# Exxon Valdez Oil Spill State/Federal Natural Resource Damage Assessment Final Report <br> Effects of Pink Salmon (Oncorhynchus gorbuscha) Escapement Level on Egg Retention, Preemergent Fry, and Adult Returns to the Kodiak and Chignik Management Areas Caused by the Exxon Valdez Oil Spill 

Fish/Shellfish Study Numbers 7B and 8B<br>Final Report

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# Effects of Pink Salmon (Oncorhynchus gorbuscha) Escapement Level on <br> Egg Retention, Preemergent Fry, and Adult Returns to the Kodiak and Chignik Management Areas Caused by the Exxon Valdez Oil Spill 

Fish/Shellfish Study Numbers 7B and 8B Final Report

Study History: Fish/Shellfish Study Numbers 7B and 8B were initiated under a 1989 detailed study plan as Fish/Shellfish Studies 7 and 8 (Injury to Pink Salmon Spawning Areas and Pink Salmon eggs and Fry in Areas Outside of Prince William Sound). In October 1990 the study was stratified owing to disparate objectives and concerns into streams within Lower Cook Inlet (F/S Study Numbers 7A and 8A), and streams within the Kodiak-Chignik areas (F/S Study Numbers 7B and 8B). Injury assessment was continued with field studies terminating after the 1991 field season. a draft report was submitted by C.O. Swanton entitled Effects of Pink Salmon (O. gorbuscha) Escapement Level on Egg Retention, Preemergent Fry, and Adult Returns to the Kodiak and Chignik Management Areas Caused by the Exxon Valdez Oil Spill. No restoration efforts were required for pink salmon populations to either of the affected areas.


#### Abstract

Potential impacts of overescapement on several life history stages of pink salmon from streams located within the Kodiak and Chignik commercial salmon fishing areas were studied. The 1989 pink salmon escapement for Kodiak was 21.0 million (odd-year escapement goal 4.7 million) and for Chignik 1.4 million fish (odd-year escapement goal 0.7 million). Measurements of egg retention, fecundity, stream residence time (stream life), total available spawning habitat, and preemergent fry densities were obtained. Egg retention was found to be positively related to spawner density; observed 1990 preemergent fry densities were significantly below predicted values for 23 Kodiak and 7 Chignik streams indicating reduced spawner success for some streams. Return per spawner analyses for Kodiak showed a significant density dependent response, however no such result was found for the Chignik data. Overall no conclusive evidence of reduced production of pink salmon adults from the 1989 escapement event was found.


Key Words: Adult returns, Chignik, egg retention, Exxon Valdez, Kodiak, Oncorhynchus gorbuscha, overescapement, pink salmon, preemergent fry.

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## EXECUTIVE SUMMARY

As a result of the 1989 Exxon Valdez oil spill, commercial salmon fishing in and around the Kodiak and Chignik areas was severely restricted throughout the 1989 season. Consequently, pink salmon escapements for these areas greatly exceeded targeted escapement objectives. Investigations were conducted within the Kodiak and Chignik Management Areas during 1989 and 1990 to determine if negative impacts on future odd-year brood line pink salmon production occurred as a result of overescapement in 1989.

The 1989 pink salmon escapements for the Kodiak and Chignik management areas were estimated to be 21.0 and 1.4 million fish, respectively; odd-year escapement objectives are 4.7 million (Kodiak) and 0.7 million (Chignik). Egg retention of spawners was found to be positively related to spawner density. Observed 1990 preemergent fry densities were significantly below those predicted from linear regression models in 23 Kodiak and 7 Chignik index streams; depressed fry production was found in 18 Kodiak and Chignik streams, collectively. Conversely, there were also Kodiak and Chignik preemergent fry index streams that had observed 1990 fry density values which were greater than average, suggesting no density dependent effect for these streams pink salmon populations. Return per spawner analyses for Kodiak resulted in a significant density dependent parameter estimate, while for Chignik no such result was found. A possible mechanism for the density dependence found within the return per spawner analysis is the effect of spawner density on egg retention with a subsequent reduction in preemergent fry yield for some Kodiak area streams.

For Kodiak, the 1991 pink salmon return was 22.0 million; predicted returns using midpoint and desired odd-year escapement goals of 4.7 and 7.0 million were 11.9 (range 3.6 to 24.4 million) and 15.7 (range 5.8 to 29.5 million) million, respectively. Had Kodiak area pink salmon escapement midpoint or desired goals been met the 1991 return could easily have been higher or lower than the actual return of 22.0 million. The Chignik area experienced a 1991 return of 1.9 million; the predicted return using the escapement goal ( 0.7 million) was 1.3 million.

