + WC_{j}.$$
(6)

The KMA peak count (PC<sub>j</sub>) to estimated total escapement (ETE<sub>j</sub>) expansion factor (PCXF<sub>89</sub>) was estimated to be 1.84 (Appendix A.1; Barrett et al. 1990); and the EDXF expansion factor was 1.06 (Appendix A.2).

The CMA PCXF<sub>89</sub> was estimated to be 1.20 (Appendix B.1; Barrett 1990); and the EDXF using data from 1969-1983 was 1.17 (Appendix B.2). Within the Chignik area for 1985-1991, estimated total escapements were derived using the AUC model (Johnson and Barrett 1988) and a 15 d stream life value (Barrett et al. 1990); adjustment for escapement distribution was also performed.

#### Total Available Spawning Habitat

Total available spawning habitat estimates for KMA and CMA pink salmon index streams employed the following equations from Cochran (1977) incorporating variance estimator modifications for systematic sampling by Wolter (1984), provided by B. Alan Johnson (Alaska Department of Fish and Game, Commercial Fisheries memorandum 9 February, 1990). Estimates for each stream's total available spawning habitat were derived by:

$$\hat{Y} = \frac{N}{n} \sum_{i=1}^{n} M_i \overline{Y}_i \quad ; \tag{7}$$

where the sample mean for the ith primary unit is equal to

$$\overline{Y}_{i} = \sum_{i=1}^{m_{i}} \frac{Y_{ij}}{m_{i}};$$
(8)

and the overall sample mean per subunit is

$$\overline{y} = \sum_{i=1}^{n} \frac{\overline{Y}_i}{n} \,. \tag{9}$$

For these equations n represents number of primary units sampled, N depicts total number of possible primary units,  $m_i$  number of subunits sampled within the i<sup>th</sup> primary unit,  $M_i$  the total number of subunits within the i<sup>th</sup> primary unit, and  $y_{ij}$  the measurement for the j<sup>th</sup> subunit within the i<sup>th</sup> primary unit. The variance estimator for total available spawning habitat is:

$$\operatorname{var} \hat{y} = N^2 \frac{(1-f_1)}{n} S_1^2 + \frac{N}{n} \sum_{i=1}^n \frac{M_1^2 (1-f_{2i}) S_{2i}^2}{m_i}, \qquad (10)$$

which includes

$$S_{1}^{2} = M_{i}^{2} \sum_{i=2}^{n} \frac{(\overline{Y}_{i} - \overline{Y}_{i-1})^{2}}{2(n-1)}$$
(11)

and

$$S_{2i}^{2} = \frac{1}{nm_{i}} \sum_{j=2}^{m_{i}} \frac{(y_{ij} - y_{i,j-1})^{2}}{2(m_{i-1})}, \qquad (12)$$

where f is the sampling fraction n/N for primary units, and  $f_{2i}$  equals  $m_i/M_i$  which is the sampling fraction of subunits within the i<sup>th</sup> primary unit. Several Kodiak index streams (Russian River, Seal Bay, Geographic, and Alinchak Creeks) did not have complete surveys conducted, while two systems had all available subunits measured (Bauman's and Big Waterfall Creeks).

Total available spawning habitat estimates for each index stream were employed to derive spawner densities by dividing the 1989 estimated total escapement by the estimated spawning habitat for that stream, expressed as number of fish per  $m^2$  of spawning habitat. Spawner density was used as an independent variable for analyses incorporating egg retention and preemergent fry data.

#### Fecundity and Length

We modeled the dependence of fecundity (F) on length (L) for the fecundity-length data collected from Litnik, E. Paramanof, Akalura, and Lake Bay as

$$F = \beta_{o} + \beta_{1}L + \beta_{2}Z_{1} + \beta_{3}Z_{2} + \beta_{4}Z_{3} + \beta_{5}Z_{1}L + \beta_{6}Z_{2}L + \beta_{7}Z_{3}L, \qquad (13)$$

where

$$Z_1 = \begin{cases} 1 & , \text{ if Litnik sample} \\ 0 & , \text{ otherwise}, \end{cases}$$

 $Z_2 = \begin{cases} 1 \ , \ if \ Paramanof \ sample \\ 0 \ , \ otherwise, \end{cases}$ 

and,

$$Z_3 = \begin{cases} 1 & \text{, if Akalura sample} \\ 0 & \text{, otherwise.} \end{cases}$$

The model given by equation (13) allows the fecundity-length data from each of the four streams to be fit by a unique linear relationship. We therefore refer to equation (13) as the "full model" for fecundity on length. In contrast is the "reduced model", which allows for only one linear equation to describe the fecundity-length relationship for data from all populations as

$$F = \beta_0 + \beta_1 L. \tag{14}$$

Our interest was not with the exact values of the parameters  $\beta_2$  through  $\beta_7$  but whether these parameters differ significantly from zero. If  $\beta_2$  through  $\beta_7$  are zero, then the full model in

equation (13) reduces to equation (14); i.e., zero values for these parameters indicate that a single linear equation is sufficient to describe the fecundity-length relationship within each of these four populations. However, if any of the  $\theta_2$  through  $\theta_7$  parameters differ from zero, then a single linear equation would erroneously describe the fecundity-length relationship within each of the four streams. We tested the adequacy of the reduced model relative to the full model by analysis of variance (Kleinbaum and Kupper 1978). The null hypothesis for the test was,

$$H_o: \beta_2 = \beta_3 = \ldots = \beta_7 = 0,$$

versus the alternative hypothesis,

$$H_a: \beta_i \neq 0$$
, for at least one i=2,3,...7.

The null hypothesis states that the fecundity-length relationship for each of the four samples falls on the same line. Deriving a single fecundity-length equation would allow for estimating pre-spawning fecundity for egg retention samples, potential egg deposition by index stream, and subsequently egg to preemergent fry survival.

### Egg Retention

Conceptually for the egg retention data, we hypothesized that there should be a positive effect of spawner density (number of fish per  $m^2$  of spawning habitat) on egg retention by individuals of a given population. To test this hypothesis we fit the data to a logistic model (Cox and Snell 1989),

$$P(ER_{i}=0) = \frac{\exp\beta_{0} + \beta_{1}\ln(D_{i})}{1 + \exp(\beta_{0} + \beta_{1}\ln(D_{i}))},$$
(15)

where  $P(ER_i=0)$  is the probability that a female from stream i retains no eggs after spawning, and  $D_i$  is the density of spawners in stream i. Negative values for the  $\theta_1$  parameter in the above model indicate that the probability of a female retaining no eggs, decreases with increasing spawner density. Therefore, negative values for  $\theta_1$  are consistent with a positive effect of spawner density on egg retention.

#### Spawner Success Of 1989 Brood Year

Linear regression analysis of preemergent fry sampling data prior to 1990 was performed for each stream to test dependence of fry density on several independent variables. Live fry density will be referred to as  $fry/m^2$ : live fry in all digs from a stream-year divided by total  $m^2$  sampled. The independent variables were:

- (1) total progeny density  $(TP/m^2)$ : live and dead fry and eggs divided by  $m^2$  sampled;
- (2) egg to fry survival: live fry divided by total progeny;
- (3) proportion of digs with progeny: number of digs with progeny divided by total number of digs;
- (4) spawning activity index: total progeny multiplied by proportion of digs with progeny;
- (5) spawning density index: total progeny divided by proportion of digs with progeny.

The fry/m<sup>2</sup> versus TP/m<sup>2</sup> relationship was consistently the most linear fit to the data, and for some streams linearity was further improved by logarithmic or square root transformations of one or both variables. Fry/m<sup>2</sup> on TP/m<sup>2</sup> models, with or without transformations, were used to test for depression in 1989 spawner success (i.e., low fry/m<sup>2</sup> relative to TP/m<sup>2</sup>). Observed 1990 TP/m<sup>2</sup> values were entered into these models to generate corresponding predicted fry/m<sup>2</sup> values and  $\alpha = 0.10$  critical values of fry/m<sup>2</sup> for depression (one-tailed) to which observed 1990 fry/m<sup>2</sup> values were compared. Probabilities of fry/m<sup>2</sup> values less than or equal to those observed in 1990 were also calculated employing the TP/m<sup>2</sup> values and regression models. There were 47 streams (KMA and CMA combined) each with a database containing greater than 3 years of data for model construction. Employing data from all years, rather than only those associated with the odd-year brood line, improved model fit and allowed more streams to be analyzed. We hypothesized that survival of eggs to live fry after deposition is likely to depend on environmental factors and progeny density rather than on brood year.

The 1989 brood year spawner success was assessed by comparing 1990  $fry/m^2$  for the index streams to odd-year brood line historical  $fry/m^2$  averages, standard deviations of the averages, and 90% confidence intervals of the averages (Appendix A.3).

## Relationship Between Spawner and Live Fry Densities

The nonparametric Mann-Whitney "U" test compares rank sums for two groups of data that have been combined and rank ordered from lowest to highest (Zar 1984). This test was employed to make several comparisons of fry yield and spawner abundance using two indices of spawner abundance:

(1) spawner density: 1989 estimated total escapement for stream i ( $ETE_{189}$ ) divided by the estimated total available spawning habitat for stream i ( $ESH_i$ ),

$$D_i = \frac{ETE_{i89}}{ESH_i}; \tag{16}$$

(2) the difference or standardized residuals  $(E_{di})$ , in standard deviations of  $ETE_{i89}$  stream i

minus the odd-year average  $(X_{eod})$  divided by the standard deviation of average odd-year brood line escapement (#SD<sub>eod</sub>):

$$E_{di} = \frac{ETE_{igg} - \overline{X}_{eod}}{SD_{eod}} \,. \tag{17}$$

This represents the 1989 escapement relative to the historical odd-year average escapement in terms of standard deviations of the historical odd-year average escapement. Our hypothesis regarding this study component is:

 $H_{o}$ : fry production increases as escapement increases for stream i versus the alternative hypothesis,

H<sub>a</sub>: fry production decreases when escapement is above some level  $\in$  '.

Alternatively stated, fry production should tend to increase with increasing escapement level unless overescapement with associated density dependent mortality of eggs and fry occurs, resulting in reduction of fry produced. Therefore, index streams with very high escapements or spawner densities, might show either an increase or decrease in fry yield relative to streams with less extreme escapements.

To test whether streams that showed depressed spawner success (i.e. lower fry/m<sup>2</sup> relative to TP/m<sup>2</sup>) by 1989 spawners had significantly higher spawner abundance than streams that did not show such depression, the ranks of the two indices were compared between the two stream groups in a one-tailed "U" test at  $\alpha = 0.10$ .

Additionally, two relationships were tested for comparing fry yields between index streams with moderate spawner density and those with high spawner density: (1)  $D_i$  vs. fry/m<sup>2</sup>; and (2) #SD<sub>esc</sub> vs.  $\#SD_{frv}/m^2$ , which are the differences in standard deviations that 1989 escapements and 1990 fry/m<sup>2</sup> were from the historical odd-year brood line average escapement, and fry/m<sup>2</sup>, respectively. To stratify streams into two groups for the first of these two-tailed "U" tests at  $\alpha = 0.10$ , individual streams were ranked in order of increasing spawner density with cumulative spawner density calculated stepwise along this ascending order (Table 3). Moderate spawner density streams were those with cumulative spawner density less than the spawner density cut off point of 1.4 fish/m<sup>2</sup>, derived from the midpoint escapement goal and spawning habitat estimates (Table 3; Appendix A.4). The moderate strata included Kodiak streams with individual spawner densities ranging from 0.09 to 4.35 fish/m<sup>2</sup>. All other streams were designated as having high spawner density (range 5.68 to 60.9 fish/m<sup>2</sup>). For the second of these tests, moderate escapement streams were those for which #SDesc was less than 2 (range for Kodiak streams was -0.94 to 1.82), and high escapement streams were those for which 2 was less than #SD<sub>esc</sub> (2.08 to 18.40 for Kodiak). All Mann-Whitney tests were conducted separately for Kodiak and Chignik streams because these areas experience considerably different abiotic Comparisons of fry yield between streams with moderate and high spawner conditions. abundances could not be made for Chignik, owing to a lack of streams with high spawner abundance.

					Cummulative	
		Stream Sp	ecific Density	Spawning		Spawner
Stream		Spawner	Live Fry	Habitat	Escapement	Density_
Name	No.	(fish/m <sup>2</sup> )	(fry/m <sup>2</sup> )	(m <sup>2</sup> )	(No. of fish)	(fish/m <sup>2</sup> )
Moderate Spaw	ner Density	- Streams (	Kodiak Managen	ent Area)		
Narrows	257-401	0.095	264.76	21,917	2,097	0.095
Karluk	255-101	0.132	0.08	853,844	111,977	0.131
Kukak	262-271	0.178	6.73	875,101	115,779	0.132
Marka	252-343	0.347	10.32	935,000	136,573	0.146
E. Paramanof	251-404	0.480	278.27	976,775	154,991	0.159
Zachar	254-301	0.889	12.06	1,094,540	258,974	0.237
Sheratin	259-371	1.714	39.85	1,164,507	378,917	0.325
Barling	258-522	1.900	51.70	1,242,067	526,340	0.424
Dakavak	262-551	1.902	0.36	1,276,702	592,250	0.464
Dog Salmon	257-403	2.290	179.86	1,414,457	907,809	0.642
Kashvik	262-604	2.348	0.81	1,534,814	1,190,409	0.776
Sid Old's	259-242	2.370	116.41	1,593,752	1,330,077	0.835
Kinak	262-451	2.415	161.66	1,616,239	1,384,404	0.857
Kiliuda	258-207	3.068	77.01	1,638,977	1,454,166	0.887
Little R.	253-115	3.097	5.62	1,701,348	1,647,333	0.968
Perenosa	251-830	3.274	72.29	1,730,571	1,743,033	1.007
Miam	259-412	3.655	118.56	1,784,208	1,939,120	1.087
American	259-231	3.910	255.97	1,877,741	2,304,503	1.227
Saltery	259-415	3.937	521.49	1,932,226	2,519,044	1.304
Jute	262-801	4.257	14.54	1,937,699	2,542,347	1.312
Hurst	259-414	4.351	102.65	1,970,451	2,684,882	1.363
High Spawner .	Density Str	eams (Kodi	ak Management	Area)		
Buskin	259-211	5.689	89.75	2,018,396	2,957,667	1.465
Terror	253-331	6.625	0.75	2,080,416	3,368,580	1.619
Uvak-202	254-202	6.479	1,070.00	2,179,998	4,013,737	1.841
Missak	262-402	6.369	131.75	2,185,651	4,049,857	1.853
Kanatak	262-802	6.883	186.68	2,208,656	4,208,207	1,905
Bauman's	253-332	7.674	240.71	2,254,698	4,561,540	2.023
Deadman	257-502	9.679	546.94	2.367.674	5.655.083	2.388
Oil	262-751	10.300	225.90	2.378.890	5,770,695	2.426
Humpy	257-701	10,380	464.79	2,550,629	7,553,891	2,962
Kaiugnak	258-542	14,400	736.78	2,558,844	7,672,214	2.998
Uganik	253-122	16.530	9.42	2,606,981	8,468,113	3.248
Big Creek	262-851	26.460	131.75	2.685.530	10,546,546	3.927
Portage	262-702	29,640	2.33	2,690,215	10,685,426	3.972
Seal	251-901	38.730	869.79	2,690,370	10,691,426	3.974
Danger	252-332	46.690	292.54	2,691,210	10,730,633	3.987
Seven Rivers	258-701	60.880	383.53	2,711,190	11,923,316	4.398
L. Waterfall	251-822	63,530	70.00	2.712.864	12.029.663	4.434
Alinchak <sup>a</sup>	262-651	160.900	27.28	2,716,230	12,571,490	4.628
Moderate Spaw	ner Density	Streams (	Chignik Manage	ment Area)		
Portage	272-842	0.062	64 22	19 107	1.200	0 063
Spoon	273-823	0.071	0 00	42 759	2,900	0.068
Ivan R.	273-722	0.169	13 69	231 759	34 900	0 151
North Fk	272-702	0 197	1 00	360 596	60 297	0 167
Amber Ck	272-703	0 264	0.00	560 637	112 207	0.107
Hook Bay	272-302	0 293	0 00	713 881	158 297	0 222
Humpback	275-502	0.462	65 59	824 175	209 297	0 254
		0.102		523,113	101,001	0.204

Table 3. Moderate and high spawner density index streams of the preemergent fry sampling program, 1989 brood year.

-Continued-

# Table 3. (page 2 of 2)

					Cummulative	
Chron		Stream Spe	cific Density	2 Spawning		Spawner
Name	No.	(fish/m <sup>2</sup> )	(fry/m <sup>2</sup> )	(m <sup>2</sup> )	(No. of fish)	(fish/m <sup>2</sup> )
Foot Bay	273-802	0.575	227.79	842,955	220,197	0.261
Ivanof	275-406	0.655	200.65	1,099,678	388,600	0.353
Agripina	272-961	0.701	279.66	1,217,443	471,192	0.387
Chiginagak 905	5 272-905	1.639	180.85	1,271,727	560,192	0.440
Chiginagak 904	272-904	1.662	0.18	1,290,972	592,192	0.459
Kumliun	272-501	3.693	2.56	1,315,206	681,691	0.518

<sup>a</sup> Based upon incomplete spawning habitat survey.

#### Return Per Spawner Analyses

Spawner-recruit curves were fit to odd brood year (1963 to 1989) pink salmon escapement and return data from the Kodiak and Chignik management areas, separately. We used the models of Ricker (1954) and Beverton and Holt (1957) as two competing models for the spawner-recruit curve. For the Ricker model, number of returns (recruits) R<sub>i</sub>, resulting from number of spawners, S<sub>i</sub>, in year i is modelled as,

$$R_i = \beta_1 S_i \exp\left(-\beta_2 S_i\right), \qquad (18)$$

while for the Beverton-Holt model the relationship is modelled as,

$$R_{i} = \frac{1}{\beta_{1} + \beta_{2} / S_{i}} .$$
 (19)

For both models the parameters  $\mathfrak{G}_1$  and  $\mathfrak{G}_2$  are constrained to be non-negative. Non-zero values of  $\mathfrak{G}_2$  in both models define a negative effect of spawner numbers on subsequent returns per spawner. In both models,  $R_i$  will be 0 when  $S_i$  is 0, but the two models differ in the density dependent response in  $R_i$  to increasing values of  $S_i$ . For the Ricker model,  $R_i$  has a maximum value of  $S_i = (\mathfrak{G}_2)^{-1}$  which is the carrying capacity of the environment for recruits, and decreases to 0 as  $S_i$  increases to  $\infty$ . The Ricker  $\mathfrak{G}_1$  represents the estimated maximum return per spawner, commonly referred to as the productivity parameter. For the Beverton-Holt model  $R_i$  increases asymptotically to  $(\mathfrak{G}_1)^{-1}$  as  $S_i$  increases to  $\infty$ .

We computed maximum likelihood estimates (MLE's) of the parameters for both models using the transform-both sides (TBS) methodology of Carrol and Ruppert (1988). Following Carrol and Ruppert, we used the modified power transformation of Box and Cox (1964) as the transforming function. MLE's were computed using the "pseudo-model" approach of Carrol and Ruppert and the nonlinear estimation routines (NONLIN) of SYSTAT (Wilkinson 1990).

#### RESULTS

#### Stream Life Estimates

Data collection from the Akalura system in 1989 suffered from poor survey conditions and high stream flow, therefore stream life was not estimated. Pink salmon stream life for East Paramanof Cr. (escapement of 20,561) and the Litnik R. (escapement of 7,477) populations averaged 8.5 and 8.3 days, respectively (Table 4).

Table 4. Stream life estimates derived using weir escapement, cumulative carcass, adjusted and not adjusted tagged fish counts from Kodiak Management Area pink salmon populations during 1989-1990.