Even though spawner density had an affect on egg retention and preemergent fry, we do not believe that a measurable impact occurred on either Kodiak or Chignik area pink salmon returns from the 1989 escapement event.

## INTRODUCTION

During 1989, numerous salmon harvest opportunities within the Kodiak (KMA) and Chignik Management Areas (CMA) were foregone owing to the Exxon Valdez Oil Spill. Pink salmon (Oncorhynchus gorbuscha) runs occur in 454 Kodiak and 107 Chignik streams and comprise $78 \%$ and $31 \%$ (1978-1988), respectively, of the Kodiak and Chignik annual salmon harvests (Malloy and Prokopowich 1992; Thompson and Owen 1992). The ex-vessel value of this harvest has averaged (1978-1988), in millions of dollars, 14.2 and 1.5 for the Kodiak and Chignik areas. The KMA odd-year midpoint pink salmon escapement goal is 4.67 million, whereas the estimated total escapement for 1989 was about 21.0 million (Barrett et al. 1990). No annual escapement within the last three decades (1963-1991) has approached this level. The CMA desired pink salmon escapement goal is 0.70 million (Probasco et al. 1987); estimated total escapement during 1989 was 1.4 million (Barrett 1990).

In addition to monetary losses, the unrealized harvests of pink salmon could potentially have resulted in overescapement of spawners with consequent depression in returns from the 1989 brood year. Redd superimposition resulting from high spawner densities can be an important cause of mortality in Pacific salmon (Gilbert and Rich 1927; Smirnov 1947; Morgan and Henry 1959). McNeil (1964) demonstrated that pink salmon egg mortality during spawning was directly related to the density of females on spawning beds in two Southeastern Alaska streams. McNeil (1964) and Beverton and Holt (1957) both surmised that the production of fry within spawning beds is limited by maximum fry yields owing to density dependent mortality of eggs and alevins. Heard (1978) presented compelling evidence that this occurred in a Southeastern Alaska stream that experienced an extremely high pink salmon escapement in 1967.

The KMA encompasses the entire Kodiak archipelago and that portion of the Alaska Peninsula draining into Shelikof Strait from Cape Douglas to Kilokak Rocks bordering Imuya Bay (Figure 1). The archipelago and Alaska Peninsula portions of the management area are each about 241 km in length while Shelikof Strait which separates the two, averages approximately 48 km in width. The commercial salmon fishery occurs within seven districts which enclose about 454 pink salmon spawning streams. Fishery managers employ aerial and foot survey escapement counts into 51 index streams as part of the inseason fishery management program; 43 index streams have preemergent fry data collected for generating preseason run forecasts (Figure 2).

The CMA includes all coastal waters and inland drainages of the northwest Gulf of Alaska extending from Kilokak Rocks bordering Imuya Bay to Kupreanof Point (Figure 3). There are five commercial salmon fishing districts which contain 107 pink salmon spawning streams. Within the CMA are 31 aerial survey index streams, 18 of which have had preemergent fry sampling conducted (Figure 4).

We employed approaches spanning several life history stages for examining whether 1989 foregone harvests led to depression in spawning success, fry yields, and the subsequent returns of pink salmon. Egg retention, fecundity, stream residence time (stream life) and estimated total escapement, total available spawning habitat, and preemergent fry data were collected during 1989 and 1990. Goals of this study were:


Figure l. Map of the Kodiak Management Acea and associated commercial salmon fishing districts.


Figure 2. Approximate locations of Kodiak Management Area aerial escapement survey, preemergent fry, egg retention, stream life, and fecundity sampling streams.


Figure 3. Map of the Chignik Management Area showing commercial salmon fishing districts.