						Stream	Life	<u>Estima</u>	<u>te (Da</u> Taggi	<u>ys)</u> ng Eve	nt			
Stream					1			2		3		4	A	.vq.
Name	Number	1989 <sup>a</sup>	1990 <sup>a</sup>	ccb	N <sub>adj</sub> c	. Adj.	<sup>N</sup> adj	. Adj.	<sup>N</sup> adj	. Adj.	N <sub>adj</sub>	. Adj.	<sup>N</sup> adj	. Adj.
Barling Cr.	258-522		7.9	15.4	10.4	15.4	7.3	15.1	7.7	14.8			8.5	15.1
Litnik R.	252-342	8.5	6.7		4.0	8.1	2.6	6.3					3.3	7.2
Pink Cr.			6.5	6.8										
Akalura Cr.	257-302	d	3.5	5.9	7.5	9.7	4.3	10.0						
Pillar Cr.	259-101			16.4										
Saltery Cr.	259-415		13.8	15.8	19.1	26.1	13.8	18.7	10.5	14.6	11.2	13.2	13.7	18.1
E.Paramanof	251-405	8.3	14.4	14.2	7.9	11.8	4.0		4.2	11.5	3.8	6.4	5.0	9.9
Average		8.4	8.8	12.4	9.8	14.2	6.4	12.5	7.5	13.6	7.5	9.8	7,6	12.6

<sup>a</sup> Area under the curve method employing foot survey and weir counts.
 <sup>b</sup> Cumulative spawner days from foot survey counts divided by cumulative carcass counts.
 <sup>c</sup> Stream life estimated using tagged fish survey data not adjusted for visibility.
 <sup>d</sup> Data excluded due to high water conditions.

• • •

The 1990 stream life estimates derived from cumulative spawner days divided by total enumerated escapement were: 7.9 d Barling Cr., 6.7 d Afgonak R., 6.5 d Pink Cr., 3.5 d Akalura Cr., 13.8 d Saltery Cr., and 14.4 d East Paramanof Cr. (Table 4); the average over all populations was 8.8 days. Stream life estimated by cumulative spawner days divided by cumulative carcass counts were: 15.4, 6.8, 5.9, 16.4, 15.8, and 14.4 days for the Barling, Pink, Akalura, Pillar, Saltery, and East Paramanof systems; the average over all populations was 12.4 days.

Estimates using live tagged fish were completed for the East Paramanof and Saltery Creek's populations using all four tag colors with about 7-10 days separating each tagging event. Escapements into these systems were 21,749 and 4,556 fish, respectively (Appendix A.5). For the other three populations only two to three tag colors were used. Adjusting estimates for tagged fish visibility (tagged fish detectable  $P_{jk}$ ) increased the unadjusted estimates by 2.6 to 10.0 days. A trend of decreasing stream life over time for the earliest to latest tagging events was observed for the East Paramanof, Saltery, and to a lesser extent Barling populations. Averages over all populations by tagging event, not adjusted were: 9.8 d, 6.4 d, 7.5 d, and 5.5 d. Adjusted estimates by event were: 14.2 d, 12.5 d, 13.6 d and 9.8 d (Table 4). A three day moving average of weir escapement counts for Barling, Saltery, and East Paramanof Creeks depicts tagging events by population, relative to escapement distribution (Figure 5).

Pink salmon stream life estimates generated from tagged fish carcass recovery data averaged over all tag colors for a population were: 17.9 d, 14.6 d, 15.6 d, and 11.4 d (Table 5). A decreasing trend in average stream life by date was observed.

## Estimated Total Escapement and Commercial Catch

Kodiak Management Area odd-year brood line  $ETE_j$ 's (1963-1991) averaged 4.9 million (range 1.21 to 21.0; Table 6). Commercial harvests excluding 1989, have ranged from 0.18 million to 15.25 million. The total return from the 1989 escapement was 21.63 million or about 1.02 returns per spawner; odd-year escapement goals range from 2.3 to 7.0 million with a midpoint of 4.67 (Appendix A.4; Barrett et al. 1990). The midpoint goal has been reached four times during 1963-1987. Escapement and run data (1969-1989) by fishing district depict the 1991 run was greater than average (1969-1987) for all districts except the Northeast District, which was slightly above average (Appendix A.6; Figure 6).

Within the CMA estimated escapements ranged from 0.21 million to 1.13 million. The 1989 escapement of 1.45 million was 2X both the escapement goal and 1963-1987 average escapement. Commercial harvests have averaged 0.7 million (range 25,000 to 1.87 million). The return from the 1989 escapement was 1.96 million (1.34 returns per spawner), whereas the odd-year average (1963-1987) is 1.41 million (Table 7). The escapement goal of 0.7 million has been exceeded five times during 1963-1987.



Figure 5. Three day moving average of daily escapement counts by system depicting tagging events in relation to escapement distribution.

Stream		<u>Stream Life</u>	Estimate	by Taggin	g_Event_	(Days)
Name	Number	1	2	3	4	
Barling Cr.	258-522	19.9	17.6	17.3		
Litnik R.	252-342	10.1	8.4			
Akalura Cr.	257-302	13.4	8.0			
Saltery Cr.	259-415	25.1	18.9	14.8	13.4	
E.Paramanof Cr.	251-405	21.4	20.2	14.8	9.5	
Average		17.9	14.6	15.6	11.4	<u>,</u>

Table 5.Stream life estimates derived for selected Kodiak pink salmon populations using color<br/>coded tagged fish carcass recoveries, 1990.

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Table 6. Kodiak Management Area odd-year pink salmon escapements, commercial catch, average weight, return, and return per spawner, 1963-1989.

Brood	Commercial Catch Estimated Mean Total Weight Peturn				
Year	Escapement	Numbers	(kg)	Return	Spawner <sup>a</sup>
1963	2,554,969	5,480,000	1.5	8,034,969	1.92
1965	2,026,981	2,886,831	1.7	4,913,812	1.19
1967	2,231,335	187,813	1.9	2,419,148	6.99
1969	3,110,525	12,492,576	1.8	15,603,101	2.13
1971	2,293,556	4,332,994	1.7	6,626,550	0.75
1973	1,213,496	511,708	1.8	1,725,204	4.30
1975	2,284,854	2,942,801	1.9	5,227,655	4.63
1977	4,334,716	6,250,667	1.9	10,585,383	3.88
1979	5,709,497	11,121,333	1.7	16,830,830	2.62
1981	5,802,258	9,183,467	1.7	14,985,725	1.44
1983	3,902,959	4,474,760	1.6	8,377,719	2.45
1985	5,710,970	3,886,501	1.6	9,597,471	1.55
1987	4,869,933	4,013,509	1.6	8,883,442	4.36
1989	21,084,539	183,235	1.3	21,267,774	1.02
1991	6,381,344	15,252,123	1.3	21,633,467	
Average	4,900,795	5,546,687	1.7 <sup>b</sup>	10,447,483	2.94 <sup>°</sup>

<sup>a</sup> Return is from brood year+2.
<sup>b</sup> Average weight excluding 1989 and 1991.
<sup>c</sup> Excluding 1989.

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Figure 6. Kodiak Management Area odd year pink salmon estimated escapements, catch, and run numbers by fishing district, 1969-1991.

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Brood Year	Estimated Total Escapement	Commercial Catch	Return	Return/ Spawner <sup>a</sup>
1963	930,032	1,662,400	2,592,432	1.73
1965	494,437	1,117,100	1,611,537	1.12
1967	447,574	108,300	555,874	6.23
1969	1,010,858	1,779,800	2,790,658	1.14
1971	549,067	612,300	1,161,367	0.42
1973	209,438	25,500	234,938	1.81
1975	313,432	66,200	379,632	5.07
1977	987,032	604,700	1,591,732	3.04
1979	1,130,518	1,876,600	3,007,118	1.72
1981	787,072	1,162,600	1,949,672	0.67
1983	209,043	321,100	530,143	3.36
1985	528,989	175,000	703,989	1.20
1987	390,309	246,800	637,109	3.79
1989	1,453,450	27,712	1,481,162	1.34
1991	788,861	1,169,248	1,958,109	
Average	682,007	730,357	1,412,364	2.41 <sup>b</sup>

Table 7.Chignik Management Area odd-year pink salmon escapements,<br/>commercial catch, return, and return per spawner, 1963-1989.

<sup>a</sup> Return is from brood year+2. <sup>b</sup> Excluding 1989.

## Estimated Spawning Habitat and Spawner Densities

Within the KMA 46 of 51 aerial survey index streams had total available spawning habitat data collected. The Karluk and Ayakulik Rivers are the largest within the KMA and provide the greatest amount of spawning habitat 829,515 m<sup>2</sup> and 364,644 m<sup>2</sup>, respectively (Table 8). However, both are even brood year dominant systems (Donnelly 1983). For odd-year dominant systems Humpy, Deadman, Kashvik, and Zachar Creeks along with the Dog Salmon River, all had greater than 100,000 m<sup>2</sup> of estimated spawning habitat. Spawner densities for these systems varied from 0.89 fish/m<sup>2</sup> (Zachar Cr.) to 10.38 fish/m<sup>2</sup> (Humpy Cr.). Spawner densities for all index streams ranged from 0.01 to 60.8 fish/m<sup>2</sup>. Coefficients of variation (CV) for most streams were less than 10% (range 0.1% to 277.0%).

Within the CMA, 14 index streams had spawning habitat estimates generated, ranging from  $17,885 \text{ m}^2$  to  $256,723 \text{ m}^2$ ; coefficients of variation were all less than 10% (Table 9). Spawner densities between systems were variable with most substantially lower than estimates from the KMA. The Amber and Foot Bay systems both had spawning habitat estimated with incomplete surveys.

## Fecundity

Of the four populations sampled for fecundity and length, East Paramanof Cr. pink salmon were the smallest in length ( $X_1$ =465 mm) and least fecund ( $X_f$ = 1,312\_eggs/female). Alternatively, Akalura Cr. fish were the largest and most fecund ( $X_1$ =483 mm; $X_f$ =1,627 eggs/female). The two northern populations (East Paramanof and Litnik) were the smallest as compared to the Akalura and Lake Bay populations located about 162 and 405 km respectively, to the south (Table 10).

Testing for differences between each population's linear relationship of fecundity on length provided an F statistic value of 9.38 with 6 and 293 df; we therefore rejected the null hypothesis (no significant difference of fecundity on length between populations) with P less than  $10^{-8}$  (Table 11; Figure 7). We determined the fecundity-length relationship differs between pink salmon from the four populations sampled. Fecundity, according to our analysis does vary positively with body length within a given locality, however no single fecundity-length relationship was applicable to all localities. Therefore, no attempt was made to apply a single fecundity-length relationship to estimate fecundity for populations sampled for egg retention or preemergent fry.

## Egg Retention

There were 4,743 carcasses examined from 34 spawning populations, 128 fish overall were designated unspawned. Egg retention from the Kodiak and Chignik pink salmon populations varied from zero to intact egg skeins. Mean egg retention for all populations ranged from 1.5 to 146.0 eggs/female with corresponding spawner densities of 0.2 to 60.8 fish/m<sup>2</sup> (Table 12).

Stre	am			Snawr	ing Habit	tat	(m <sup>2</sup> )	1989 Snawner Density
Name	Number	Length	(km)	Est.	SD	CV	m <sub>i</sub> a	(No. Fish/m <sup>2</sup> )
Afognak District				•				ung <u></u>
Malina Cr.	251-105	4.8		5,524	330.8	5.9	60	0.73
E. Paramanof Cr.	251-404	5.6		41,346	917.3	2.2	60	0.48
Big Waterfall Cr.	251-821	0.5		1,488	411.2	27.6	7	3.09
Portage Cr.	251-830	2.4		29,223	1,120.1	3.8	60	,C
Seal Bay Cr.	251-902	1.2		103	39.2	38.1	16	_D,C
Danger Cr.	252-332	6.4		839	138.9	16.6	60	46.73
Alognak R.	252-342	2.4		27,864	481.9	1.7	60	1.49
Marka Cr.	252-343	6.2		58,724	1,416.2	2.4	60	0.35
Northwest Kodiak	District							
Sheratin Cr.	252-371	6.4		67.847	1,287.6	1.9	60	_b,c
Little R.	253-115	14.4		62.370	1,223,2	1 9	60	3 10
Uganik R.	253-122	7.6		46,903	3,536.9	7.5	60	8,76
Terror R.	253-331	14.8		61,192	2,032 5	3 3	60	6 64
Baumans Cr.	253-332	6		44 646	1 733 1	7 9	48	7 91 C
Browns Cr.	254-204	5 1		40 177	1 642 4	⊿ 1	60	1 59
Uvak Cr.	254-202	5 6		97 818	1 112 2	1 1	60	5 59
Zacher Cr.	254-301	12.8		116,852	2,391.7	2.0	60	0.89
Southwest Kodiak	District							
Karluk R.	255-101	34.4		829,515	5,437.2	<0.1	60	0.13
Ayakulik R.	256-201	38.4		364,644	3,460.6	<0.1	60	0.12
Sturgeon R.	256-401	22.4		178,278	1,668.2	<0.1	60	0.01
Alitak Bay Distri	ct							
Narrows Cr.	257-401	5.6		21.532	803.8	37	60	0 10
Dog Salmon R.	257-403	6.8		137,754	2.059.0	1 5	60	2 29
Deadman Cr.	257-502	8.8		110,465	2,080 5	1 9	60	9 90
Humpy Cr.	257-701	14.4		171,739	2,041.9	1.1	60	10.38
Eastside Kodiak D	istrict							
Kiliuda Cr	258-207	3 6		22 727	1 405 0	6 7	60	2 06
Barling R	258-522	5.0 5.4		75 210	1,400.0	1 0	60	1.00
Kajugnak Cr	258-542	2.4		7 924	1,470.0	1.7	60 CO	1 = 10
Seven R	258-701	10.0		10 500	740.5	2.7	c 0	13.12
Miam Cr	259-412	10.0		IJ, J00	1 206 7	2.0	60	00.00
Hurst Cr	259-414	11 6		30,000	1,200.7	2.2	60 60	3.05
Saltry Cr.	259-415	6.8		53,695	922.3	1.7	60	4.59
Northeast Kodiak .	District							
						_		
Pillar Cr.	259-102	1.6		2,541	196.2	7.7	60	16.90
BUSKIN K.	259-211	7.2		47,945	1,688.9	3.5	60	5.69
Sid Old's R.	259-231	8.2		93,533	1,869.1	1.9	60	3.20_
Russian R.	259-242	3.8		58,938	1,406.3	2.4	48	_0,0
American R.	259-231	9.1		93,383	1,625.2	1.7	60	3.91

# Table 8. Estimates of total available spawning habitat and 1989 spawner densities for Kodiak<br/>Management Area aerial survey escapement index streams.

-Continued-

Str	ream		Spawn	ing Habi	1989 Spawner Density		
Name	Number	Length (k	n) Est.	SD	CV	m <sub>i</sub> a	(No. Fish/m <sup>2</sup> )
Mainland Distric	t						
Kukak Cr.	262-271	2.4	21,257	654.1	3.1	60	0.17
Missak Cr.	262-402	2.8	5,292	713.7	13.5	53	6.82
Kinak Cr.	262-451	6.0	21,738	980.8	4.5	60	2.50
Geographic Cr.	262-501	1.2	63	174.7	277.2	12	_b
Dakavak Cr.	262-551	8.4	34,635	2,802.4	8.1	60	1,90
Kashvik Cr.	262-604	8.8	117,682	1,688.6	1.4	60	2.40
Alinchak Cr.	262-651	6.8	3,317	329.2	9.9	48	b
Portage Cr.	262-702	3.6	4,543	618.8	13.6	60	30.57
Oil Cr.	262-751	4.4	10,966	673.7	6.1	60	10.54
Jute Cr.	262-801	3.6	5,473	468.7	8.5	60	4.25
Kanatak Cr.	262-802	2.4	23,005	1,083.7	4.7	60	6.88
Big Cr.	262-851	9.2	77,704	1,424.2	1.8	60	26.74

<sup>a</sup> Number of secondary units sampled.
<sup>b</sup> Streams that were not escapement surveyed in 1989.
<sup>c</sup> Entire stream reach available for spawning was measured.
<sup>d</sup> Incomplete spawning surveys, therefore estimates are approximate.

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## Table 9. Estimates of total available spawning habitat and 1989 spawner densities for Chignik Management Area aerial survey escapement index streams.

St	ream		Spaw	ning Habi	m <sup>2</sup> )	1989		
Name	Number	Length (km)	Est.	SD	CV	mi <sup>a</sup>	Spawner Density (No. Fish/m <sup>2</sup> )	
Eastern Distric	t							
Ocean Bch. Cr. Chiginagak 904 Chiginagak 905 Agripina Cr. Amber Cr.	272-801 Cr. 272-904 Cr. 272-905 272-961 272-702	9.2 1.2 7.5 11.2 21.6	80,013 19,245 54,283 115,699 200,041	2,146.6 1,304.1 971.6 1,863.3 2,399.3	2.7 6.8 1.8 1.6 1.2	60 60 60 48	0.17 1.66 1.64 2.33 0.26	
Central Distric	t							
Hook Bay Cr. Kumliun Cr. North Fork Cr.	272-302 272-501 272-514	18.3 9.9 30.6	153,244 24,234 128,837	3,022.7 551.8 1,925.4	1.9 2.3 1.5	60 60 60	0.30 3.69 0.20	
Perryville Dist	rict							
Ivanof Cr. Humpback Cr.	275-406 275-502	28.2 15.6	256,723 110,294	1,971.7 1,564.7	<.1 1.4	60 60	0.66 0.46	
Western Distric	t							
Ivan Cr. Foot Bay Cr. Spoon Cr. Portage Cr.	273-722 273-802 273-823 273-823 273-842	24.0 2.0 4.5 6.4	189,000 17,885 23,652 19,106	2,960.0 891.5 466.5 593.1	1.6 5.0 2.0 3.1	60 57 60 60	0.17 0.60 0.07 0.06	

<sup>a</sup> Number of secondary units sampled.

Stream		Length (mm)					Fecundity (No. of eggs)					
Name	Area	Mean	Median	SD	<u>Rai</u> Min.	Max.	Mean	Median	SD	<u>Ra</u> Min.	Max.	N
Litnik R.	Kodiak	472	470	26.6	420	530	1,445	1,457	255.3	930	2,127	50
E.Paramanof Ci	r. Kodiak	465	465	19.4	416	501	1,312	1,304	232.5	862	1,960	50
Akalura Cr.	Kodiak	483	485	26.5	426	564	1,627	1,625	324.2	892	2,448	100
Lake Bay Cr.	Chignik	478	485	38.2	376	581	1,587	1,631	301.5	708	2,279	101

Table 10. Pink salmon length and fecundity statistics from populations within the Kodiak and Chignik Management Areas, 1989.

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Table 11. Analysis of variance for the full model regression of fecundity on length and partitioning of sums of squares (SS) for the full model into two components: (1) SS explained by the reduced model alone; and, (2) the additional SS explained by the full model ("Full!Reduced").

Source	df	SS	MS	F
Full Model				
Regression Residual	7 293	134,373 x 10 <sup>2</sup> 159,028 x 10 <sup>2</sup>	1,919,613.0 54,275.61	35.36
Reduced Model				
Regression Residual	1 299	103,802 x 10 <sup>2</sup> 189,599 x 10 <sup>2</sup>	103,802 x 10 <sup>2</sup> 63,410.946	163.70
Full   Reduced				
Regression Residual	6 293	30,571 x 10 <sup>2</sup> 159,028 x 10 <sup>2</sup>	5,095 x 10 <sup>2</sup> 54,275.61	9.38



Figure 7. Pink salmon length-fecundity relationships from four populations studied, 1989.

Table 12. Egg retention statistics and spawner densities observed for Kodiak and Chignik area pink salmon populations, 1989.

0.5		1989	Egg Retention						
Stream		Spawner Density				Ra	nqe	Percent	
Name	Number	(No. of Fish/m <sup>2</sup> )	Mean	Medi	an SD	Min.	Max.	zero eggs	N
Kodiak Area Stream	15					-			
Narrows Cr.	257-401	0.10	15.7	0.0	44.1	0.0	267	57.0	86
Karluk R.	255-101	0.13	75.1	1.0	177.8	0 0	1 122	49 0	147
Marka Cr.	252-343	0.35	2 0	0.0	11 3	0.0	131	77 3	150
E. Paramanof Cr.	251-404	0.48	40.4	3 0	197 3	0.0	1 858	21 7	157
Browns Cr.	254-204	1.59	14 5	0.0	84 9	0.0	943	55 0	150
Barling Cr.	258-522	1,96	11 9	0.0	48 9	0.0	359	70 0	150
Dog Salmon R.	257-403	2 29	7 2	2 0	18 1	0.0	127	70.0	150
Portage Cr.	251-830	3 27	26.2	2.0	57 7	0.0	137	30.0	150
Miam Cr.	259-412	3 65	11 0	2.5	37.2	0.0	376	34.0	100
American R	259-231	3 91	11.0 77 E	0.0	43.3	0.0	370	43.7	126
Hurst Cr	259-414	4 35	22.3	2 0	211 5	0.0	1 240	04.7	150
Saltry Cr	259-415	1.55	71.J	2.0	211.5	0.0	1,249	28.0	150
Buskin R	259-211	5 4 9	24.0	2.0	152.1	0.0	1,309	38.0	150
Terror R	253-331	5.65	24.0	7 0	34.6	0.0	1 007	58.9	145
Kanatak Cr	262-802	6.04	32.3	1.U	50.0	0.0	1,097	45.4	152
Baumans Cr	252 322	7 91	10.3	75 5	59.2	0.0	402	5.3	150
Uganik P	255 552	0 70	140.0	33.5	249.8	0.0	1,4/3	20.0	150
Deadman P	253-122	8.75	19.3	0.0	54.8	0.0	385	59.3	150
Humpy Cr a	257-502	9.90	23.2	1.5	54.4	0.0	520	38.2	144
Kaiugnak Cr	257-701	10.38	132.0	6.0	221.2	0.0	1,167	39.1	64
Dillar Cr.	208-042	15.12	19.5	0.0	63.2	0.0	429	54.1	148
Pillar Cr.	259-102	16.90	22.7	0.0	93.8	0.0	987	52.4	147
Danger Cr.	252-332	46.73	9.4	0.0	24.9	0.0	143	38.8	49
Seven R.	258-701	60.88	146.0	55.0	193.8	0.0	920	22.3	148
Chignik Area Strea	ms								
Ocean Beach Cr.	272-801	0.17	1.9	0.0	7.4	0.0	59	72.2	108
Ivan R.	273-722	0.17	2.3	0.0	6.3	0.0	47	57.7	78
Amber Bay Cr.	272-702	0.26	6.7	0.0	54 1	0 0	655	57 3	150
Hook Bay Cr.	272-302	0.29	1.5	0.0	7 2	0.0	87	66.9	157
Humpback Cr.	275-502	0.46	23 4	1 0	146 4	0.0	1 437	43 6	156
Foot Bay Cr.	273-802	0.60	15.3	1.0	60.4	0 0	397	40 0	45
Ivanof Cr.	275-406	0.66	13 6	ô š	71 9	0.0	840	50.0	162
Chigniagak 905 Cr.	272-905	1.64	16 6	0 0	70 9	0.0	690	50.0	146
Chiginagak 904 Cr	272-904	1 66	57	1 0	20.0	0.0	174	45 0	101
Agripina Cr.	272-961	2 33	60.7	2 0	159 9	0.0	970	40.0	146
Kumulin Cr.	272-501	3 69	8 7	<u> </u>	30 5	0.0	074	34.2 59 C	160
		0.00	0.5	0.0	33.0	0.0	242	20.0	197

<sup>a</sup> There were 84 individuals designated as being unspawned; these fish were excluded from analysis.

Kodiak populations had the largest range of egg retention and spawner densities. The Chignik egg retention data were more variable between populations.

Maximum likelihood parameter estimates (MLE) for the logistic model of spawner density on percent of sample with zero eggs was  $\theta_0$  -0.042 (95% CI -0.023 to 0.107). The estimated  $\theta_1$  value was -0.118 (95% CI of -0.155 to -0.082; Figure 8). These parameter estimates were consistent with the hypothesized positive effect of spawner density on egg retention (i.e., as spawner density increases the probability of a fish retaining zero eggs decreases).

## Success Of 1989 Spawners and Brood Year

Observed 1990 preemergent  $fry/m^2$  was significantly below that predicted by the  $fry/m^2$  vs. TP/m<sup>2</sup> linear regression models for 23 Kodiak and 7 Chignik streams (Tables 13 and 14), indicating that spawner success was poor. Thirteen streams in Kodiak and two in Chignik did not show significant depression in  $fry/m^2$  but eight of these had lower than predicted  $fry/m^2$ . An alternative method for both KMA & CMA index streams of employing escapement to predict preemergent  $fry/m^2$  was discarded (Figure 9).

During 1990, preemergent fry/m<sup>2</sup> was greater than the odd-year brood line historical average in 23 streams in Kodiak (Table 13). For 13 of these, fry/m<sup>2</sup> was greater than the upper limit of the 90% confidence intervals of the average fry/m<sup>2</sup>. In view of  $\#SDfry/m^2$  values (standardized residuals), which compare 1990 fry yield to past yield and variability in that yield, 1990 fry production was extremely high in the Paramanof, Uyak-202, and Saltery systems (Table 13; Figure 10). Fry yield was also very high in Seal Bay, Danger, and Narrows Creeks, even though the first of these three showed significantly depressed fry/m<sup>2</sup>. Of the 13 other streams that had positive  $\#SD_{fry}/m^2$  values, 11 had significantly depressed fry/m<sup>2</sup> as compared to predicted values. The 1990 fry/m<sup>2</sup> was less than the odd-year brood line average in 13 Kodiak streams, and was significantly low in 11 of these. The most negative  $\#SD_{fry}/m^2$  value was only -1.42 for Little Waterfall Creek. The average 1990 fry/m<sup>2</sup> for index streams stratified by fishing district was greater than historical averages in all Kodiak districts except the Mainland, where the average of the historical averages was slightly higher. Average 1990 fry/m<sup>2</sup> values were substantially greater for the Northwest Kodiak, Alitak, and Eastside Kodiak Districts, but only slightly higher for the Afognak and Northeast Kodiak Districts.

For the CMA populations, 1990 fry/m<sup>2</sup> was greater than the odd-year average in only Ivanof River and Portage Creek. The Ivanof value was greater than the 90% CI of the average, but the fry/m<sup>2</sup> was significantly depressed as compared to the predicted value (Table 14). The 1990 fry/m<sup>2</sup> was less than average in the other seven streams, was significantly depressed in six of these, and was below the 90% CI of the average for the Ivan and Spoon Rivers. The deviations of 1990 fry/m<sup>2</sup> from average fry/m<sup>2</sup> were low for Chignik streams; the greatest  $\#SD_{fry}/m^2$  value being that of Spoon Creek -1.56 (Table 14; Figure 10). The average of the 1990 fry/m<sup>2</sup> values for Chignik streams were substantially lower than the average of the historical averages.



Figure 8. Pink salmon egg retention represented as percent of females sampled that retained no eggs as a function of spawner density.

	_	Linear Regression Results							Historical Odd-vear Brood Line Frv/m <sup>2</sup> D			
	-		Number			1990 Fry/m <sup>2</sup>	1	Number	Fry/m <sup>2</sup>			
Stream		_	Sampling	<u>^</u>			Pb.	Years	-		Standardize	ed
Name	Number	Model No <sup>a</sup> .	Years	r <sup>2</sup>	Predicted	Observed	Pred.< Obser.	Averaged	Mean	SD	Residuals	2. 90% CI
Afognak Distr	ict											
E. Paramanof	251-404	3	13	0.798	252.9	278.3	0.6046	11	21.16	25.2	10.20	7.39-34.92
L. Waterfall	251-822	2	8	0.682	532.2	70.0	0.0076	4	484.25	291.5	-1 42	141 24-827 26
Perenosa	251-030	4	14	0.765	289.7	72.3	0.0016	12	219.13	126.3	-1 16	153 64-284 61
Seal Bay	251-901	1	13	0.859	701.3	459.2	0.0054		315 04	236 7	2 34	120 29-509 79
Danger	252-332	1	21	0.849	259 9	292 5	0 7179	11	89 74	91 7	2.24	42 29-137 09
Marka	252-343	4	9	0.954	10.9	10.3	0.2896	0	07,74	01.7	2.10	42.33-137.03
Average						234.5 <sup>d</sup>			174.43 <sup>e</sup>	2		
Northwest Kod	iak Dístri	ct										
Sheratin	259-371	3	22	0.578	153.3	39.8	0.0529	11	40.63	33.3	-0.02	21.54~59.72
Uganik	253-122	1	22	0.720	132.4	9.4	0.0193	12	66 13	68 7	-0.81	30 51-101 74
Terror	253-331	3	22	0.387	56 5	0.7	0.0036	12	11 46	12 0	-0.03	4 79-19 13
Bauman's	253-332	1	23	0 915	843 6	240 7	0.0062	12	212 20	22.0	-0.03	4.70-10.11
Ivak-202	254-202	1	22	0 979	947 6	1 070 0	>0.0002	12	313.30	223.3	-0.32	196.55-430.20
Uvuk - 203	254-203	1	5	0.000	252 7	220 0	0.0034	12	207.04	157.7	5.15	1/5.88-339.39
Zachar	254-301	1	22	0.746	92.0	12.1	0.0024	12	207.59	17.9	0.11	0.00-431.29-
Average						230.1			124.64 <sup>e</sup>			
Alitak Bay Di	strict											
Narrows	257-401	4	13	0.810	309.3	264.8	0 3671	12	91 77	69.1	2 50	55 92-127 62
Dog Salmon	257-403	4	22	0.982	159.5	179.9	0 7241	12	104 60	113 0	0.67	46 00-163 20
Deadman	257-502	4	22	0.985	664 4	546 9	0 0241	12	472 65	196 3	0.60	220 96 524 44
Humpy	257-701	i	23	0.876	742.6	464.8	0.0008	12	220.05	173.5	1.41	126.11-313.99
Average						364.1			212.27 <sup>e</sup>			
Bastside Kodi	ak Distric	t										
Kiliuda	258-207	1	22	0.960	82.28	77.0	0.3925	10	34.62	40.4	1 05	11 47-57 76
Barling	258-522	1	20	0.782	132 25	51 7	0 0223	12	99 49	93.9	-0.51	AP 22.150 74
Kaiugnak	258-542	ī	22	0.899	814.35	736.8	0 2127	12	241 35	209.7	1 61	R1 57-401 12
Seven Rivers	258-701	ī	22	0.668	556 19	383 6	0 0534	10	201 71	150 1	1.01	202 26 266 20
Reaver	259-365-	î		0 995	573 40	540 0	0.0100	14	204.13	120.1	0.62	202.76-366.70
Miam	259-412	ī	12	0.667	119 20	110 6	0.2402	*	4/3.01	403.5	0.19	0.00-948.33*
Unret	200-414	1	21	0.007	110.20	100 0	0.5025	0	10 7-			
Saltery	259-415	1	21	0.344	20/./	102.8	<0.0005	10	49.36	56.4	0.95	16.65-82.06
Durcery	202-413	1	<i>66</i> 			541.5	U.3335	12	94.73	67.5	6.32	59.73-129.72
Average						317.739			158.02 <sup>e</sup>			

# Table 13. Results of linear regression analyses of 1990 preemergent fry sampling data (1989 brood year), and historical odd-year brood line live fry density data for the Kodiak Management Area.

-Continued-

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	_	L	Linear Regression Results						orical Odd-year Brood_Line Fry/m <sup>2</sup> Data			
			Number			1990 Fry/m <sup>2</sup>		Number			Fry/m <sup>2</sup>	
Stream			Sampling	2			P <sup>D</sup> .	Years	Standardized			
Name	Number	Model No <sup>a</sup> .	Years	r²	Predicted	Observed	Pred.< Obser.	Averaged	Mean	SD	Residuals <sup>C</sup>	. 90% CI
Northeast Ko	odi <b>ak</b> Distri	ct										
Monashka	259-101	4	9	0.576	75.4	2.7	0.0085	5	39.65	48.8	-0.76	22.87-56.43
Buskin	259-211	3	22	0.734	268.3	89.7	0.0326	12	143.01	122.1	-0.44	79.72-206.30
American	259-231	1	23	0.978	325.5	256.0	0.0021	12	142.93	124.4	0.91	78.43-207.42
3id Old's	259-242	1	23	0.907	104.4	116.4	0.7024	12	93.02	70.2	0.33	56.63-129.41
verage						116.2			115.75 <sup>e</sup>			
fainland Dis	strict											
Kukak	262-271	1	6	0.806	49.4	6.7	0.0320	3	52.09	47.6	-0.95	0.00-132.34 <sup>f</sup>
lissak	262-271	1	12	0.939	226.1	131.7	0.0008	6	99.51	73.5	0.44	39.06-159.96
linak	262-451	1	12	0.886	238.0	161.7	0.0030	6	142.61	80.9	0.24	76.05-209.17
akavak	262-551	4	14	0.983	18.5	0.4	<0.0005	7	59.84	51.6	-1.07	21.94-97.74
ashvik	262-604	2	14	0.844	7.1	0.8	0.1497	7	44.50	59.5	-0.73	0.79-88.21
linchak	262-651	1	10	0.843	137.5	27.3	0.0025	6	18.68	28.0	0.31	4.21-33.15
ortage	262-702	1	10	0.928	3.4	2.3	0.4890	4	19.52	35.9	-0.48	0.00-61.74 <sup>1</sup>
anatak	262-802	1	9	0.964	176.9	186.7	0.6421	5	173.78	161.5	0.08	19.76-327.80,
lig Creek	262-851	1	9	0.998	155.9	131.7	0.0013	5	115.03	166.2	0.10	0.00-273.51 <sup>f</sup>
werage						72.1			81.09 <sup>e</sup>			

<sup>a</sup> Linear regression model was live fry density (fry/m<sup>2</sup>) vs. total progeny density (TP/m<sup>2</sup>) with or without square root or logarithm transformation of either or both variables. The model numbers represent (dependent variable first): (1) fry/m<sup>2</sup> vs. TP/m<sup>2</sup>; (2) square root fry/m<sup>2</sup> vs. TP/m<sup>2</sup>; (3) square root fry/m<sup>2</sup> vs. square root TP/m<sup>2</sup>; and (4) logarithm fry/m<sup>2</sup> vs. logarithm TP/m<sup>2</sup>.
<sup>b</sup> The probability, according to 1990 TP/m<sup>2</sup> and the linear regression relationship, that fry/m<sup>2</sup> in 1990 would be ≤ the observed 1990 fry/m<sup>2</sup>. Probabilities <0.10 indicate significantly low 1990 fry/m<sup>2</sup>.
<sup>c</sup> Standarized Residuals=(1990 fry/m<sup>2</sup> - historical mean fry/m<sup>2</sup>)/standard deviation of historical mean fry/m<sup>2</sup>.
<sup>d</sup> Average calculated excluding Marka Creek.

<sup>e</sup> Weighted average.

f Zero was assigned to lower confidence limits that were negative numbers.

<sup>g</sup> Average calculated excluding Miam Creek.

## Table 14. Results of linear regression analyses of 1990 preemergent fry sampling data (1989 brood year), and historical odd-year brood line live fry density data for the Chignik Management Area.

		L	inear Regr	ession R	esults	Historical Odd-year Brood Line Fry/m <sup>2</sup> Data						
	_		Number			990 Fry/m <sup>2</sup>		Number			Fry/m <sup>2</sup>	
Stream			Sampling	n			Prob <sup>D</sup> .	Years		St	candardize	d
Name	Number	Model No <sup>a</sup> .	Years	r²	Predicted	Observed	Pred.< Obser.	Averaged	Mean	SD	Residuals	5 <sup>C</sup> . 90% CI
Hook Bay	272-302	1	14	0.820	69.2	0.0	<0.0005	11	18.77	27.8	-0.67	0.00-44.71 <sup>a</sup>
Kumliun	272-501	1	5	0.974	75.2	2.6	0.0009	4	45.93	40.3	-1.08	0.00-120.95 <sup>a</sup>
Chiginagak 905	272-905	1	6	0.995	261.1	180.8	0,0007	2	209.22	83.2	-0.34	0.00-463.29 <sup>a</sup>
Ivan	273-722	3	17	0.958	33.8	13.7	0.0498	9	111.61	113.4	-0.86	31.61-191.61
Foot Bay	273-802	1	11	0.977	304.3	227.8	0.0148	6	265.77	192.7	-0.20	102.58-428.95
Spoon	273-823	3	9	0.932	8.1	0.0	0.0206	5	99.09	63.6	-1.56	39.09-159.08
Portage	272-842	1	17	0.576	41.9	64.2	0.8035	6	40.34	40.0	0.60	14.49-66.19
Ivanof	275-406	3	17	0.983	267.2	200.6	<0.0005	9	99.66	94.3	1.07	39.44-159.88
Humpback	275-502	3	17	0.979	79.0	65.6	0.2900	9	185.38	249.7	-0.48	14.34-356.41
Average						68.7			119.52			

<sup>a</sup> Linear regression model was live fry density (fry/m<sup>2</sup>) as the dependent variable versus total progeny density (TP/m<sup>2</sup>) as the independent variable, with or without square root or logarithm transformation of either or both variables. The model numbers represent: (1) fry/m<sup>2</sup> vs. TP/m<sup>2</sup>; (2)square root fry/m<sup>2</sup> vs TP/m<sup>2</sup>; (3) square root fry/m<sup>2</sup> vs square root TP/m<sup>2</sup>; and (4) logarithm fry/m<sup>2</sup> vs logarithm TP/m<sup>2</sup>.

<sup>b</sup> The probability, according to 1990 TP/m<sup>2</sup> and the linear regression relationship, that  $fry/m^2$  in 1990 would be  $\leq$  the observed 1990 fry/m<sup>2</sup>. Probabilities <0.10 indicate significantly low 1990 fry/m<sup>2</sup>.

<sup>c</sup> Zero assigned to lower confidence limits that were negative numbers.

<sup>d</sup> Weighted average.



Figure 9. Pink salmon preemergent fry/m<sup>2</sup>as a function of escapement by index stream: (A) Kodiak; and (B) Chignik.



Figure 10. Preemergent fry standardized residuals as a function of 1989 escapements: (A) Kodiak; and (B) Chignik.

#### Relationship Between Spawner Abundance and Live Fry Density

There were a wide range of spawner densities experienced within the KMA index streams during 1989 with most escapements higher than the historical average; differences ranged from small fractions to large numbers of standard deviations (Table 15). Few streams had escapements that were lower than average, all by less than one SD. For CMA streams, escapements were below the 1963-1989 escapement average in six of nine preemergent index streams with Chiginagak-905 and Hook Bay Creeks receiving escapements greater than one SD above average (Table 16).

Of the Mann-Whitney comparisons of spawner abundance indices between streams that had depressed fry/m<sup>2</sup> (i.e. low fry/m<sup>2</sup> relative to TP/m<sup>2</sup>) and streams that did not, only the comparison of spawner densities between these stream groups in Chignik yielded significant evidence (P=0.087) that depression was associated with increased spawner density (Table 17). However, the comparison of  $\#SD_{esc}$  values between these Chignik groups provided a test statistic far below the critical value. Comparisons of these indices between stream groups for Kodiak produced test statistics close to but below the critical values. Moderate and high spawner density stream groups in Kodiak had significantly different 1989 fry/m<sup>2</sup> values (P=0.025). The average of the 1989 fry/m<sup>2</sup> values found in moderate spawner density streams (91.796) was much lower than the average value for high spawner density streams (291.41). Comparison of  $\#SD_{esc}$  vs.  $\#SD_{fry}/m^2$  values between moderate and high escapement stream groups in Kodiak yielded a test statistic well below the critical value.

#### Spawner-Recruit

The Ricker and Beverton-Holt models fit to the Kodiak return per spawner data are essentially equivalent. What is important in the context of this study is that the Kodiak data indicates that for both models the  $6_2$  or density dependent parameter is non-zero (Table 18; Figure 11). In contrast, parameter estimates for the Ricker and Beverton-Holt models based on Chignik data do not differ significantly from zero, suggesting that the dependence of return on number of spawners is weak over the range of observed data (Table 18; Figure 12). The predicted KMA 1991 pink salmon return from the odd-year midpoint escapement goal of 4.67 million (Ricker model) provided a return of 11.9 million fish or 2.5 returns per spawner (95% prediction interval: 5.8 to 29.5 million; 0.8 to 4.2 returns per spawner). The maximum difference between replacement and the fitted curve occurs around 7.5 million (Figure 11). The CMA predicted return from the 0.7 million escapement goal was 1.3 million or 1.8 returns per spawner).

Table 15. Preemergent fry sampling index streams classified as having moderate and high escapements within the Kodiak Management Area, 1989 brood year.