Figure 4. Chignik Management Acea aerial escapement survey, preemergent fry, egg retention, and fecundity sampling index streams.
(1) estimate the numbers of spawning pink salmon by "index" stream for systems outside Prince William Sound where historical fry density data exist. These include 43+ streams in the Kodiak Island/Shelikof Strait mainland area and 18 streams in the Chignik area;
(2) produce a catalog of aerial photographs and detailed maps of pink salmon spawner distribution for index streams in the Kodiak and Chignik areas;
(3) determine abundance of pink salmon eggs and preemergent fry. Inclusive of deriving a fecundity-length relationship for selected odd-year Kodiak and Chignik populations; estimating egg retention for populations utilizing selected preemergent fry index streams; and determine total available spawning habitat for these streams;
(4) estimate or derive indicies of overwinter mortality (egg to preemergent fry) of pink salmon eggs;
(5) determine reductions, if any, in pink salmon returns from the 1989 escapement event;
(6) identify potential alternative methods and strategies for restoration of lost use, populations, or habitat where injury is identified.

## OBJECTIVES

(1) estimate the numbers of spawning pink salmon by "index" stream for systems outside Prince William Sound where historical fry density data exist. These include $43+$ streams in the Kodiak Island/Shelikof Strait mainland area and 18 streams in the Chignik area;
(2) produce a catalog of aerial photographs and detailed maps of pink salmon spawner distribution for index streams in the Kodiak and Chignik areas;
(3) determine abundance of pink salmon eggs and preemergent fry. Inclusive of deriving a fecundity-length relationship for selected odd-year Kodiak and Chignik populations; estimating egg retention for populations utilizing selected preemergent fry index streams; and determine total available spawning habitat for these streams;
(4) estimate or derive indices of overwinter mortality (egg to preemergent fry) of pink salmon eggs;
(5) determine reductions, if any, in pink salmon returns from the 1989 escapement event;
(6) identify potential alternative methods and strategies for restoration of lost use, populations, or habitat where injury is identified.

## METHODS

## Stream Life

Stream life for this study is defined as the time span from when an individual pink salmon enters freshwater (counted through a weir or is tagged) until the individual dies (Bocking et al. 1988). It is represented as an average over all fish within a population. During 1989, fish counting weirs were installed on the Akalura, E. Paramanof, and Litnik systems to obtain daily and total escapement counts by species (Figure 2). On a minimum of two and maximum of three day rotational basis, stream foot surveys were conducted and live and dead pink salmon enumerated. Pillar Creek located about 12 km distant from the city of Kodiak, was foot surveyed only. Each survey performed had fish visibility conditions rated as affected by turbidity, water level, and cloud cover.

During 1990, weir stations were operated on the E. Paramanof, Litnik, Pink, Barling, Saltery, and Akalura streams. Daily, upstream migrating salmonids were identified and enumerated by species. Extending through about four weeks of each population's migration, color coded (unique color for each week) 30.5 cm long floy tags were affixed about 2.5 cm below the base of the dorsal fin to about 150 upstream migrants/population/week. Fish used for tagging were captured in a $2.7 \mathrm{~m} \times 5.0 \mathrm{~m}$ trap constructed of aluminum weir panels attached to the upstream side of the weir. After tagging, fish were released upstream of the weir; tag color and number of tagged
fish released by date and location were recorded (Table 1). Foot surveys were conducted on a two to three day rotational basis beginning usually a day after tagging, with live and dead tagged (by color code) and untagged fish counted and recorded. Pillar Creek was foot surveyed only. For 1990, foot survey methods were modified so that after being counted, carcasses were removed from the stream bed and gravel bars to prevent double counting on subsequent surveys.

## Total Escapement and Commercial Catch Estimation

Pink salmon escapements into Kodiak and Chignik management area streams have been assessed via aerial surveys using fixed wing aircraft and observers for over 30 years. Foot surveys and weir counts are also used for total escapement estimation. Modal "peak" pink salmon counts for a given stream-year are assumed to represent some unknown fraction of the total escapement for each species. Aerial and foot survey conditions were subjectively rated by the observer from poor to excellent depending upon fish visibility within a stream (David Prokopowich, Alaska Department of Fish and Game, Kodiak, personal communication). For this study the historical odd numbered year (odd-year brood line) survey databases for the years 1963-1991 from Kodiak and Chignik were employed to estimate total area wide pink salmon escapements (refer to Data Analysis). The 1989 pink salmon estimated total escapements and catch figures for the KMA were obtained from Barrett et al. (1990) and for the CMA from Barrett (1990).

Total KMA commercial catch numbers for 1963-1985 were extracted from Manthey et al. (1986) and for the years 1987-1991 from the ADF\&G Division of Commercial Fisheries fish ticket summary reports. For annual KMA total commercial catch, Kitoi Bay hatchery produced pink salmon were not included. CMA catch numbers 1963-1989 were obtained from Thompson and Owen (1992); catch figures for 1991 were provided by David Owen (personal communication, Alaska Department of Fish and Game, Kodiak).