Stream		Odd.	-Year Esc	apements	Escapement Years Averaged <sup>a</sup>				
Name	Number	Mean	SD	1989	(1963-1989)	Esc.b	Fry/m <sup>2C</sup>		
Streams Class	lfied as hav	ing Modera	te Escape	ment in 198	39				
Narrows Cr.	257-401	9,540	7,926	2,097	11	-0.939	2.50		
Seal Bay Cr.	251-901	16,627	17,281	6,000	8	-0.615	2.34		
Kinak Cr.	262-451	70,124	51,702	54,327	9	-0.306	0.24		
L. Waterfall	251-822	125,736	161,451	106,347	4	-0.120	-1.42		
Kukak Cr.	262-271	4,767	8,113	3,802	11	-0.119	-0.95		
Dakavak Cr.	262-551	40,966	46,037	46,037	14	0.542	-1.07		
E.Paramanof C:	c. 251-404	12,029	14,287	20,056	12	0.562	10.20		
Danger Cr.	252-332	20,047	21,711	39,207	14	0.883	2.48		
Sid Olds R.	259-242	76,654	46,612	139,668	12	1.352	0.33		
Perenosa Cr.	251-830	43,115	37,596	95,700	12	1.399	-1.16		
Monashka Cr.	259-101	5,718	5,606	14,720	6	1.606	-0.76		
Barling Cr.	258-522	73,541	43,206	147,423	14	1.710	-0.51		
Saltry Cr.	259-415	106,916	58,975	214,541	14	1.825	6.32		
Streams Classi	ified as hav	ing High E	scapement	in 1989					
Kiliuda Cr.	258-207	19,501	24,151	69,762	12	2.081	1.05		
Zachar Cr.	254-301	43,056	27,337	103,983	14	2.229	0.11		
Buskin R.	259-211	82,622	79,360	272,785	14	2.396	-0.44		
Dog Salmon R.	257-403	117,882	68,953	315,559	14	2.867	0.67		
Missak Cr.	262-402	7,765	8,520	36,120	10	3.328	0.44		
Sheratin R.	259-371	46,123	20,818	119,943	6	3.546	-0.02		
American R.	259-231	83,179	75,003	365,383	14	3.763	0.91		
Kanatak Cr.	262-802	32,785	28,689	158,350	11	4.377	0.08		
Uyak 202 R.	254-202	170,582	97,247	645,157	14	4.880	5.15		
Uyak 203 R.	254-203	59,559	48,415	303,600	12	5.041	0.11		
Kaiugnak Cr.	258-542	25,997	16,997	118,323	14	5.432	1.61		
Hurst Cr.	259-414	23,142	20,493	142,535	11	5.826	0.95		
Kashvik Cr.	262-604	57,073	36,251	282,600	11	6.221	-0.73		
Terror R.	253-331	95,579	45,623	410,913	14	6.912	-0,83		
Portage Cr.	262-702	17,682	16,911	138,880	10	7.167	-0.48		
Uganik R.	253-122	125,317	92,153	795,899	14	7.277	-0.83		
Humpy Cr.	257-701	250,989	173,451	1,783,196	14	8.834	1.41		
Deadman R.	257-502	187,443	101.643	1,093,543	14	8.915	0.58		
Bauman's Cr.	253-332	21,903	22,980	353,333	13	14,423	-0.32		
Seven River's	258-701	140.099	70.374	1,192.683	14	14,957	0,62		
Alinchak R.	262-651	26,444	29,894	541,827	13	17.241	0.31		
Big Creek	262-851	99,027	107,563	2,078,433	13	18.402	0.10		

<sup>a</sup> Escapement years are odd-years only for which escapement counts were available. <sup>b</sup> ((1989 escapement - mean escapement)/ standard deviation of mean escapement). <sup>c</sup> ((1990 fry/m<sup>2</sup> - mean fry/m<sup>2</sup>)/ standard deviation of mean fry/m<sup>2</sup>).

Preemergent fry sampling index streams classified as having moderate and high Table 16. escapements within the Chignik Management Area, 1989 brood year.

Streen		00	ld-Year Esc	apements	Escapement Years		съ.
Name	Number	Mean	SD	1989	(1963-1989) <sup>a</sup>	Esc.b	Fry/m <sup>2C</sup>
Streams Classified	as havin	g Moderate	Escapemen	it in 1985	)		
Spoon Cr.	273-823	7,237	5,685	1,700	10	-0.974	-1.56
Ivan Cr.	273-722	211,287	185,450	320,000	10	-0.967	-0.86
Foot Bay Cr.	273-802	20,505	15,015	10,800	10	-0.646	-0.20
Portage <sup>°</sup> Cr.	273-842	7,930	10,755	1,200	10	-0.626	0.60
Cape Kumliun Cr.	272-501	124,223	103,381	89,499	8	-0.336	-1.08
Humpback Cr.	275-502	59,432	43,164	51,000	10	-0.195	-0.48
Ivanof R.	275-406	128,598	89,116	168,403	10	0.447	1.07
Chiginagak 905 Cr.	272-905	33,157	38,645	89,000	10	1.445	-0.34
Streams Classified	as havin	g High Esc	apement in	1 <i>989</i>			
Hook Bay Cr.	272-302	13,865	14,018	45,000	10	2.221	-0.67

<sup>a</sup> Escapement years are odd-years only for which escapement counts were available.
<sup>b</sup> ((1989 escapement - mean escapement)/ standard deviation of mean escapement).
<sup>c</sup> ((1990 fry/m<sup>2</sup> - mean fry/m<sup>2</sup>)/ standard deviation of mean fry/m<sup>2</sup>).

Stream Gro	ups C	ompared	Management Area	Ranked <sup>a</sup> Test Variable	Test <sup>b</sup> Statistic	Critical Value (P=0.10)
Nondepressed fry/m <sup>2</sup> (n=14)	vs.	Depressed fry/m <sup>2</sup> (n=23)	Kodiak	#SD <sub>esc</sub>	193	203
Nondepressed fry/m <sup>2</sup> (n=2)	vs.	Depressed fry/m <sup>2</sup> (n=7)	Chignik	#SD <sub>esc</sub>	7	10
Nondepressed fry/m <sup>2</sup> (n=14)	vs.	Depressed fry/m <sup>2</sup> (n=21)	Kodiak	Spawner Density	183	186
Nondepressed fry/m <sup>2</sup> (n=14)	vs.	Depressed fry/m <sup>2</sup> (n=21)	Chignik	Spawner Density	11 (p=0.0875)	10
Moderate Spawner Density (n=18)	vs.	High Spawner Density (n=21)	Kodiak	fry/m <sup>2</sup>	270 (p=0.0246)	248
Moderate Escapement (n=13)	vs.	High Escapement (n=22)	Kodiak	#SD <sub>fry</sub> /m <sup>2</sup>	151	192

Table 17. Results of Mann-Whitney "U" test comparisons for preemergent fry sampling program index stream groups, Kodiak and Chignik Management Areas.

<sup>a</sup> #SD<sub>esc</sub>= (1989 escapement- historical average escapement)/ standard deviation of historical escapements; spawner density= 1989 escapement/ estimated total available spawning habitat; #SD<sub>fry</sub>/m<sup>2</sup> = (1989 fry/m<sup>2</sup>-historical average fry/m<sup>2</sup>)/standard deviation of historical fry/m<sup>2</sup> values.

<sup>b</sup> Test statistic value greater than the critical value indicates significant difference between the compared groups; significance level is stated in parenthesis.

Management		Para	Parameter				
Area	Model	Number	Estimate	SE	95% CI		
Kodiak	Ricker	61	3.2500	0.7300	1.6500 to 4.8400		
Kodiak	Ricker	6 <sub>2</sub>	0.0053	0.0022	0.0004 to 0.0102		
Chignik	Ricker	6 <sub>1</sub>	2.8800	1.3500	-0.0900 to 2.8800		
Chignik	Ricker	<sup>6</sup> 2	0.0610	0.0570	-0.0640 to 0.1850		
Kodiak	Beverton-Holt	61	0.0031	0.0013	0.0002 to 0.0061		
Kodiak	Beverton-Holt	6 <sub>2</sub>	0.2580	0.0940	0.0510 to 0.4650		
Chignik	Beverton-Holt	6 <sub>1</sub>	0.0400	0.0340	-0.0350 to 0.1160		
Chignik	Beverton-Holt	6 <sub>2</sub>	0.2640	0.2190	-0.2180 to 0.7460		

Table 18.Ricker and Beverton-Holt model's parameter estimates for odd-year brood line<br/>(1963-1989) spawner-recruit data from the Kodiak and Chignik Management Areas.



Figure 11. Spawner-recruit curves (A) Ricker model with 95% prediction intervals and (B) Beverton-Holt model for Kodiak Management Area odd-year broodline pink salmon, 1963-1989.



Figure 12. Spawner-recruit curves for Chignik Management Area odd-year broodline pink salmon, 1963-1989.

## DISCUSSION

Stream life estimates generated from the various methods employed averaged 8.4 to 17.9 days and are similar to estimates generated from other studies (Perrin and Irvine 1990; Sharr and Sharp 1990). Pink salmon stream life appeared to decrease as the run progressed which is consistent with results from studies conducted in Southeast Alaska (Kingsbury 1977; Thomason and Jones 1984). Estimates generated using tagging data from Prince William Sound ranged from 7.2 to 21.9 days, whereas stream life reported by weekly strata had a grand mean of 13.9 days (range 9.9 d to 17.1 d); estimates decreased over time within a system (Sharr and Sharp 1990).

Stream life estimate variability between methods both within and among streams made selection of a single value difficult. Tagging data estimates adjusted for proportion of detectable tags had the least amount of perceived bias when compared to other methods. Given the wide range of estimated values and the influence stream life has on escapement estimation using the AUC approach and aerial survey data, we felt using a 15 d value was prudent. Had we incorporated more streams of various sizes within our study, a suite of values tailored to run timing and stream size similar to the approach used by Sharr and Sharp (1990) would have been instituted. Regardless, Perrin and Irvine (1990) suggest caution be exercised when using stream life estimates for escapement estimation which transcend areas, streams, or years.

Total available spawning habitat estimates for KMA index streams ranged from less than 1,000  $m^2$  (Danger Cr.) to 829,515  $m^2$  (Karluk R.). For some streams, estimates were considered approximate owing to incomplete spawning ground surveys. Streams having CV values greater than 10% were those where habitat surveys were incomplete. Spawning habitat estimates generated for CMA streams varied from 17,000 to 250,000  $m^2$ ; coefficients of variation were all less than 10%. Spawner density (D<sub>i</sub>) estimates are precise as indicated by the spawning habitat CV values. However, our inability to accurately stratify populations which have bimodal run timing into early and late run components suggests that Di estimates for some spawning stocks might be overestimates. Alternatively, some spawning reaches had densities that were probably greater than those expressed in terms of the entire stream, as within stream direct measures for specific spawning areas were not estimated. In light of the escapement and total available spawning habitat estimates, the D<sub>i</sub>'s are not without error. We believe this error is relatively minor and imparts negligible influence on the egg retention and preemergent live fry analyses.

Fecundity estimates and the variability encountered between the four populations corresponds well with information on pink salmon presented by Heard (1991). Heard states that fecundity values deviate considerably between stocks, years, and regions. Therefore, our results are not atypical. Fecundity-length relationships for the four populations investigated were significantly different from one another. The lack of a single fecundity-length relationship severely limited our analysis of egg retention and preemergent fry survival.

For the egg retention data collected (Kodiak and Chignik populations combined), we found a highly variable relationship between mean egg retention and spawner density. A trend of increasing mean egg retention as a function of spawner density for most populations (Kodiak and

Chignik combined) was suggested by the data. However, the number of eggs retained was highly variable within samples and a known, but unspecifiable, relationship between fecundity and length further reduced clarity of this perceived trend. Interpretability and model fit was improved by transforming egg retention into a binary variable i.e., a value of zero when no eggs were retained and a value of one when eggs were retained. The index of egg retention for each population was expressed as the percent of each sample having zero eggs retained. Mean egg retention for the Kodiak pink salmon populations ranged from 1.5 to 146.0 eggs per female. Obtaining a statistically significant parameter estimate indicated the existence of a positive effect of spawner density on egg retention. Invariably, analyses were biased by omitting fish classified as "unspawned". However, ascertaining whether a fish had attempted to spawn prior to death or had died from kidney failure, oxygen depravation, or an epizootic proved unrealistic.

Few investigators have reported a relationship between spawner density (number of fish/m<sup>2</sup>), egg retention, and subsequent fry survival with pink salmon. Hanavan and Skud (1954) showed that for *in-situ* pen enclosures with density controlled at 1.68 fish/m<sup>2</sup>, mean egg retention was 409.4 eggs/female, while for densities of 3.7 fish/m<sup>2</sup> egg retention averaged 624.2 eggs/female. Corresponding egg to fry survival was reported as 10.1% and 7.7%, respectively. Hanavan and Scud (1954) cautioned against using this information for calculating optimum spawner densities, but felt that crowding was influential on subsequent fry survival. No causal mechanism other than mechanical disruption of previously deposited fertilized eggs was discussed. McNeil (1964), citing *in-situ* experiments conducted within two Southeast Alaska streams, reported that spawning success (measured as number of females spawning successfully within a given area) was greatest at 330 females per 100 m<sup>2</sup> of spawning habitat. Schroder (1973), relating density to spawning success with Chum salmon reported that spawning pairs were free from aggressive behavior only 8% of the time at spawner densities of 0.9 females/m<sup>2</sup>, while mating activity decreased or ceased at densities above this level. For sockeye salmon data is limited, but Manzer and Miki (1986) have suggested that increased spawner density may cause increased egg retention.

Heard (1991) reports that pink salmon egg retention as a percentage of fecundity ranges from less than 1% to 40%. A value intermediate between these two would limit egg deposition and reproductive potential for a population. We found that mean egg retention for the streams with the highest spawner densities (140, 132, and 146 eggs/individual with spawner densities of 7.9, 10.3 and 60.9 fish/m<sup>2</sup>, respectively) when evaluated using a fecundity value of 1,400 eggs/female, represent less than 11% of each populations total egg capacity. This would represent an insignificant impact on reproductive potential in light of superimposition and other avenues of egg loss. However, if populations with the highest levels of egg retention also had substantial loss to a populations potential egg deposition. Results from the preemergent fry study component discount that this occurred for KMA spawning populations during 1989.

The foregone salmon harvests of 1989 resulted in the largest escapement on record for the Kodiak Management Area (L. Malloy, ADF&G Kodiak, pers. comm); only five Kodiak index streams had less than average escapement. Brood year spawner success (predicted live  $fry/m^2$  from TP/m<sup>2</sup>, relative to observed 1990 live  $fry/m^2$ ) by fishing district was apparently very good in the Northwest Kodiak, Alitak, and Eastside Kodiak Districts, and near average within the

Afognak, Northeast Kodiak, and Mainland Districts (Table 13). The 1989 brood year seems to have had poor success in Chignik but escapement for streams in this area were generally low (Mike Thompson, Alaska Department of Fish and Game Anchorage, personal communication). Brood year success was apparently good to exceptionally good in most Kodiak index streams (when evaluated independently) but was good in only two Chignik index streams. Some Kodiak streams and most Chignik streams showed only fair to low brood year success but none had particularly poor success. Significant depression in 1990 fry/m<sup>2</sup> relative to TP/m<sup>2</sup> indicated poor spawner success in most Kodiak and Chignik streams. Evidently, only the Uyak-202 River and Saltery Creek had exceptionally good spawner success.

Of the 23 Kodiak streams that had depressed spawner success, 12 had good brood year success (better than average  $fry/m^2$ ) and 11 had poor brood year success (less than average  $fry/m^2$ ), suggesting that overescapement with density dependent mortality occurred in some streams. High spawner density streams had significantly higher  $fry/m^2$  than moderate spawner density streams because the moderate group contained streams with very low spawner densities and because the cumulative spawner density goal (1.4 fish/m<sup>2</sup>) used to stratify the streams into the two groups may have been too low. Had the cut-off point been higher, perhaps 3.0, the two stream groups would have been much more similar in fry yields, suggesting that the Kodiak midpoint escapement goal is too low to obtain maximum fry yield. Moderate and high escapement stream groups were stratified at a level that was probably more extreme, and these groups did not differ significantly in their deviations from average  $fry/m^2$ . A well defined predictive relationship between parent year escapement and preemergent  $fry/m^2$  produced would have undoubtedly simplified the analyses.

Although Kodiak pink salmon escapements were very large in 1989 and the spawner abundance indices tended to be somewhat higher for streams that showed significant depression than for streams that did not, there was no conclusive evidence that depression in fry/m<sup>2</sup> was directly caused by spawner density. Most Chignik index streams had low spawner abundance and depressed spawner success, although environmental conditions in this area are generally more extreme than in Kodiak and may have resulted in depressed 1989 spawner success. However, the spawner abundance indices quantify spawner abundance in entire streams and may not reflect spawner abundance in the sampled spawning riffles. Only 4 Kodiak streams had predicted 1990 fry/m<sup>2</sup> lower than the historical average fry/m<sup>2</sup>, suggesting that egg deposition was greater than usual in most sampled spawning riffles.

Donnelly (1983) and Eggers et al. (1991) have reported significant density dependent relationships for Kodiak pink salmon populations. Both the egg retention and preemergent fry study components suggest that a density dependent response occurred between spawning and emergent fry for some Kodiak systems. This is partially substantiated with existence of a statistically significant density dependent parameter estimate for both the Ricker and Beverton-Holt models as applied to the Kodiak data.

The KMA and CMA catch data are assumed to be precise. Escapements estimated using a 15 d stream life value either directly (AUC approach) or indirectly (expansion factors) coupled with error associated with aerial survey data are of potential concern. Donnelly (1983) used an expansion factor of 2.5 to convert aerial survey peak counts to estimated total escapement; expansion factors used herein, when combined were similar to this value. Sharr and Sharp

(1990) reported from preliminary analysis that escapements not corrected for stream life and aerial survey bias were 38% below corrected estimates. Walters and Ludwig (1981) consider escapement measurement errors less than or equal to 30% to be "accurate". This level of error is probably realistic relative to evaluating the Kodiak and Chignik escapement data. Return per spawner analyses for both areas did not account for this error and probably led to the narrow confidence intervals for the parameter estimates. Significant parameter estimates from Kodiak, and non-significant estimates from Chignik could be explained by larger measurement error within the Chignik data than for Kodiak. Had we been able to quantify and correct the escapement estimate bias for both Kodiak and Chignik escapement data, a more concrete return per spawner relationship for both areas would have resulted.

Overall, no conclusive evidence of reduced production of pink salmon adults from the 1989 escapement was found in our study.

## CONCLUSIONS

It is apparent that both egg retention and preemergent fry analyses demonstrated impact caused by spawner densities experienced within some streams during 1989. Evaluation of the overall 1989 brood year success using spawner-recruit analyses also demonstrated a density dependent response to the 1989 escapement. However, the variability in the Kodiak pink salmon return per spawner data made numerical estimation of the damage caused by this escapement event tenuous. The major question remaining unanswered is whether the 1989 odd-year escapement event has caused depressed production from the even-year brood line; since 1989, even-year brood line production has been well below forecast. This forecast is one of the most accurate for any salmon species in Alaska. We feel after evaluating even-year preemergent fry indices and escapement numbers that the causal mechanism for this depressed production is marine derived.

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APPENDIX

Appendix A.1.	Pink sal	lmon pe	ak a	nd estim	ated total	escape	ement o	count	ts by
	fishing	district	for	streams	assessed	using	aerial	and	foot
	surveys	within	the	Kodiak I	Managem	ent Ar	ea, 19	89.	

	Numbe	_	
	Peak Count	Escapement	
Stream No.	(PC)	(ETE)	ETE/PC <sup>a</sup>
Afognak District			
251-101	0	0	
251-102	409	778	1.90
251-103	4,000	7,604	1.90
251-104	300	570	1.90
251-105	1,200	4,022	3.35
251-109	500	951	1.90
251-201	774	1,471	1.90
251-202	651	1,238	1.90
251-206	9	17	1.90
251-207	200	380	1.90
251-301	0	Ō	
251-403	200	1,351	6.76
251-405	75	143	1.90
251-406	27,500	27,500	1.00
251-407	8,425	8,425	1.00
251-502	225	225	1.00
251-503	0	0	
251-504	615	615	1.00
251-505	325	325	1.00
251-505A	424	424	1.00
251-506	274	274	1.00
251-507	287	287	1.00
251-510	2,000	2,000	1.00
251-811	41	41	1.00
251-812	2	2	1.00
251-813	2,150	2,150	1.00
251-821	0	0	
251-826	0	0	
251-829	0	0	
251-832	75	143	1.90
251-901	1,500	6,000	4.00
252-103	1,150	2,186	1.90
252-104	325	618	1.90
252-301	1,500	1,500	1.00
252~302	50,000	95,052	1.90
252-303	0	0	

-Continued-

E.

	Numbers of Fish				
		Estimated Total			
<b>6</b> , <b>1</b>	Peak Count	Escapement			
Stream No.	(PC)	(ETE)	ETE/PC <sup>a</sup>		
252-305	3,000	3,000	1.00		
252-306	150,000	285,157	1.90		
252-307	60,000	114,063	1.90		
252-308	3	6	1.90		
252-309	60,000	114,063	1.90		
252-317	1	8	8.00		
252-318	150,000	285,157	1.90		
252-319	0	0			
252-323	150,000	285,157	1.90		
252-331	2,356	2,356	1.00		
252-331A	1,429	2,717	1.90		
252-331B	1,156	2,198	1.90		
252-332	30,000	39,207	1.31		
252-333	15,000	28,516	1.90		
252-335	189	359	1.90		
252-337	3	6	2.00		
252-338	2	4	1.90		
252-339		13	1.90		
252-343	10,000	20,794	2.08		
Subtotal	738,282	1,349,069	1.83		
Northwest Kodiak	District				
253-114	2,500	3,707	1.48		
253-115	84,000	193,167	2.30		
253-121	8,000	13,600	1.70		
253-122	520,000	795,899	1.53		
253-141	1,800	3,422	1.90		
253-142	900	4,400	4.89		
253-311	1,500	2,852	1.90		
253-313	12,000	12,000	1.00		
253-321	6,000	11,406	1.90		
253-331	367,000	410,913	1.12		
253-332	325,000	353,333	1.09		
253-333	800	2,740	3.43		
254-103	2,000	3,802	1.90		
254-105	1,400	1,400	1.00		
254-201	60,000	60,000	1.00		
254-202	300,000	645,157	2.15		
254-203	165,000	251,953	1.53		
254-204	34,000	63,820	1.88		
234-203	500	1,500	3.00		
254-200	1,000	1,000	1.00		
254-207	1,000	2,200	2.20		
25-2-200	1 500	2,000	2.00		
254-209	1,500	1,500	1.00		
254-210	I,500	1,500	1.00		
254-212	2,500	2,500	1.00		
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Numbers of Fish						
		Estimated Total				
	Peak Count	Escapement				
Stream No.	(PC)	(ETE)	ETE/PC <sup>a</sup>			
<u> ,</u>		·····				
254-301	95,000	103,983	1.09			
254-302	3,000	3,000	1.00			
254-401	48,000	91,250	1.90			
254-402	2,500	4,753	1.90			
254-403	100	190	1.90			
254-404	1,500	4,833	3.22			
254-405	400	1,920	4.80			
259-363	0	D				
259-365	236,000	432,100	1.83			
259-366	8,700	16,539	1.90			
259-368	7,000	13,307	1.90			
259-371	39,000	119,943	3.08			
259-372	4,500	14.700	3.27			
259-373	1,800	1,800	1 00			
259-381	1,000	1,000	2.00			
259-382	54 000	126 000	2 33			
259-383	54,000	120,000	2.55			
259-391	5 400	5 477	1 01			
200-302	3,400	5,477	1 90			
255-352	2,000	7,804	1.90			
207-094	3,200	3,200	1.00			
233-333	200	344	1.90			
259-397	300	1,500	5.00			
259-398A	35	68	1.90			
259-398B	5	10	1.90			
259-399	3	6	1.90			
Subtotal	2,420,630	3,803,498	1.57			
Southwest Kodiak	District					
256-101	0	0				
256-102	0	0				
256-202	0	ō				
256-301	1.500	2.852	1.90			
256-302	-,•	-,				
256-303	Õ	Ő				
256-303A	n n	Ő				
256-401	1 250	1 692	1 35			
256-402	3,000	5,800	1.93			
Subtotal	5,750	10,344	1.80			
Alitak Bay District						
257-101	0	0				
257-102	1,000	1,000	1.00			
257-301	-, ····	-,				
257-303	600	1.400	2 33			
257-401	550	2 097	3 81			
257-402	300	570	1.90			
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-Continued-
	Numbe	ers of Fish	
		Estimated Total	
	Peak Count	Escapement	
Stream No.	(PC)	(ETĒ)	ETE/PC <sup>a</sup>
·			
257-405	6,000	6,000	1.00
257-502	410,000	1,093,543	2.67
257-503	87,000	139,417	1.60
257-504	32,000	67,133	2.10
257-601	12,000	24,800	2.07
257-602	35,000	72,333	2.07
257-603	55,000	62,962	1.14
257-604	123,000	212,733	1.73
257-605	7,000	14,260	2.04
257-701	690,000	1,783,196	2.58
257-702	94,000	103,400	1.10
257-702A	65,000	134,333	2.07
257-703	0	0	
Subtotal	1,618,450	3,719,177	2.30
Eastside Kodiak	District		
250 101	2 600	12 080	4 05
258-101	2,600	12,880	4.95
258-201	4,700	28,420	6.05
	200	271	1.30
258-203	150	120	1.00
258-204	200	227	1 1 2 2
258-205	4 100	7 652	1 07
258-207	4,100	69 763	1.07
258-207	60,500	09,702	1.10
258-208	1 600	1,647	1.83
258-210	1,000	1,000	1.00
258-211	400	647	1 62
250 211	400	10 900	2 70
258-213	28 500	43 700	1 53
258-301	2800	2 800	1.00
258-302	3 500	3 500	1.00
258-303	2,300	5,500	3.00
258-304	3 000	10 247	3.00
258-305	575	1,648	2 87
258-306	150	430	2 87
258-307	1 100	3 520	3 20
258-401	1,100	3,520	5.20
258-511	Ő	ů	
258-512	11.000	11.000	1 00
258-513	0	,	
258-514	10,800	35,800	3.31
258-521	15,000	42,437	2.83
258-522	53,300	147,423	2.77
258-531	. 0	0	
258-532	Õ	ō	
258-533	Ō	Ō	
258-541	30,500	57,133	1.87

Appendix A.1.	(page 5 of 8)
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	Numb	Numbers of Fish								
		Estimated Total								
	Peak Count	Escapement	/							
Stream No.	(PC)	(ETE)	ETE/ PC <sup>4</sup>							
258-541A	0	0								
258-542	78,000	118,323	1.52							
258-544	17,000	19,257	1.13							
258-551	2,600	10,450	4.02							
258-544	1,000	1,533	1.53							
258-555	300	620	2.07							
258-601	4,000	8,267	2.07							
258-602	50,400	81,256	1.61							
258-603	42,000	81,200	1.93							
258-701	450,000	1,192,683	2.65							
258-702	68,000	146,467	2.15							
258-703	47,000	65,033	1.38							
258-704	10 000		1 74							
258~705	19,000	25,367	1.34							
250~700	100	190	1 80							
258-707	100	190	1.90							
258-803	0	0								
258-804	Ő	0								
258-805	ő	õ								
258-807	ő	õ								
258-808	100	200	2.00							
258-809	0	0								
258-810	0	0								
258-851	0	0								
258-852	100	100	1.00							
258-901	1,800	3,720	2.07							
257-902	17,000	35,133	2.07							
258-903	3,000	3,000	1.00							
258-904	2,000	2,000	1.00							
259-401	3,600	6,844	1.90							
259-403	100	190	1.90							
259-411	2,000	3,802	1.90							
259-412	61,000	196,087	3.21							
259-414	96,000	142,535	1.48							
259-415	81,000	251,741	3.11							
259-415	6,500	12,357	1.90							
259-417	300	1,300	2.50							
259-417A	0	0								
259-418A	900	ויל ו	1 90							
259-418B	900	±,,,±±	1.50							
259-419	9	ő								
259-420	ō	ō								
259-421	0	0								
259-422	20,200	47,160	2.33							
259-423	700	1,820	2.60							
259-424	46,000	103,478	2.25							
259-425	6,300	16,647	2.64							
259-426	800	1,521	1.90							
259-427	200	200	1.00							
259-428	0	0								
Subtotal	1,369,077	3,076,233	2.25							
		Continued-								

	Numb	_	
	• -	Estimated Total	
Stream No.	Peak Count (PC)	Escapement (ETE)	ETE/PC <sup>a</sup>
Northeast Kod:	iak District		
259-101	8,000	11,519	1.44
259-102	42,100	42,950	1.02
259-105	4,000	7,298	1.82
259-221	22,000	41,823	1.90
259-222	36,500	108,498	2.97
259-223	113,000	329,627	2.92
259-231	152,000	365,383	2.40
259-233	500	951	1.90
259-235	3,700	5,427	1.47
259-242	46,000	139,668	3.04
259-243	47,700	90,680	1.90
200-244	10,200	10 201	1.00
259-245	10,200	300	1.90
259-251	39.400	74,901	1 90
Subtotal	538,000	1 251 015	2 33
Mainland Dist	rict	1,231,013	4.55
Maimiand Dist			
262-101	100	190	1.90
262-102	12,000	21,693	1.81
262-103	600	1,141	1.90
262-104	0	0	
262-105	0	0	
262-106	0	0	
262-107	2 600	C 480	1 00
202-100	3,800	6,480	1.80
262 105	Ü	0	
262-111	0	0	
262-113	3 800	7 224	1 90
262-151	0,000	0	2.30
262-152	15,000	28,516	1.90
262-153	48,000	91,250	1.90
262-154	0	0	
262-155	800	1,521	1.90
262-201	3,000	12,000	4.00
262-202	0	0	
262-203	0	0	
262-204	0	0	
262~205	42,000	79,844	1.90
262-206	6,000	11,406	1.90
202-207	U 1 0 0	0	1 00
202-204 262-271	2 000	5 0 0 0 Nrt	1 80
262-271	∠,000	5,002	1.30
262-273	0	0	
262-301	0	0	
262-351	9.000 <sup>°</sup>	17.109	1.90
262-352	3,800	7,224	1.90
	-	-	

	Numb	_	
		Estimated Total	
Stream No.	Peak Count (PC)	Escapement (ETE)	ETE/PC <sup>a</sup>
262-401	3,800	7,224	1.90
262-402	19,000	36,120	1.90
262-451	37,000	54,327	1.47
262-501	200,000	380,209	1.90
262-502	0	0	
262-503	1,000	1,901	1.90
262-504	10,000	10,000	1.00
262-505	10,000	10,000	1.00
262-551	38,000	65,910	1.73
262-552	0	0	
262-601	0	0	
262-602	0	0	
262-603	0	0	
262-604	80,000	282,600	3.53
262-605	10,000	22,667	2.27
262-606	20	38	1.90
262-651	445,000	541,827	1.22
262-652	162,000	201,400	1.24
262-653	230,000	230,000	1.00
262-654	0	0	
262-655	300	620	2.07
202-050	13,600	33,320	2.45
202#7UL	44,000	92,000	2.09
262~702	67,000	138,880	2.07
262-703	42,000	42,000	1.00
262-704	20,000	27,203	1.36
262-705	12,000	19,867	7.00
262-760	73 000	115 612	1 50
262-752	10,000	11 560	1 16
262-753	3 500	£ 554	1 90
262-801	22,000	23 303	1 06
262-802	76 000	158 350	2 08
262-803	500	500	1 00
262-804	200	200	1 00
262-850	5.000	5.000	1 00
262-851	1,320,000	2,078,433	1.57
262-852	91,000	110.580	1.22
262-853	66,000	124,200	1.88
262-854	10,000	12,000	1.20
262-856	100	190	1.90
262-858	1,900	2,489	1.31
262-859	0	0	
262-860	0	0	
262-861	9,000	18,600	2.07
262-862	200	200	1.00
262-863	0	0	
262-864	0	0	
262-865	50	95	1.90
262-866	800	1,521	1.90
262-868	100	190	1.90

-

	Numb		
Stream No.	Peak Count (PC)	Escapement (ETE)	ETE/PC <sup>a</sup>
262-951 262-952	24,000 28,000	45,625 53,229	1.90 1.90
Subtotal	3,335,870	5,256,284	1.58
Total	10,026,059	18,465,619	1.84

<sup>a</sup> Estimated total escapements derived by the AUC method included fractions of whole numbers which the ETE/PC ratio reflects in several instances.
<sup>b</sup> Estimate includes estimates of fish spawning downstream of the counting weir.

	Peak Count North N												
Stream No.	1969	1971	1973	1975	1977	1979	1981	1983	1985	1987	Mean (1969-87)	1989	Index Stream
Afognak Dis	trict												
251-101	0	3,200	200	0	0	5,000	450	600	1,200	7,500	1,815	C	)
251-102	25	0		0	60	600		0	-		114	409	)
251-103	0	50			30	675		170	0	0	132	4,000	)
251-104		10			0			0		0	3	300	)
251-105	1,000	100	3,800	900	10,000	19,000	16,500	13,500	3,500	25,000	9,330	1,200	) *
251-106		0	0			. 0	0		,	•	. 0		
251-201	0	100		0	20	1,124	0	7,000	400		1,081	774	
251-201 <sup>a</sup>	0	0			0	. 87	0	400		0	70		
251-202	0	100	0		600	450		3,300	20		639	651	L
251-203					1,200	600	7,500	•			3,100		
251-204						68	. 0				. 34		
251-301	1,600	7,200			0	8,000	28,300	1,500	7,599	5,500	7,462	0	)
251-302	25	300	700	0	0	15,000	11,000	. 0	. 0	· 0	2,703		
251-403	8,200	1,500	800	1,500	500	1,500	8,000	7,000	400	0	2,940	200	}
251-404	2,900	1,300	3,000	200	1,300	13,000	26,000	13,000	9,900	6,500	7,710		*
251-405	0	0		0				0	10	-	2	75	5
251-406		0		0	0	15					4	27,500	)
251-407		0				3,499					1,750	8,425	5
251-501					0				0		0		
251-502					0	5					3	225	5
251 <b>-503</b>				0	0	0					0	0	)
251-504				0	0	0					0	615	5
251-505				0	0	0					0	305	5
251-601	0	0		0	0	0	0	0	0		0		
251-701	0	0			0		0	0	0		0		
251-702	0	0			0	0	100	0	87		27		
251-703					0	0	0	0	0		0		
251-706									3		3		
251-821	0			500	15		1,800	2,000	0	0	616	0	)
251-822 <sup>C</sup>							61,193	21,670	199,211	29,093	77,092	106,347	7 🔸
251-823					0		0	0	0	0	0		
251-824	0				0		0	0	0		0		
251-825						0	138	0			46		
251-826				0	10		73	0	0		17	C	}
251-827					0	12	452	800	5,000	0	1,044		

Appendix A.2. Pink salmon aerial survey odd-year peak counts, Kodiak Management Area, 1969-1987.

Appendix A.2.	(page 2 of 11)

						Peak Cou	int						
Stream No.	1969	1971	1973	1975	1977	1979	1981	1983	1985	1987	Mean (1969-87)	1989	Index Stream
251-828	100			0	0	0	581		375		176		
251-829						0	1,205	0	0		301	(	D
251-830	25,000	25,000	11,000	21,000	21,600	71,000	56,000	18,000	33,145	12,094	29,384		*
251-831								533	26,904	202	9,213		
251-901	0	3,323		1,688	14,314	10,000	29,000	5,500	15,000		9,853	1,500	o 🔹
251-901A						0					0	·	
251-902							500	0	3,000		1,167		
251-903A								0		0	0		
252-101		300		90		300	4,400	0	9,000		2,348		
252-102		50		0	0	700	8,500	0	11,000		2,893		
252-301									10,000		10,000	1,500	D
252-302								100		0	50	50,000	Э
252-305								85		0	43	3,000	0
252-306								2,550	15,000	0	5,850	150,000	
252-307								213	5,000	0	1,738	60,000	
252-308										0	0	3	3
252-309							6,000			800	3,400	60,000	D
252-311	0			25		0	500				131		
252-312	50			90	563	284	1,200				437		
252-314A					500						500		
252-315	25			200		90	1,200				379		
252-316	250	300		438	718	1,062					554		
252-317	0			0	1,409	1,012	100			0	420	1	1
252-318				0	208	4	0		0		42	150,000	0
252-319	0			0	25	0		0		0	4	Ċ	2
252-321	0			18		850	0				217		
252-323					0	300		0	10,000	0	2,060	150,000	C
252-324		3,800		8,643	6,661	15,988	40,834	0	25,000	153,482	31,801		
252-331	600		200	0	158	3,200	18,500	1,096	200	1,200	2,795	2,356	5
252-332	7,600	5,800	3,000	2,100	21,000	24,000	44,029	11,000	13,000	2,700	13,423	30,000	) *
252-333	0			127		512	0	0	-	900	257	15,000	0
252-335			100	0	0		0	0		0	17	189	9
252-341	0			0	3	799	1,000	0			300		
252-342 <sup>D</sup>	48,000	13,600	38,000	2,800	84,000	19,020	17,508	84,956	8,860	8,780	32,552	41,611	1 *
252-343	12,000	12,500	0	0	29,000	41,500	38,700	34,500	9,800	7,100	18,510	10,000	• 0

Peak Count													
Stream No.	1969	1971	1973	1975	1977	1979	1981	1983	1985	1987	Mean (1969-87)	1989	Index Stream
Northwest .	Kodiak Di	strict				<b>*</b> • • • • • • • • • • • • • • • • • • •							
253-114	100		0		0	4	250	50			67	2,500	0
253-115	0		2,000		11,500	36,000	82,000	32,000	18,300	41,000	27,850	84,000	b +
253-121	500	1,500	1,000	0	50	1,600	6,000	1,600		16,000	) 3,139	8,000	0
253-122	50,000	37,000	53,000	77,500	44,000	99,000	75,000	128,000	114,000 :	190,000	86,750	520,000	*
253-123	0				0		0	. 0			. 0	,	
253-141								50			50	1,800	0
253-201										O	) 0		
253-203										0	) 0		
253-311							100	350			225	1,500	0
253-313			0		362	60	0	150	1,200	0	253	12,000	0
253-321	1,500	2,500	3,000		325	4,800	11,700	1,400		2,600	) 3,478	6,000	0
253-322				20	2	375	0	0			79		
253-323			0		362	60	0	0		C	) 70		
253-331	46,000	40,000	22,000	43,500	56,000	80,000	92,000	42,250	80,000	72,000	57,375	367,000	• 0
253-332	23,000	6,000	3,000	550	5,900	18,100	44,500	8,100	21,000	15,000	) 14,515	325,000	0 •
253-333					35	0	0	1,200	4,100	100	906	800	0
253-351	0					0		0			0		
253-352	200					80		0			93		
254-103			0		300	1,800		190	80		474	2,000	0
254-104					12,075		600				6,338		
254-105							3,000	700	30		1,243	1,400	0
254-201	400		10	13,500	13,500	19,000	15,000	6,000	4,500	1,600	8,168	60,000	0
254-202	103,000	65,000	50,000	72,000	111,000	147,000	136,000	180,000	175,500	96,500	) 113,600	300,000	0 •
254-203	13,000	30,000	4,000	24,000	24,130	58,000	58,000	62,500	80,000	30,000	38,363	165,000	0 *
254-204	2,600	400	800		4,500	5,200	32,000	4,500	7,100	11,000	7,567	34,000	0 •
254-205			0	1,200	1,950	400	0	0	0	C	) 444	500	0
254-206				500	1,350	300	0	0	0	C	) 307	1,000	0
254-207			0		8,500	1,500	0	0	40	300	1,477	1,000	0
254-208					1,480	350	300	60	30	500	) 453	1,000	0
254-209					640		0	0	0		160	1,500	0
254-210					1,900		0	0	0		475	1,500	0
254-211					560		0	0			187	5,000	0
254-212					390		0	600	2,300		823	2,500	0
254-213								220	1,500		860		
254-301	18.000	14.000	11,000	35,500	35.500	43.000	50.000	30.000	4.000	30.000	27.100	95.000	n •

Appendix A.2. (page 3 of 11)