## Total Available Spawning Habitat

Numerous aquatic habitat inventory sampling designs exist; however most are directed at estimating size of resident fish populations and impacts of land use practices (Platts et al. 1983; Frissel 1986; Murphy et al. 1987; Hankin and Reeves 1988). The use of visual classification of habitat units (Hankin and Reeves 1988) and substrate size (Shirazi and Seim 1979) are proven alternatives to designs relying upon direct measurements. According to habitat suitability index models developed by Raleigh and Nelson (1985), substrate size and water velocity have the largest control over spawning success of pink salmon, while substrate embeddedness is also thought to be influential (Platts et al. 1983). We employed visual classification of substrate size, flow velocity, and stream depth coupled with direct stream width measures for estimating total available pink salmon spawning habitat. The paucity of data for Kodiak and Chignik streams regarding habitat unit types (i.e. pools, riffles, and glide areas) precluded using a stratified systematic sampling design.

Table 1. Tagging dates and numbers of tagged fish released for estimating pink salmon stream life, 1990.

| Tagging Event | System |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Barling |  |  | Litnik |  |  | Akalura |  |  | Saltery |  |  | E. Paramanof |  |  |
|  | N | Date | $\begin{aligned} & \text { Tag } \\ & \text { Color } \end{aligned}$ | N | Date | $\begin{aligned} & \text { Tag } \\ & \text { Color } \end{aligned}$ | N | Date | $\begin{aligned} & \text { Tag } \\ & \text { Color } \end{aligned}$ | N | Date | $\begin{aligned} & \text { Tag } \\ & \text { color } \end{aligned}$ | N | Date | $\begin{aligned} & \text { Tag } \\ & \text { Color } \end{aligned}$ |
| 1 | 150 | 8/6-7 | Orange | 150 | 8/7 | Orange | 134 | 8/27 | Orange | 150 | 8/7-8 | Orange | 150 | 7/28 | Orange |
| 2 | 150 | 8/15 | Blue | 150 | 8/18 | Blue | 140 | 9/3-5 | Blue | 151 | 8/24-15 | Yellow | 150 | 8/9-10 | Yellow |
| 3 | 150 | 8/22-23 | Yellow |  |  |  |  |  |  | 150 | 8/22-23 | Blue | 150 | 8/17-18 | Blue |
| 4 |  |  |  |  |  |  |  |  |  | 150 | 8/29-30 | Pink | 150 | 8/23-24 | Pink |

Streams utilized by pink salmon for spawning within the KMA and CMA are generally small (less than 10 km in length) second and third order streams with low base flows typically occurring in July-August. Spawning habitat data from 45 KMA and 14 CMA index streams were collected during 1989-1990. Utilizing enlargements of United States Geological Survey (USGS) $1: 250,000$ topographic maps, individual stream maps were constructed on which fishery managers demarcated observed reaches of historical pink salmon spawner distribution for each stream. Total stream length (km) was measured from maps with a calibrated map wheel. Braided channels, mainstem tributaries, and intertidal areas were included in length measures where applicable. A two stage systematic sampling design incorporating each stream's large scale linear trends in gravel size and flow regime was employed to select a maximum of 60 onemeter strip transects from each stream. Total stream length was divided into 300 m primary units and assigned a number ( 001 to N). Five primary units were randomly selected employing a random number table, and within each primary unit, a systematic sample of 12 one-meter subunits, spaced 25 m apart were chosen for measurement.

Location of selected primary units was accomplished using maps and helicopter instrumentation. Stream width was measured every 25 m providing 12 stream width measurements per selected primary unit. Measures of stream width were to the nearest 2.5 cm from bank to bank where water depth was greater than 15 cm using a hip chain; islands were excluded. Designation of percent spawning habitat (recorded to nearest $10 \%$ ) was visually estimated as the overall area of a one meter strip transect with designations founded upon ranges of channel substrate size, stream depth, and velocity (Table 2). Gravel embeddedness was evaluated based upon whether extensive force was needed to loosen substrate materials. Suitable substrate size ranges of 0.6 to 13.7 cm ; stream depth greater than 15 cm ; and a water velocity range of 0.3 to $0.9 \mathrm{~m} / \mathrm{s}$ were used in this study component. Habitat was deemed unsuitable for spawning if visual determinations were outside of these ranges (Andrew and Geen 1960; Chambers 1956; Dvinin 1952; Krueger 1981; Neave 1966; Raleigh and Nelson 1985; Wilson et al. 1981).

## Fecundity and Egg Retention

Fecundity data (number of eggs per individual) were collected from three KMA streams (Akalura, E. Paramanof, and Litnik) and one CMA (Lake Bay) system in 1989. Sampling was conducted at fish counting weirs for Kodiak streams and with a beach seine measuring 50 mx 2.5 m with 110 mm stretch mesh for the Lake Bay system. Each fish was measured for length (mid-eye to fork-of-tail) to the nearest five mm, and egg skeins extracted. Afterward, each skein was immersed in boiling water and lightly teased to separate individual ovaries, and direct counts of eggs recorded. A total of 301 individuals, 200 from Kodiak, and 101 from Chignik were sampled.

Egg retention data from 34 spawning populations ( 23 KMA and 11 CMA ) were collected during 1989 with a maximum of 150 postspawning females sampled from each population (run). Sampling was structured systematically by selecting every third carcass encountered where concentrations were visually estimated to be less than 1,000 and every fifth carcass for concentrations greater than 5,000 . This approach was conceived because of the clumped distribution of carcasses within each stream. All samples were collected after the peak spawning

Table 2. Parameter values employed for estimating pink salmon total available spawning habitat, Kodiak and Chignik Management Area streams.

| Spawning Habitat Parameter | Reported Range |  |  | Source |
| :---: | :---: | :---: | :---: | :---: |
|  | Minimu |  | Maximum |  |
| Water Velocity | 0.4 | to | $0.8 \mathrm{~m} / \mathrm{s}$ | Dvinin (1952) |
|  | 0.45 | to | $0.73 \mathrm{~m} / \mathrm{s}$ | Hourston and Mackinnon (1957) |
|  | 0.19 | to | $0.66 \mathrm{~m} / \mathrm{s}$ | Wilson et al. (1981) |
|  | 0.10 | to | $1.32 \mathrm{~m} / \mathrm{s}$ | Graybill (1979) |
|  | 0.33 | to | $0.85 \mathrm{~m} / \mathrm{s}$ | Andrew and Geen (1960) |
| Average | 0.29 | to | $0.87 \mathrm{~m} / \mathrm{s}$ |  |
| Optimum | 0.30 | to | $0.90 \mathrm{~m} / \mathrm{s}$ | Raleigh and Nelson (1985) |
| Substrate Partical Size | 0.04 |  | 25 cm (dia.) | Wilson et al. (1981) |
|  | 0.3 |  | 10 cm (dia.) | Andrew and Geen (1960) |
|  | 0.3 |  | 10 cm (dia.) | Chambers (1956) |
|  | 2.0 | to | 10 cm (dia.) | Lucas (1960) |
| Average | 0.66 | to | 13.75 cm (di |  |
| Optimum | Unknown |  |  |  |
| Water Depth | 0.2 m |  |  | Dvinin (1952) |
|  | 0.2 | to | 7 m | Chambers (1956) |
|  | 0.28 | to | 0.78 m | Graybill (1979) |
|  | 0.37 | to | 0.63 m | Wilson et al. (1981) |
| Average | 0.26 | to | 2.8 m |  |
|  | 2.1 | m |  | Raleigh and Nelson (1985) |

period from gravel bars and banks of each stream. The body cavity of each carcass was inspected and direct counts of retained eggs and length (mid-eye to fork-of-tail) recorded to the nearest 1 mm . Carcasses exhibiting signs of predation and those deemed unspawned were noted.