-Continued-

						Peak Cou	int						
Stream No.	1969	1971	1973	1975	1977	1979	1981	1983	1985	1987	Mean (1969-87)	1989	Index Stream
254-302	0		5,000	0	3,200	6,000	300	0		C	) 1,813	3,000	)
254-401	2,000	20,000	0		45,000	18,000	0	30,000	0	C	) 12,778	48,000	2
254-402						400	100	40			180	2,500	)
254-403					400	150	50	20			155	100	)
254-404							2,800	7,100	750	450	) 2,775	1,500	)
254-405							2,300	100			1,200	400	נ
255-102			80		7,770	129		0			1,995		
259-362					125						125		
259-363	0		0	0	0	0	200	200	0	C	) 44	C	)
259-364	1,400		50		400	12,000	4,500	400	2,500	2,100	) 2,919		
259-365	8,300	4,000	8,000	11,000	19,300	29,600	58,250	18,000	35,000	29,000	22,045	236,000	)
259-366	1,500		300		150	16,500	4,300	600	1,600	4,700	3,706	8,700	)
259-367					4		120			1,400	508		
259-368					73	2,700	20			50	) 711	7,000	)
259-369					250		0			200	) 150	-	
259-371	11,200	6,400	4,000	12,000	18,000	38,000	20,500	18,000	40,900	15,000	18,400	39,000	} *
259-372						60	0	200		2,000	565	4,500	)
259-373							20				20	1,800	)
259-381						0	6,000	0		0	1,500	0	)
259-382	4,700	2,300	1,800	2,650	3,500	16,200	4,000	2,800	6,600	3,780	4,833	54,000	)
259-383	1,700	200	500	0	500	3,800	0	2,900		800	1,156	. 0	)
259-391	1,500		500	400	300	8,600	20,000	5,500	7,400	2,000	5,133	5,400	)
259-392	2,500		200	900	2,500	1,300	6,500	0	3,000		2,113	4,000	)
259-393	200		0	150	200	20	250	600	300	0	) 191		
259-394								700	5,400	1,900	2,667	3,200	)
259-395								0		900	450	286	5
259-397									4,500		4,500	300	)
Southwest R	odiak Di	strict											
arr 101a	<u>^</u>												
255-101 256 2018	2 100	0	57	210	52,319	82,973	51,248	38,902	41,232	24,222	32,351	109,880	) •
256-201-	2,100	0	839	2,568	3,/16	10,878	4,358	17,702	3,000	7,819	5,298	_	*
256-202			0	0	0	10		1,400			282	0	)
200-301			0	0	3,500	3,000	13,000	3,000	2,600	3,440	3,568	1,500	)
256-302			0	0	0	0	50	800		0	121	0	)
200-303		4 000	U	~	0	0	0	1,100	10,500	0	1,657	0	}
256-401 256 402	0	4,000	0	0	2,000	45,000	0	0	4,500	0	5,500	1,250	) •
200-402		U		U	5,000	400		4,500	5,000		2,483	3,000	)
						Contin	und	<b>_</b> _					

						<u>Peak Co</u>	unt						
Stream No.	1969	1971	1973	1975	1977	1979	1981	1983	1985	1987	Mean (1969-87)	1 1989 S	index tream
Alitak Bay	District												
257-101		0			0	0		0		C	0	0	
257-102	0	0			0	0	0	0	0	0	0	1.000	
257-301	0					1,500	0	Ó	-	_	375	_,	
257-302	14,000		742	9,494	13.380	100	Ó	9.000	2.500	22.791	8.001	49 608	
257-303	0			0	2,500	0	õ	600	_,		388	600	
257-304	35,000	15,000	2,197	22.000	18,000	5.277	1,506	424	5.235	1.010	10.565	000	
257-305	·	•	,	,	.,	-,	_,	10	0,200	1,010	10		
257-401	6,000	5.000	1,000	200	2.500	3.830	16.000	3.000	8.000		5.059	550	•
257-402	25	400	2,777	2,000	7.000	6,000	3,000	1.000	0	2.200	2,440	300	
257-403	60,000	63,000	22,000	38,000	98,000	85,178	157.000	72,000	141.000	55,993	79,303	315 559	*
257-404	19,000	19,000	8,000	3,500	,	17,000	50	5,000	,	,	10,221	5157005	
257-405	5,000			4,500	4,000	2,400	4,000	4,500			4,067	6.000	
257-406					300	•	3,500	700			1,500	0,000	
257-407					400		. 0	24			141		
257-409							0	0			0		
257-501			0	10	100			0	800		182		
257-502	140,000	100,000	40,000	76,700	141,000	97,500	180,000	148,000	166,500	130,700	122,040	410,000	*
257-503	15,000	4,000	500	340	8,000	7,000	13,000	8,200	9,000	6,100	7,114	87,000	
257-504			0	0		3,500		2,000	350		1,170	32,000	
257-505								0			. 0	,	
257-601	1,500	0	450	0	2,800	4,500	50	2,000	40,000	0	5,130	12,000	
257-602	1,700	4,000	400	900	3,000	6,000	3,800	8,700	35,000	25,000	8,850	35,000	
257-603	6,000	1,000	4,000	200	3,000	10,500	2,500	10,100	48,300	24,500	11,010	55,000	
257-604	2,000	400	1,000	2,400	4,500	6,000	800	15,500	58,000	0	9,060	123,000	
257-605	0	0	0	0	0		0	1,000	2,000		375	7,000	
257-701	44,000	122,000	45,000	71,300	92,000	306,000	240,000	142,000	273,000	208,500	154,380	690,000	٠
257-702			500	4,200	10,500	900		1,600	8,000	0	3,671	94,000	
257-703	0			0	600	6,000		5,000	90	0	1,670	0	
Bastside Ko	odiak Dis	trict											
258-101	1,100	200	0			6.300	9.000	200	1.100		2 557	2 600	
258-201	2,000		Ō	0		3,500	2,000	800	5,200	100	1,700	4 700	
258-202	. 15	100	250	ň		1,000	2,000 N	200	2,200	400	2,.00	200	
258-203			•	วด้		1,000	ő	~ ~ ~ ~ ~ ~ ~ ~ ~ ~ ~ ~ ~ ~ ~ ~ ~ ~ ~ ~	,0	100	200	2.00	

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													- 1
Stream No.	1969	1971	1973	1975	1977	1979	1981	1983	1985	1987	Mean (1969-87)	1989 :	Index Stream
258-204	0	200	0	100		0	1,000	200	0	C	167	200	)
258-205		0	300	0		0	0	0	100	3,300	463	300	)
258-206	2,500	3,000	700	200	1,350	4,400	9,000	6,000	13,400	6,080	4,663	4,100	)
258-207	500	2,500	1,500	3,700	2,700	22,200	47,000	8,800	12,600	11,180	11,268	60,500	)
258-208			0	0	200	0	0	0	100	0	38	900	)
258-209	1,000	50	500	0		1,200	3,100	2,500	1,200	100	1,072	1,600	)
258-210	0	0	0	0		0	5,000	0		0	625	0	)
258-211	0	1,000	10	0	400	0	0	0	50	0	146	400	)
258-212	0		0	0	0	0	0	0	0	0	0	4,000	)
258-213						150	0	0	0	C	30	28,500	)
258-301	0				0	0	0	600	500		183	2,800	)
258-302	0	200	10		2,600	0	0	300	400	0	390	3,500	)
258-303	0		0	0	0	0	0	0	0		0	2	2
258-304	1,100		20		13,000	11,000	6,300	4,200	2,800	200	4,828	3,000	}
258-305	0		0	0	200	700	75	0	0		122	575	5
258-306	0	0	0	0	0	0	0	0			0	150	}
258-307						0	0	0	0	0	0	1,100	}
258-401	500	0	1,000	0	500	0	0	0	400	0	240	0	)
258-402					580	2,000		150	20		688		
258-403									800		800		
258-404									900	0	450		
258-511	2,200	5,400	300	0	600	3,500	8,600	3,000	3,600	0	2,720	0	)
258-512	0		100		400	3,000	0	0	300	0	475	11,000	)
258-513	0		50		2,760	0	0	100	700	0	451	0	)
258-514							0	500	1,200	0	425	10,800	)
258-515							8,000	0	0	0	2,000		
258-516									0	0	0		
258-521	1,900	1,900	5,000	13,500	28,000	6,000	22,000	22,000	15,000	3,400	11,870	15,000	)
258-522	35,000	23,000	23,000	39,400	48,500	67,500	52,000	78,000	77,000	14,100	45,750	53,300	) ;
258-523	1,500	0	0	0	30	100	0	700	50		264		
258-531	0			0	60	0	5,000	400	1,100		937	0	)
258-532	0			0	0	80	0	200	25		44	0	)
258-533	0				0	0	0	25		0	4	0	)
258-541	4,000	3,000	150	250	7,000	13,000	11,000	6,000	3,200	500	4,810	30,500	)
258-542	25,000	1,500	1,100	1,600	13,000	17,000	22,000	23,100	21,000	11,000	13,630	78,000	) 1
258-543	0	0		0		0	4,000	0	10		573	-	
258-544						0	1.200	500	3.800	2.200	1.540	17.000	)

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						Peak Co	unt						
Stream No.	1969	1971	1973	1975	1977	1979	1981	1983	1985	1987	Mean (1969-87)	1989	Index Stream
258-551	7,000	0	3,000	0	400	500	34,000	2,000	1,800		5,411	2,60	0
258-552	3,800		3,000		14,000	11,500	17,000	4,000	. 0	5,360	7,333		
258-553	8,000	0	200		15,000	9,500	16,000	6,000		850	6,944		
258-554	100		0	0	600	300	2,500	500			571	1,00	0
258-555	25		0		50	0	, 0	200			46	30	ō
258-601	400			0	200	200	0	40	3,000		549	4.00	0
258-602	22,000	10,000	1,500	2,500	9,000	33,000	2,500	30,800	10,000	8,800	13,010	50,40	ō
258-603	5,200	. 0	13,600	0	10,000	9,000	. 0	10,000	12,000	480	6,028	42.00	ò
258-701	39,000	54,000	24,000	86,000	53,000	100,700	128,000	86,000	60,000	164,400	79,570	450,00	0 *
258-702	8,000		0	0	9,000	8,000	10,600	4,000	5,200	8,580	5,931	68,00	0
258-703					6,000	10,000	5,600	18,000	400	,	8,000	47,00	0
258-704					0			800	0		267		0
258-705							0		6		3	19,00	0
258-706							10,000				10,000		0
258-801	17,000		600	0	4,500	1,200	800	0			3,443		0
258-801A				20		0	0				7		
258-801B							3,400				3,400		
258-802				0	20	0			200		55		
258-803				0	0	2		0	0		0	1	0
258-803A		0		0	0	15					4	1	0
258-804		0		0	90	0		0	0		15		0
258-804A		0		0	0	0					0		
258-805		0		0	0	0		0	0		0	1	0
258-806				0	0	0		0	0		0		
258-807									13		13	1	0
258-808				0	100	100	0	450			130	10	0
258-809				0		20		200			73	i	0
258-810				0		700					350	i	0
258-811				0							0		
258-851								1,500	500		1,000		0
258-852									40		40	10	0
258-853									0	0	) 0		
258-901								200	1,000		600	1,00	0
258-902								6,000	8,000		7,000	17,00	0
258-903								150	3,000	0	1,050	3,00	0
258-904								0	0		0	2,00	0
259-401	20,000	21,000	15,000	14,000	0	23,000	3,800	5,400	11,800	150	) 11,415	3,60	0

			<u>1</u>			Peak Cou							
Stream No.	1969	1971	1973	1975	1977	1979	1981	1983	1985	1987	Mean (1969-87)	1989	Index
			1979	1973		1979	1901		1905		(1)0)-0//	1989	Scream
259-402	10	0	0	0		0		120	0	(	0 16		
259-403	0	60	100	100		0		20	2,400	(	335	100	0
259-404	0	130	0	50		200		0	10	(	3 49		
259-405	0		0			0		0	0	(	0 C		
259-410	2,100	7,000	2,900	12,000	100	14,000	4,100	3,800	6,600	12,000	6,460		
259-411	10,200	1,500	0	260	100	5,000	2,000	400	14,500	2,000	3,596	2.000	0
259-412	4,000	3,000	1,100	12,000	11,900	71,000	15,380	17,500	40,000	22,500	0 19,838	61,000	0 *
259-413	0	0		300		0			10		62		
259-414	20,000	30,000	8,300	7,700	33,000	17,000	6,050	3,500	1,500	11.100	13,815	96,000	0 *
259-414A									50		50		-
259-415 <sup>d</sup>	85,000	57,000	25,000	46,000	144,000	68,000	57,000	39,000	28,400	39,687	7 58,909	81,000	0 *
259-416	10,000	3,000	1,000	6,600		8,600	3,000	7,500	7.700	· (	5,267	6,500	0
259-417			0	20		10	0	500	400	C	133	500	0
259-418A	100	0	100	0	0	0	1,000	3,000	9,000	C	1,320	900	0
259-418B	0	0	0	0	0	0	1,100	1,000	20	C	212	(	0
259-419	0		0	0	0	0	0	0	80	200	31	(	0
259-419A						445	0	0	250	(	) 139		-
259-420				0			0				0	(	0
259-421	0			0		2,400	0	10	30		349	(	0
259-422	5,000	5,000	2,000	4,000		7,800	5,500	9,000	7.200	630	5,126	20.200	0
259-423	2,100	2,000	0	6,000		3,000	500	4,000	2,000	1,700	2,367	700	0
259-424	3,600	1,800	3,600	5,400	22,000	57,300	14,400	6,400	4,550	11,000	13,005	46.000	D
259-424A					6,200			,	,	,	6,200	,	-
259-425	600	0	0			800	0	300			283	6,300	0
259-426	0	0	100	0		500	0	0		75	5 84	800	0
259-427			0			100	200	0	0	C	50	200	D
259-428	0	550	3,500		0	0	0	0	0		506	(	Ď
Northeast	Kodiak Di	strict											
259-101	3,500	0	50	600	3,800	3,300	1,300	1,250	8,700	295	5 2,280	8,000	0
259-101A	300		500		200	400	2,500	650	2,800		1,050		
259-102	500	500	1,000	4,500	4,800	850	400	430	5,040	300	1,832	42,000	0
259-103									-	C	, 0		
259-211	66,500	7,900	26,350	22,000	54,000	61,000	88,000	53,000	166,688	27,892	2 57,333		*
259-221	3,000	2,000	1,000	-	300	1,400	1,400	300	4,000	300	1,522	22,000	0
259-222	16,340	3,100	2,250	6,000	12,000	24,200	5,600	2,000	10,400	18,200	10.009	36,500	0
	•	•				,		_,					-

						Peak Cou	int						
Stream No.	1969	1971	1973	1975	1977	1979	1981	1983	1985	1987	Mean (1969-87)	1989	Index Stream
259-223 259-224	15,100	4,400	1,500	11,250	6,000	19,550	10,000	5,500	20,400	10,000	10,370	113,00	0
259-231 259-232	52,000	18,900	11,300	0	51,900	45,300	54,000	71,000 0	141,000	112,800	55,820	152,00	0
259-233 259-234 259-235		1,000		20	200	0	0 5,500	0	0	1,000	278 5,500 250	50	0
259-242 259-243	37,000 5,000	40,000 300	13,000	16,000 400	41,000	80,000 1,100	73,000 0	30,020	69,400	66,100 0	46,552	46,00	0
259-244 259-245 259-246					0			0 1,390	4,600 0	0 0 0	0 1,997 0	12,600 10,200 300	0 0 0
259-251 259-252 259-253	13,000 300	4,000	600 50	5,600 0 300	4,200 500	11,500 500	1,500 0 0	2,800 260 0	7,800 60 300	12,000 14,000 10	6,300 1,841 122	39,40	0
259-254 259-255	4,000 1,700	1,500 4,000	200 300	1,300 3,200	2,700 11,000	4,000 60	650 2,600	3,000	10,500 0	9,950 0	3,780 2,540		
Mainland D	istrict												
262-101 262-102 262-103		0 0				200 600 0	50 0 30	100 300 50	900 0 2,200	0	250 150 380	10) 12,000 600	0 0 0
263-104 262-105 262-106		0				0	0	0	0	0	0 0		0
262-106A 262-107		0				4,000	0	0	1,500	0	1,500		0
262-108 262-151 262-151	0	50 100	0	7,400	0	300	0	0	600	0	158 750	3,60	0 0
262-151A 262-152 262-152A	70,000	15,000	15,000 0	11,000 0	40,000 500	25,000	40,000	11,000	25,000	22,000	27,400 167	15,00	0
262-153 262-153A 262-153B	99,000	20,000	500 0	27,000 0	70,000 0 250	19,050	42,000	33,000	15,000	9,000	33,455 0 250	48,00	0
262-154	0	0	0	40	300	0	0	0	0	0	34	l.	0

						<u>Peak</u> Cou	nt						_
Stream No.	1969	1971	1973	1975	1977	1979	1981	1983	1985	1987	Mean (1969-87)	1989 \$	Index Stream
262-155				0	50		+ _ + _ + _ + _ + _ + _ + _ + _	·····	0	100	) 38	800	)
262-156				150	30						90		
262-157				0	0						0		
262-158				0	0						0		
262-159				0	200						100		
262-201	25	0	0	0	2,000	600	0	200	200	3,000	) 603	3,000	)
262-202	0	0	6,000	0	0	0	0	0	0	200	) 620	0	)
262-203	10,000	0	1,500	0	3,000	0	1,500	2,000	2,000	C	2,000	0	) *
262-204	500	0	0	0	600	100	1,100	300	0	C	) 260	0	)
262-205	2,000	1,000	500	2,500	3,500	6,000	11,000	450	7,700	3,400	) 3,805	42,000	) *
262-253	0	0	0	0	0	0	50	0	1,700		194		
262-254	20	100		0		0	0	20	300	C	) 55	100	)
262-255	1,500	0	50	0	100		0	220	75		243		
262-256	0	0	50	0	100		0	16	150		40		
262-271	0	0	0	15,000	5,000	3,000	2,000	500	2,000	1,000	2,850	2,000	) *
262-272	0	0	0	0	6,000	300	500		900	C	) 856	0	) *
262-301	0	0	0	0	3,600	0	0	0	0	C	) 360	0	}
262-302							0				0		
262-351	0	0	0	0	5,000	3,000	6,000	2,600	2,500	600	0 1,970	9,000	)
262-352						6,000	0	0			2,000	3,800	)
262-401	0	0	0	0	1,500	1,200	9,300	500	2,700	100	) 1,530	3,800	)
262-402	3,000	100	400	4,200	10,500	5,000	14,000	2,000	3,000	C	) 4,220	19,000	) *
262-451	13,000		500	34,000	55,000	55,000	89,000	45,000	45,000	6,500	38,111	37,000	) *
262-452		200	800		0	3,000	200	0	0		600		
262-453							0	1,100	0		367		
262-501	3,000	500	1,000	0	15,000	21,000	51,200	4,500	16,500	14,000	12,670	200,000	) *
262-502					200	0	6,000	200		700	) 1,420	0	)
262-551	18,000	1,000	3,000	29,000	76,000	77,000	31,000	8,000	17,000	18,000	27,800	38,000	) *
262-552						0	0	0	0	C	0	0	}
262-601	0	0			0	0	0	0	0		0	0	)
262-602	2,000	5,000	2,000	0	3,000	0	0	0		2,000	) 1,556	0	)
262-603	200	0	6,000	0	200	0	0	0	1,900	, C	) 830	0	)
262-604	14,000	18,000	9,500	16,000	60,000	38,700	37,000	33,000	45,000	63,000	) 33,420	80,000	) *
262-605					4,300	1,200		0	1,200	700	1,480	10,000	)
262-606							0			(	0 0	20	)
262-651	7,000	6,000	6,000	5,200	28,000	17,000	11,000	12,000	42,000	51,000	18,520	445,000	) *
262-652	1,300	800	500	20	0	300	1,500	2,000	15,000	28,000	) 4,942	162,000	)

						Peak Cou	int						
Stream No.	1969	1971	1973	1975	1977	1979	1981	1983	1985	1987	Mean (1969-87)	1989	Index Stream
262-653	1,200	2,500	500	10		250	5,100	2,800	7,600	18,000	) 4,218	230,000	)
262-654	3,300	5,000	2,500	2,000	9,000	0	0	0	2,000	1,100	2,490	. (	)
262-655		6,000				700	0		•	1,400	2,025	300	)
262-656							0			600	300	13,600	)
262-701	250	500	300	450	2,000	500	2,000	2,200	7,000	5,400	2,060	44,000	)
262-702	4,300	2,000	3,000		7,000	12,000	14,000	6,800	16,000	31,000	10,678	67.000	- ) *
262-702A	0	0	0	0	0	. 0	0	. 0	, 0		) 0	,	
262-703	0	0	0	120		0	0	0	0	C	) 13	42.000	)
262-704	50	0	0	50	300	0	5,000	1,600	0	2.000	900	20,000	)
262-705		0		0		300	250	1.200	2,000	4,200	1.136	12,000	)
262-750		0		0		300	250		•	,	138	,	-
262-751	4,000	2,500	5,000	2,200		5,700	15,000	3,100	31,000	13,000	9,056	73,000	) *
262-752	1,500	0	0	3,500		400	0	200	0	4,700	) 1.144	10.000	)
262-801	600	1,000	300	80	1,500	1,200	500	800	1,200	2,600	) 978	22,000	) *
262-802	16,000	6,000	2,000	3,000	10,500	40,000	50,000	6,500	14,000	21,000	16,900	76,000	) *
262-803	0	100		0	200	200	. 0	, 0	150		) 72	500	}
262~804										C	) 0	200	)
262-851	21,000	4,000	5,000	13,000	112,000	184,000	65,000	40,000	69,000	141,000	65.400	1.320.000	) *
262-851A										Ċ	) . 0		
262-852	2,000	600	0	2,600	4,000	4,500	16,000	5,000	11,000	12,500	5,820	91,000	)
262-853	3,800	0	0	200	600	500	2,500	6,000	4,200	8,200	2,600	66,000	)
262-854	500	0	0	0	300	100	20	50	2,800	8,600	) 1,237	10,000	)
262-854A					25						25	•	
262-855	0	0		4	0		0	15		C	) 3		
262-856	500	0	0	0		6,200	3,200	4,000	8,000	3,200	) 2,789	100	)
262-856A						2,600		3,000	7,800	19,000	8,100		
262-856B						400		200	400	1.00	275		
262-857	0	0	0	10	0	0	0	0			1		
262-858	0	5,000	0	8,700	0	2,000	0	0	0	5,000	2,070	1,900	)
262-859	0	6,000		0	300	200		500	0		1,000		)
262-859A					75	0		0		4,500	) 1,144	-	
262-951					650					-	650	24,000	)
Reference of the second s													

Appendix A.2. (page 11 of 11)

 <sup>a</sup> Total escapement counted through fish weirs.
 <sup>b</sup> Afognak River fish weir was moved downstream in 1986. Counts preceeding were adjusted to account for fish spawning between the two weir sites.

<sup>c</sup> Little Waterfall (251-822) escapement counts pre-1981 were excluded owing to additional spawning habitat created by a fishpass.
 <sup>d</sup> Saltery Creek (259-415) escapement counts reflect estimates of fish spawning downstream of the weir.

STREAM	1965	1967	1969	1971	1973	1975	1977	1979	1981	1983	1985	1987	Average 1965-1987	90% CI 1965-1987	1989
Alinchak					0.0		60.2	0.0	2.0	49.0	0.9		18.7	4.2-50.4	27.3
American	8.8	85,4	150.4	93.2	83.6	219.1	361.4	33.3	174.6	0.0	126.6	378.8	142,9	78.4-207.4	256.0
Barling		13.4	36.0	0.0	20.1	63.3	107.3	39.0	168.2	289.9	147.7	210.3	99.6	48.2-150.9	51.7
Bauman's	119.4	353.8	107.0	113.0	108.1	317.2	375.9	401.5	188.1	252.9	560.4	863.2	313.4	196.5-430.2	240.7
Beaver									1,037.0	493.9	171.4	192.1	473.6	0-948.3 <sup>a</sup>	549.4
Big Creek					4.7		138.9	36,5		0.0		395.0	115.0	0-273.5 <sup>a</sup>	131.8
Buskin	64.7	11.5	46.3	35.9	31.8	51.5	164.3	141.8	374.8	277.5	248.4	267.6	143.0	79.7-206.3	89.8
Dakavak					85.1	19.5	111.2	0.3	85.8	0.0		117.0	55.6	21.9-97.7	0.4
Danger		157.6	31.1	227.9	0.0		156.0	16.2	67.0	10.2	56.9	174.5	89.7	42.4-137.1	292.5
Deadman	283.8	403.5	351.1	725.4	218.7	532.2	691,5	238.3	755.2	242.4	349.6	400.3	432.7	330.9-534.4	546.9
Dog Salmon	73.8	254.7	46.7	88.2	0.0	83.0	67,2	21.2	49,6	120.4	50.1	400.3	104.6	46.0-163.2	179.9
Нитру	5.1	281.7	80.2	144.3	88.7	422.4	508.0	35,9	228.1	220.1	135.1	491.1	220.0	126.1-313.9	464.8
Hurst			96.3	172.7	22.7	12.7	69.5	34.6	0.0	0.0	0.1	84.9	49.4	16.7-82.1	102.9
Kaiugnak	60.4	908.8	68.4	0.9	21.6	213.6	768.2	9.8	295.2	55.5	421.4	72.2	241.3	81.6-401.1	736.8
Kanatak							48.2	124,9	452.9	159.2		83.6	173,8	19.8-327.8	186.7
Karluk															0.1
Kashvik					5.3	12.5	24.0	20.5	2.3	164.9		82.1	44.5	0.8-88.2	0.8
Kiliuda			81,6	14.1	3.6	19.3	0.8	11,8	90.1	5.0	104.5	15.5	34.6	11.5-57.8	77.0
Kinak						136.2	280.4	179.6	105.8	111.1		42.7	142.6	76.1-209.2	161.7
Kukak								106.8	29.3	20.2			52.1	0-132.3 <sup>a</sup>	6.7
Little															5.6
Little Wtfl.									586.2	839.5	339.0	172.3	484.2	141.2-827.3	70.0
Marka															10.3
Miam															118.6
Missak						130.1	126.0	59.4	63.3	3,6		214.6	99.5	39.1-160.0	131.8

Appendix A.3. Odd-year pink salmon preemergent fry densities (live fry/m<sup>2</sup>) by stream and brood year, Kodiak Management Area, 1965-1989.

1903	1967	1969	1971	1973	1975	1977	1979	1981	1983	1985	1987 1	965-1987	1965-1987	1989
						3,6	21.0	50.6	119.8	3.3		39.6	22.9-56.4	2.7
77.3	225.3	123.0	47.2	0.0	70.5	185.8	4.8	27.6	107.0	96.8	135.9	91.8	55.9-127.6	264.0
	0.0	2.8	27.9	15.2	6.1	13.3	26,1	85.5	44.6	5.7	5.5	21.2	7.4-34.9	278.3
137.9	363.2	428.2	347.0	146.9	306.6	111.4	234.2	251.9	46.1	216.9	39.2	219.1	153.6-284.6	72.3
						73.2	0.0	0.0	4.8			19.5	0-61.7 <sup>a</sup>	2.3
35.2	76.3	33.9	34.2	72.4	126.6	194.8	81.5	163.2	46.7	44.4	227.5	94.7	59.7-129.7	521.5
						616.0	99.0	331.5	89.8	582.2	171.7	315.0	120.3-509.8	869.8
115.1	210.0	304.7	325.2	137.5	338.0	260.4	173.7	595.8	135.0	569.2	252.0	284.7	202.8-366.7	383.5
10.1	67.4	92.0		46.6	15.8	89.4	21.9	20.9	7.0	68.7	7.1	40.6	21.5-59.7	39.8
1.4	100.2	163.6	134.2	43.7	91.1	31.8	50.1	44.5	183.8	44.4	227.5	93.0	56.6-129.4	116.4
1.2	4.1	7.1	11.9	7.4	47.0	11.7	7.0	21.6	2.0	0.2	16.2	11.4	4.8-18.1	0.8
4.2	96.5	9.9	20.9	166.7	92.1	67.1	154.2	172.9	7.1	0.4	1.4	66.1	30.5-101.7	9.4
59.5	175.2	313.8	187.7	70.9	377.8	258.5	199.1	605.5	233,4	162.5	447.7	257.6	175.9-339.4	1,070.0
						123.0	0.5	340.9	700.3	0.0	80.8	207.6	0-431.3 <sup>a</sup>	238.0
0.0	0.0	0.0	0.0	0.0	3.6	34.7	0.0	45.7	0.0	0.0	38.0	10.2	0.9-19.4	12.1
-	77.3 137.9 35.2 115.1 10.1 1.4 1.2 4.2 59.5 0.0	77.3       225.3         0.0         137.9       363.2         35.2       76.3         115.1       210.0         10.1       67.4         1.4       100.2         1.2       4.1         4.2       96.5         59.5       175.2         0.0       0.0	$\begin{array}{cccccccccccccccccccccccccccccccccccc$	77.3       225.3       123.0       47.2         0.0       2.8       27.9         137.9       363.2       428.2       347.0         35.2       76.3       33.9       34.2         115.1       210.0       304.7       325.2         10.1       67.4       92.0       1         1.4       100.2       163.6       134.2         1.2       4.1       7.1       11.9         4.2       96.5       9.9       20.9         59.5       175.2       313.8       187.7         0.0       0.0       0.0       0.0	77.3 $225.3$ $123.0$ $47.2$ $0.0$ $0.0$ $2.8$ $27.9$ $15.2$ $137.9$ $363.2$ $428.2$ $347.0$ $146.9$ $35.2$ $76.3$ $33.9$ $34.2$ $72.4$ $115.1$ $210.0$ $304.7$ $325.2$ $137.5$ $10.1$ $67.4$ $92.0$ $46.6$ $1.4$ $100.2$ $163.6$ $134.2$ $43.7$ $1.2$ $4.1$ $7.1$ $11.9$ $7.4$ $4.2$ $96.5$ $9.9$ $20.9$ $166.7$ $59.5$ $175.2$ $313.8$ $187.7$ $70.9$ $0.0$ $0.0$ $0.0$ $0.0$ $0.0$	77.3 $225.3$ $123.0$ $47.2$ $0.0$ $70.5$ $0.0$ $2.8$ $27.9$ $15.2$ $6.1$ $137.9$ $363.2$ $428.2$ $347.0$ $146.9$ $306.6$ $35.2$ $76.3$ $33.9$ $34.2$ $72.4$ $126.6$ $115.1$ $210.0$ $304.7$ $325.2$ $137.5$ $338.0$ $10.1$ $67.4$ $92.0$ $46.6$ $15.8$ $1.4$ $100.2$ $163.6$ $134.2$ $43.7$ $91.1$ $1.2$ $4.1$ $7.1$ $11.9$ $7.4$ $47.0$ $4.2$ $96.5$ $9.9$ $20.9$ $166.7$ $92.1$ $59.5$ $175.2$ $313.8$ $187.7$ $70.9$ $377.8$ $0.0$ $0.0$ $0.0$ $0.0$ $0.0$ $3.6$	$\begin{array}{c ccccccccccccccccccccccccccccccccccc$	$\begin{array}{cccccccccccccccccccccccccccccccccccc$						

<sup>a</sup>Zero was assigned to lower confidence intervals which were negative numbers.

• · · ·		Even Yea	r Indexed		Od Od	<u>d-Year Escape</u>	ment goals	
<u>Index Stream</u>		<u>    Escapeme</u>	nt qoals	Index	ed"	ES	timated Total	
Name	No.	Minimum	Desired	Minimum	Desired	Minimum	Desired	Mid Point
Afognak Distr	ict							
Malina	251-105	20,000	60,000	5,000	15,000	9,200	27,600	18,400
Paramanof	251-404	10,000	30,000	5,000	15,000	9,200	27,600	18,400
L. Waterfall <sup>b</sup>	251-822	15,000	45,000	15,000	45,000	15,000	45,000	30,000
Discoverer	251-830	20,000	60,000	20,000	60,000	36,800	110,500	73,700
Pauls Bay <sup>b</sup>	251-831	3,000	9,000	3,000	9,000	3,000	9,000	6,000
Seal Bay	251-901	5,000	15,000	5,000	15,000	9,200	27,600	18,400
Big Danger	252-332	15,000	45,000	10,000	30,000	18,400	55,200	36,800
1arka	252-334	30,000	90,000	10,000	30,000	18,400	55,200	36,800
Litnik <sup>b</sup>	252-342	30,000	90,000	10,000	30,000	10,000	30,000	20,000
Subtotal		148,000	444,000	83,000	249,000	129,200	387,700	258,500
Northwest Dis	trict							
Sheratin	253-371	15,000	45,000	10,000	30,000	18,400	55,200	36,800
Bauman's	253-333	5,000	15,000	5,000	15,000	9,200	27,600	18,400
Ferror	253-331	40,000	120,000	30,000	90,000	55,200	165,800	110,500
Jganik	253-122	80,000	240,000	70,000	210,000	128,900	386,800	257,800
Little River	253-115	40,000	120,000	15,000	45,000	27,600	82,800	55,200
Zachar	254-301	40,000	120,000	20,000	60,000	36,800	110,500	73,700
Browns Lgn.	254-204	40,000	120,000	5,000	15,000	9,200	27,600	18,400
Jyak-202	254-202	50,000	150,000	50,000	150,000	92,100	276,300	184,200
Jyak-203	254-203	5,000	15,000	15,000	45,000	27,600	82,900	55,200
Subtotal		315,000	945,000	220,000	660,000	405,000	1,215,500	810,200

Appendix A.4. Pink salmon escapement goals for odd and even year populations for Kodiak Management Area index streams employing peak counts and estimated total escapement.

-Continued-

		Even Ye	ar Indexed		Od	ld-Year Escape	ment goals	
Index Stream	ı	Escapem	ent qoals	Inde	xed <sup>a</sup>	Es	timated Total	
Name	No.	Minimum	Desired	Minimum	Desired	Minimum	Desired	Mid Point
Southwest Dis	trict							
Karluk <sup>b</sup> Sturgeon Ayakulik <sup>b</sup>	255-101 256-401 256-201	800,000 50,000 400,000	1,600,000 150,000 800,000	20,000 5,000 5,000	60,000 15,000 15,000	20,000 9,200 5,000	60,000 27,600 15,000	40,000 18,400 10,000
Subtotal		1,250,000	2,550,000	30,000	90,000	34,200	103,600	68,000
Alitak Bay Di	strict							
Narrows Dog Salmon Deadman Humpy	257-401 257-403 257-502 257-701	2,000 50,000 40,000 70,000	6,000 150,000 120,000 210,000	2,000 60,000 60,000 90,000	6,000 180,000 180,000 270,000	3,700 60,000 110,500 165,800	11,000 180,000 331,500 497,300	7,400 120,000 221,000 331,500
Subtotal		162,000	486,000	212,000	636,000	340,000	1,019,500	679,900
Northeast Dis	trict							
Sid Olds American Buskin <sup>b</sup>	259-242 259-231 259-211	30,000 30,000 60,000	90,000 90,000 180,000	30,000 30,000 50,000	90,000 90,000 150,000	55,200 55,200 50,000	165,800 165,800 150,000	110,500 110,500 100,000
Subtotal		120,000	360,000	110,000	330,000	160,400	481,600	321,000
Eastside Dist	rict							
Seven Rivers Kaiugnak Barling Kiliuda	258-701 258-542 258-522 258-207	40,000 10,000 30,000 20,000	120,000 30,000 90,000 60,000	40,000 10,000 30,000 10,000	120,000 30,000 90,000 30,000	73,700 18,400 55,200 18,400	221,000 55,200 165,800 55,200	147,300 36,800 110,500 36,800

		Even Yea	ar Indexed			Odd-Year Escar	ement qoals	••
<u>Index Stream</u>	L	<u>Escapeme</u>	ent qoals	Inde	exed <sup>a</sup>	E	<u>stimated Tota</u>	<u>l</u>
Name	No.	Minimum	Desired	Minimum	Desired	Minimum	Desired	Mid Point
Saltervb	259-415	20.000	60.000	30,000	90.000	30.000	90.000	60,000
Miam	259-412	20,000	60,000	10,000	30,000	18,400	55,200	36,800
Hurst	259-414	10,000	30,000	10,000	30,000	18,400	55,200	36,800
Subtotal		150,000	450,000	140,000	420,000	232,500	697,600	465,000
Mainland Dist	rict							
Big River	262-152	10,000	30,000	10,000	30,000	18,400	55,200	36,800 <sup>C</sup>
Village	262-153	15,000	45,000	15,000	45,000	27,600	82,900	55,200 <sup>C</sup>
Cape Chiniak	262-205	5,000	15,000	3,000	9,000	5,500	16,600	11,000 <sup>C</sup>
Big Hallo	262-203	2,000	6,000	2,000	6,000	3,700	11,000	7,400 <sup>C</sup>
Kukak	262-271	3,000	9,000	2,000	6,000	3,700	11,000	7,400
Missak	262-402	5,000	15,000	3,000	9,000	5,500	16,600	11,000
Kinak	262-451	20,000	60,000	20,000	60,000	36,800	110,500	73,700
Geographic	262-501	4,000	12,000	4,000	12,000	7,400	22,100	14,700
Dakavak	262-551	25,000	75,000	20,000	60,000	36,800	110,000	73,700
Kashvik	262-604	25,000	75,000	25,000	75,000	46,000	138,100	92,100
Big Alinchak	262-651	30,000	90,000	20,000	60,000	36,800	110,500	73,700
Portage	262-702	15,000	45,000	10,000	30,000	18,400	55,200	36,800
Oil	262-751	15,000	45,000	10,000	30,000	18,400	55,200	36,800
Jute	262-801	2,000	6,000	1,000	3,000	1,800	5,500	3,700
Kanatak	262-802	10,000	30,000	10,000	30,000	18,400	55,200	36,800
Big Creek	262-851	70,000	210,000	60,000	180,000	110,500	331,500	221,000
Subtotal		256,000	768,000	215,000	645,000	395,700	1,187,100	791,800
Grand Total		2,401,000	6,003,000	1,010,000	3,030,000	1,697,000	5,092,600	3,394,400
Estimated tot	al Escape	ement <sup>d</sup>				2,341,900	7,027,800	4,684,300 <sup>e</sup>

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 <sup>&</sup>lt;sup>a</sup> Non weir systems indexed escapements are peak count derived.
 <sup>b</sup> Systems with escapements enumerated through weirs.
 <sup>c</sup> Index streams for which total available spawning habitat was not estimated; these streams escapement goals were not included in deriving overall spawner density for all index streams.

d Production from these 51 index streams account for 72.70% of the total odd-year pink salmon production founded upon indexed escapements during 1969-1987. Estimated total escapements were generated by expanding the minimum, desired, and mid-point total estimated escapements for index streams by a factor of 1.38 (Barrett et al. 1990).

<sup>&</sup>lt;sup>e</sup> The mid-point estimated total escapement figure minus the escapement figures for those streams that did not have total available spawning habitat estimated divided by the total habitat estimate for all Kodiak Management Area index streams (4,573,900/3,254,454) provides a spawning density of 1.4 fish/m<sup>2</sup> of spawning habitat based upon the mid-point escapement goal for odd-year brood line pink salmon.

	Weired System Counts												
	Af	<u>oqnak</u>	<u> </u>	k	<u>   Barl</u>	Barling		ira	E. Paramanof		Sal	tery	
Date	Daily	Cum.	Daily	Cum.	Daily	Cum.	Daily	Cum.	Daily	Cum.	Daily	Cum.	
/10	1	1											
/11	0	1											
/12	0	1											
/13	0	1									3	3	
/14	2	3									0	3	
/15	0	3									0	3	
/16	0	3							2	2	0	3	
/17	0	3							10	12	2	5	
/18	2	5							23	35	0	5	
/19	0	5							13	48	10	15	
/20	1	6							10	58	0	15	
/21	0	6							0	58	4	19	
/22	16	22							190	248	31	50	
/23	5	27							54	302	6	56	
/24	0	27			10	10			243	545	30	86	
/25	0	27			3	13			171	716	21	107	
/26	1	28			2	15			89	805	0	107	
/27	63	91			8	23			654	1,459	68	175	
/28	370	461			16	39	2	2	223	1,682	59	234	
/29	88	549			19	58	0	2	142	1,824	110	344	
/30	13	562			18	76	0	2	66	1,890	90	434	
/31	93	655			2	78	0	2	1,116_	3,006	41	475	
/1	689	1,344	3	3	0	78	0	2	1,263 <sup>a</sup>	4,269	98	573	
/2	150	1,494	0	3	0	78	0	2	1,263 <sup>a</sup>	5,532	20_	593	
/3	323	1,817	0	3	51	129	0	2	1,263 <sup>a</sup>	6,795	62 <sup>a</sup>	655	
/4	313	2,130	14	17	24	153	0	2	1,263 <sup>a</sup>	8,058	62 <sup>a</sup>	717	
/5	125	2,255	28	45	84	237	0	2	498	8,556	62 <sup>a</sup>	779	
/6	123	2,378	31	76	98	335	0	2	511	9,067	62 <sup>a</sup>	841	
/7	1,124	3,502	94	170	64	399	0	2	1,188 1	.0,255	117	958	
/8	1,995	5,497	283	453	302	701	1	3	95 1	.0,350	239	1,197	

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Appendix A.5. Pink salmon daily weir escapement counts for systems selected for stream life estim	ition, 1990.
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## Appendix A.5. (page 2 of 3)

	Weired System Counts										
	Afoqnak	Pink	Barling	Akalura	<u>E. Paramanof</u>	Saltery					
ate	Daily Cum.	Daily Cu	n. Daily Cum.	Daily Cu	um. Daily Cum.	Daily Cum.					
/9	1,472 6,969	17 47	0 231 932	0	3 27 10,377	226 1,423					
/10	295 7,264	4 47	4 335 1,267	0	3 59 10,436	100 1,523					
/11	1,635 8,899	995 <sup>a</sup> 1,46	5 250 1,517	6	9 0 10,436	100 1,623					
/12	654 9,553	156 1,62	1 260 1,777	6 1	15 0 10,436	55 1,678					
/13	770 10,323	113 1,73	4 323 2,100	0 1	15 0 10,436	117 1,795					
/14	935 11,258	30 1,76	4 370 2,470	6 2	21 117 10,553	258 2,053					
/15	1,181 12,439	172 1,93	5 379 2,849	3 2	24 51 10,604	39 2,092					
/16	3,287 15,726	192 2,12	3 804 3,653	0 2	24 4 10,608	199 2,291					
/17	2,966 18,692	26 2,15	4 237 3,890	7 3	31 71 10,679	27 2,318					
/18	304 18,966	54 2,20	3 200 4,090	28 5	59 79 10,683	137 2,455					
/19	1,447 20,443	163 2,37	L 653 4,743	40 9	99 143 10,826	126 2,581					
/20	1,216 21,659	131 2,50	2 913 5,656	7 10	06 908 11,734	1 2,582					
/21	500 22,159	29 2,53	L 988 6,644	13 11	19 185 11,919	55 2,637					
/22	246 22,405	61 2,592	2 575 7,219	4 12	23 1,470 13,389	95 2,732					
/23	255 22,660	27 2,61	9 829 8,048	4 12	27 830 14,219	115 2,847					
/24	317 22,977	4 2,62	3 447 8,495	60 18	37 643 14,862	85 2,932					
/25	182 23,159	90 2,71	3 839 9,334	112 29	99 369 15,231	62 2,994					
/26	60 23,219	266 2,97	9 703 10,037	368 66	57 690 15,921	118 3,112					
/27	102 23,321	44 3,023	3 652 10,689	463 1,13	30 653 16,574	37 3,149					
/28	68 23,389	40 3,063	3 181 10,870	51 1,18	31 244 16,818	76 3,225					
/29	10 23,399	1 3,064	492 11,362	183 1,36	54 212 17,030	64 3,289					
/30	14 23,413	3 3,06	7 186 11,548	38 1,40	02 478 17,508	154 3,443					
/31	92 23,505	10 3,07	7 539 12,087	176 1,57	78 147 17,655	135 3,578					
/1	155 23,660	1 3,07	3 705 12,792	18 1,59	96 868 18,523	141 3,719					
/2	99 23,759	114 3,19:	2 359 13,151	186 1,78	32 605 19,128	132 3,851					
/3	671 24,430	997 4,18	9 1,453 14,604	198 1,98	30 1,549 20,677	91 3,942					
/4	267 24,697	13 4,203	2 417 15,021	23 2,00	03 400 21,077	96 4,038					
/5	141 24,838	2 4,204	4 584 15,605	126 2,12	29 212 21,289	117 4,155					
/6	96 24,934	13 4,21	7 1,005 16,610	38 2,16	57 159 21,448	8 4,163					
/7	374 25,308	27 4,24	£ 570 17,180	38 2,20	05 85 21,533	247 4,410					

-Continued-

	Weired System Counts											
	A	Afoqnak		Pink		Barling		Akalura		<u>E. Paramanof</u>		tery
Date	Daily	Cum.	Daily	Cum.	Daily	Cum.	Daily	Cum.	Daily	Cum.	Daily	Cum.
9/8	582	25,890	10	4,254	434 1	7,614	292	2,497	188	21,721	29	4,439
9/9	366	26,256	1	4,255	819 1	18,433	86	2,583	0	21,721	25	4,464
9/10	76	26,332	1	4,256	218 3	18,651	17	2,600	3	21,724	8	4,472
9/11	395	26,727	31	4,287	164 1	.8,815	35	2,635	0	21,724	15	4,487
9/12	531	27,258	2	4,289	0 1	18,815	46	2,681	25	21,749	10	4,497
)/13	55	27,313	0	4,289	36 1	L8,851	2	2,683	0	21,749	34	4,531
9/14	377	27,690	8	4,297			0	2,683			8	4,539
9/15	6	27,696	0	4,297			0	2,683			5	4,544
)/16	112	27,808	0	4,297			1	2,684			10	4,554
)/17							0	2,684			2	4,556
9/18							0	2,684				
)/19							1	2,685				
9/20							0	2,685				
9/21								·				
Total		27,808		4,297	1	L8,851	, <u>, , , , , , , , , , , , , , , , , , </u>	2,685		21,749		4,556

<sup>a</sup> Counts represent estimates while weir was inoperable due to high water conditions.

District	Brood Year	Estimated Escapement	Commercial Catch	Run
Afognak				
_	1969	186,747	32,819	219,566
	1971	165,335	3,389	168,724
	1973	91,279	57,911	149,190
	1975	90,474	119,150	209,624
	1977	340,793	161,474	502,267
	1979	580,287	782,546	1,362,833
	1981	984,370	1,440,345	2,424,715
	1983	422,129	192,784	614,913
	1985	975,721	1,132,382	2,108,103
	1987	597,819	505,497	1,103,316
	1989	1,991,845		1,991,845
	1991	472,029	1,112,192	1,584,221
Average	(1969-1987)	443,495	442,830	886,325
Northwes	t Kodiak			
	1969	685.383	653.549	1,338,932
	1971	535,828	133,210	669,038
	1973	397.817	218,327	616,144
	1975	690,221	1,155,439	1,845,660
	1977	1,023,479	1,670,335	2,693,814
	1979	1,565,862	3,040,943	4,606,805
	1981	1,716,753	2,600,000	4,316,753
	1983	1,371,842	1,364,787	2,736,629
	1985	1,452,625	1,801,177	3,253,802
	1987	1,334,032	1,329,254	2,663,286
	1989	5,653,387		5,653,387
	1991	792,347	3,787,647	4,579,994
Average	(1969-1987)	1,077,384	1,396,702	2,474,086
Southwes	t Kodiak			
	1969	2,100	6,669	8,769
	1971	9,347	294	9,641
	1973	896	4,261	5,157
	1975	2,778	11,923	14,701
	1977	80,571	8,188	88,759
	1979	206,975	15,234	222,209
	1981	86,081	8,090	94,171
	1983	81,841	753	82,594
	1985	97,832	88,270	186,102
	1987	40,079	241,954	282,033
	1989	168,972		168,972
	1991	188,005	1,375,838	1,563,843
Average	(1969-1987)	60,850	38,564	99,414

Appendix A.6. Kodiak Management Area odd-year pink salmon escapement, catch, and run numbers by fishing district, 1969-1991.

District	Brood Year	Estimated Escapement	Commercia Catch	ll Run
Alitak B	ay			
	1969	816,069	3,775,182	4,591,251
	1971	780,024	100,896	880,920
	1973	309,780	49,932	359,712
	1975	550,886	235,711	786,592
	1977	961,780	961,673	1,923,453
	1979	1,330,071	1,664,410	2,994,481
	1981	1,460,981	2,073,629	3,534,610
	1983	1,029,028	1,428,526	2,457,554
	1985	1,864,240	1,057,940	2,922,180
	1987	877,906	916,883	1,794,789
	1989	4,483,424		4,483,424
	1991	662,294	2,373,521	3,035,815
Average	(1969-1987)	998,076	1,226,478	2,224,555
Eastside	Kodiak			
	1969	707,874	6,704,242	7,412,116
	1971	482,507	592,876	1,075,383
	1973	299,784	91,799	391,583
	1975	550,094	382,177	932,271
	1977	885,092	2,215,285	3,100,377
	1979	1,373,155	3,685,457	5,058,612
	1981	1,310,255	2,456,641	3,766,896
	1983	959,501	783,039	1,742,540
	1985	916,113	81,673	997,786
	1987	746,297	817,847	1,564,144
	1989	3,088,875		3,088,875
	1991	1,329,082	5,650,427	6,979,509
Average	(1969-1987)	823,067	1,781,104	2,604,171
Northeas	t Kodiak	401 086	1 363 800	1 607 006
	1909	421,000 194 142	163 216	357 359
	1971	100 543	16 211	116 754
	1975	176 900	218 793	355 693
	1977	130,900	135 921	513 801
	1979	510 039	458,116	968,155
	1981	458 266	416,920	875,186
	1983	330,144	193.880	524,024
	1985	840.387	203.409	1,043,796
	1987	600,770	276,657	877,427
	1989	1.520.402	/	1,520,402
	1991	524,455	296,438	820,893
Average	(1969-1987)	397,016	334,592	731,608

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Distric	Brood t Year	Estimated Escapement	Commercia Catch	l Run	
Mainland	1				
	1969	709,324	65,562	774,886	
	1971	254,828	153,319	408,147	
	1973	168,016	24,356	192,372	
	1975	437,996	270,804	708,800	
	1977	1,267,433	343,295	1,610,728	
	1979	1,284,305	623,117	1,907,422	
	1981	1,222,730	271,758	1,492,488	
	1983	569,060	183,735	752,795	
	1985	1,022,058	261,059	1,283,117	
	1987	1,239,439	228,238	1,467,677	
	1989	7,614,276		7,614,276	
	1991	2,231,218	1,166,188	3,397,406	
Average	(1969~1987)	817,519	242,524	1,060,006	

## Appendix B.1. Pink salmon peak counts and estimated total escapement for Chignik Management Area streams, 1989<sup>a</sup>.

	Doola	Numbers of Fish			
	Peak	Estimated lotal			
Stream No.	(PC)	ESCapement (ETE)	ETE/PC <sup>b</sup>		
Eastern District	:				
272-701	3,200	4,636	1.44		
272-702	53,000	53,000	1.00		
272-703	17,000	26,983	1.59		
272-720	0	0			
272-721	10,900	14,000	1.28		
272-722	0	0			
272-723	0	0			
272-801	10,900	14,080	1.29		
272-802	9,000	16,018	1.78		
272-802A	200	226	1.13		
272-802B	750	848	1,13		
272-803	36,000	36,000	1.00		
272-804	10,600	10,600	1.00		
272-805	19,000	19,000	1.00		
272-821	0	0			
272-822	0	0			
272-823	0	0			
272-84	0	0			
272-843	0	0			
272-844	0	0			
272-845	0	0			
272-900	2,300	2,300	1.00		
272-901	10,000	10,000	1.00		
272-902	22,000	49,270	2.24		
272-903	11,400	15,480	1.36		
272-903A	18,400	21,200	1.15		
272-903B	52,000	52,000	1.00		
272-904	32,000	32,000	1.00		
272-905	89,000	89,000	1.00		
272-906	18,000	20,340	1.13		
272-921	3,200	3,616	1.13		
272-922	8,000	8,000	1.00		
272-923	9,000	9,000	1.00		
272-941	0	0			
272-961	53,000	82,592	1.56		
272-961A	8,500	8,500	1.00		
272-961B	135,000	202,667	1.50		
272-961C	25,000	58,333	2.33		
272-962	4	5	1.25		
272-962A	350	817	2.33		
272-962B	0	0			
272-963	10,500	10,500	1.00		
		••••			
Subtotal	678,204	871,011	1.28		
		-	-		

		Numbers of Fish	
	Peak	Retimated Total	
	Count	Escapement	
Stream No.	(PC)	(ETE)	ETE/PC <sup>b</sup>
			·
Central District			
272-201	210	210	1.00
272-202	0	0	
272-202A	600	903	1.50
272-2028	200	200	1.00
272-204	16,600	24,460	1.47
272-205	5	5	1.00
272~206	230	230	1.00
272-302	45,500	45,500	1.00
272-501	63,000	89,499	1.42
272-502	0	0	
272-502A	0	0	
272-503	0	0	
272-504	0	0	
272-505	0	0	1 00
272-506	300	300	1.00
272-507	0	0	
272-508	0	0	1
272-509	7,300	7,300	1.00
272-510	1,350	1,350	1.00
2/2-511	0	0	
272-511A	500	500	1.00
272-5118	0	0	
272~512			
272-514	23,000	25,497	1.14
272-310	1 700	19,093	1.00
272-602	1,700	2,275	1 1 0
272-004	200		1 1 2
272-605	3,000	1 000	1.13
272-000	,000	1,800	±.00
Subtotal	186,495	225,004	1.20
Chignik Bay Distr	ict		
271-100	0	0	
271-101A	3,000	3,000	1.00
271-101B	4,000	4,000	1.00
271-102A	. 0	0	
271-102B	0	0	
271-102C	0	0	
271-103	250	250	1.00
271-104	4,800	4,800	1.00
271-105	400	400	1.00
271-106	1,020	1,020	1.00
Subtotal	13,470	13,470	1.00

	N			
	Peak	Estimated Total		
	Count	Escapement		
Stream No.	(PC)	(ETE)	ETE/PC <sup>D</sup>	
Western Distr	ict			
273-702	2,900	2,900	1.00	
273-720	0	0		
273-722	32,000	32,000	1.00	
273-723	1,500	1,500	1.00	
273-802	10,800	10,800	1.00	
273-821	0	0		
273-822	2	2	1.00	
273-823	1,700	1,700	1.00	
273-842	1,200	1,200	1.00	
273-843	1,700	1.700	1.00	
273-844	_,,,,,	-,		
273-845	110	110	1.00	
273-941	4.500	5,982	1.33	
		- ,		
Subtotal	56,412	57,894	1.03	
Perryville Di	strict			
275-401	4.200	4,760	1.13	
275-402	9.400	9,400	1.00	
275-403	0	0		
275-404	3.800	5.813	1.53	
275-405	3,000	0		
275-406	161 000	168 403	1 04	
275-408	200	207	1 03	
275-502	51 000	51 000	1 00	
275-503	40	45	1 12	
275-504	3 700	3 700	1 00	
275-505	19 000	19 000	1 00	
275-506	2 380	2 618	1 10	
275-600	2,300	103	1 03	
275-601	1 600	2 370	1 48	
2,2 001		2,3,4 	±,40	
Subtotal	256,420	267,419	1.04	
Grand Total	1,191,001	1,434,798	1.20	

- <sup>a</sup> Table adapted from Barrett (1990).
   <sup>b</sup> Estimated total escapements derived by the area-under-the-curve (AUC) method included fractions of whole numbers which the ETE/PC ratio reflects in several instances.
- <sup>c</sup> Factor used to expand peak counts to Estimated total escapement for the years 1963-1983.

Stream					Pea	ik Count	· · ·						
Name	NO.	1969	1971	1973	1975	1977	1979	1981	1983	1985	1987	Mear	n 1989
Bastern Distric	t												•
Aniakchak	272-605	100	0	1,200	1,450	1,500	0	2,700	1,000	0	2,500	1,045	5,000
Cape Agutka	272-606	2,000	2,000	1,000	1,200	900	13,000	5,800	100	0	4.200	3.020	1.800
Main Creek	272-702	2,000	1,000	1,000	3,700	2,300	13,500	5,800	9,000	13,600	11,100	6,300	53,400
Northeast	272-703	4,500	2,000	1,100	350	5,000	7,000	3,300	2,600	9,000	5,500	4,035	19,000
Yantarni	272-721	1,000	0	300	100	1,700	14,000	13,500	3,600	12,000	13,000	5,920	11.000
Ocean Beach	272-801	4,000	0	450	400	1.000	1,000	10,500	2,600	4,500	9.350	3,380	-11,400
Nakalilok	272-804	6,000	1,000	500	2,500	2,100	12,000	6,000	4,800	15,000	1,400	5,130	10,650
Chiginagak	272-902	1,000	2,500	300	400	2,100	11.000	5,000	400	2,500	3,200	2,840	22,550
Chiginagak R.	272-903	0	0	150	2,500	1,400	400	6.000	1.300	1,000	32,900	4 565	18,800
Chiginagak-904	272-904	5,500	1,000	500	650	3,100	7.000	6.950	1,700	5,000	11,000	4 240	32 100
Chiginagak-905	272-905	2,400	4,000	800	1.300	20,800	61,000	25,000	3,100	15,000	20,700	15 410	99,000
Agripina	272-961	2,500	6,000	4.200	2.050	3,200	23,500	13,400	5,000	30,000	1 000	9 085	135 000
Glacier	272-962	0	0	0	150	-,	,~_0	50	1,300	30,000	5 500	700	350
Kilokak	272-963	200		50	250	200	1,100	300	300		9,000	1,140	11,000
Central Distric	t												
Thompson Valley	272-204	19.000	6.500	2,300	8.700	24.500	35 000	6 500	1 520	0		11 550	16 600
Hook Bay	272-302	17,000	10,000	3,700	1 900	15 300	42 700	12,000	200	2 000	8 050	11,338	47 100
Cape Kumliun	272-501	,	51,000	23,300	10,000	113,000	153 000	35 000	200	118 100	0,050	LI,203	67,100
Bear	272-505	0	,0	,	0	500	100,000	100	2 000	110,100	17	261	03,245
Rudv's	272-509	200	õ	ő	ñ	5 800	12 000	700	2,000	0	12	201	7 300
North Fork	272-514	4,000	ō	1,900	350	4,400	12,700	14,000	3,500	5,000	6,650	5,250	23,180
Western Distric	t												
Coal Cape	273-702	40 000	8 000	1 065	13 550	78 500	50 000	94 900	11 200	0	11 700	20.001	2 000
Ivan	273-722	255,000	130,000	35,000	73 100	236,000	85 000	80,000	12 200	20 000	12 900	23,301	2,300
Foot	273-802	14 000	30,000	7 000	8 700	13 000	9,600	10,000	1 200	20,000	12,800	33,910	38,700
Spoon	273-823	6 500	7 000	7,000	3 500	4 100	7 000	6 700	1,200	5,000	5,340	10,384	12,920
Portage	273-842	21,000	10,000	14 000	400	3,100	17 500	6,700	200	200	30	3,613	1,800
Seal Bay	273-843	7,500	5,000	1,500	5,000	3,000	200	9,000	1,000	5,000	500	3,770	1,200
Perryville Dist	rict												
Kupreanof	275-401	6.500	3.500	200	650	3 000	28 000	14 000	3 500			7 410	4 300
Smokey Hollow	275-402	-, <b>u</b>	-,	200	50	1 500	20,000	14,000	2,000	160	1 700	1,313	4,200
Wasco's	275-404	4.000	3,000	200	800	1,400	2.000	0	2 000	250	27 800	440	3,400
		-,•	_,			2,100	2,000	0	2,000	2.00	27,000	4,120	3,000

Appendix B.2. Pink salmon aerial survey odd-year peak counts, Chignik Management Area, 1969-1987.

Stream			Peak Count										
Name	No.	1969	1971	1973	1975	1977	1979	1981	1983	1985	1987	Mear	n 1989
Ivanof Humpback	275-406 275-502	105,000 59,000	20,000 8,000	18,100 5,000	20,800 8,100	51,800 48,200	89,000 59,000	18,000 39,000	32,000 11,000	155,000 25,000	123,400 19,800	63,310 28,510	166,000 51,000
Total Peak Con	unt <sup>a</sup> (TPC)	589,900	311,500	125,515	172,600	652,800	767,800	440,700	119,120	325,200	466,232	405,861	882,095
Total Escapem Peak	ent <sup>b</sup> (TE)	624,450	324,800	158,970	238,100	749,800	858,800	598,200	158,900				
TPC/TE€		94%	968	79€	728	87%	89€	748	75%⊦				

a Total peak counts for 31 Chigink Management Area index streams.
 b Total escapement for all aerial surveyed streams obtained from Thompson and Owen (1991).
 c Average of TPC/TE for 1969-1983 is 83.25%.