## Preemergent Fry Sampling

Preemergent sac fry sampling of KMA index streams has been conducted on an annual or everyother year basis, although sampling did not actually occur in all scheduled years for most streams. The CMA preemergent fry sampling program was discontinued after 1983, however a database of at least eight odd-years were available for analysis. Sampling took place in late February through mid-April, after the period of hatching and early fry development but prior to any significant emergence and emigration having occurred. Numbers of sampling sites (spawning riffles) per stream varies from 2 to 15 , and was directly proportional to escapement and stream size. Sampling site selection was based on spawner distribution and habitat usage recorded from aerial surveys; the same riffles were sampled each year. Normally, 10 samples (digs) were taken in an " X "-shaped configuration at a randomly selected site in each riffle. Digs were made with a cylindrical frame benthic sampler that captures material forced out of the substrate by a gas powered hydraulic pump and washed downstream into a five foot long net with attached codend. An aluminum mesh covered the front half of the frame to exclude material washed from upstream, outside of the dig area. The substrate area sampled was 0.18 $\mathrm{m}^{2}$, and digs were made to a depth of approximately 15 to 47 cm for 1 to 3 minutes. Live and dead fry and eggs were counted separately, and stream temperature, predator presence, stage of fry development, number of egg fragments, and evidence of stream bed scouring and shifts were noted.

## DATA ANALYSIS

## Stream Life

Estimates of average pink salmon stream life during 1989 for the Litnik and E. Paramanof populations were derived from foot survey and cumulative weir count data, and an area-underthe spawner abundance curve (AUC) method (Johnson and Barrett 1988). Within the computer program employed for the AUC model, a stream life value was estimated iteratively until the resulting escapement estimate based on survey counts (converted to spawner days) converged on the total cumulative weir count, or "true" escapement. A spawner day represents each day that a fish is alive within the surveyed stream reach.

For foot survey data collected during 1990, several approaches were used to estimate average stream-life: (1) total number of spawner days derived from foot survey counts (live fish counts) divided by total escapement counted through the weir; (2) total spawner days obtained from foot surveys (live fish counts) divided by total carcass count; (3) iterated AUC (Johnson and Barrett 1988) method employing tagged fish data (live fish counts) where each tagged fish observed is
treated as representing a spawner day; (4) iterated AUC method with tagging data adjusted for estimated proportion of tagged fish detectable; (5) iterated AUC method employing tagged carcass counts and total number of carcasses of a color code.

Adjustment of tagging data was founded upon the assumption that not all live tagged fish within the surveyed stream reach are visible to the observer (Perrin and Irvine 1990). We further assumed that: (1) the proportion of live tagged fish from the $\mathrm{j}^{\text {th }}$ release that were detectable was independent of survey date and release date; and (2) all fish from the initial release of a tag color remained alive for at least one day after their release. Let $\mathrm{N}_{\mathrm{i}, \mathrm{j}, \mathrm{k}}$ denote the number of tagged fish from the $j^{\text {th }}$ release for the $\mathrm{k}^{\text {th }}$ stream that remain alive i days after release, noting that $\mathrm{N}_{0, \mathrm{j}, \mathrm{k}}$ is the number of tagged fish released. Also, let $\mathrm{Ct}_{\mathrm{i}, \mathrm{j}, \mathrm{k}}$ equal the survey count of live fish from the $j^{\text {th }}$ release in the $k^{\text {th }}$ stream i days after release. From assumptions (1) and (2) above, the estimated proportion of live tagged fish from the $j^{\text {th }}$ release detectable in stream $k\left(\mathrm{P}_{\mathrm{jk}}\right)$ was estimated by:

$$
\begin{equation*}
\hat{P}_{j k}=\frac{C t_{1, j, k}}{N_{0, j, k}} . \tag{1}
\end{equation*}
$$

Then, we estimated $\mathrm{N}_{\mathrm{i}, \mathrm{j}, \mathrm{k}}$ for i greater than one by:

$$
\begin{equation*}
\hat{N}_{i, j, k}=\frac{C t_{i, j, k}}{\hat{P}_{j, k}} \tag{2}
\end{equation*}
$$

## Estimated Total Escapement

Historically within the KMA and CMA "peak" aerial and foot survey counts have been used to index pink salmon escapements. Owing to a number of factors peak counts represent only a fraction of the estimated total escapement (Cousens et al. 1982). Total pink salmon escapements for Kodiak and Chignik streams during 1989 were estimated using aerial and foot survey counts, an AUC model (Johnson and Barrett 1988) and a 15 d stream life value (Barrett et al. 1990; Barrett 1990). During 1989 for both areas more streams were surveyed and with greater frequency than any year on record. Performance of the AUC model improves (increased accuracy of escapement estimates) as survey frequency increases.

For the KMA odd-year 1963-1991 stream survey databases, both the frequency and number of streams surveyed were substantially less than in 1989, therefore the methods employed during 1989 were not used. However, 1989 estimated total escapements were employed for expanding odd-year (1963-1991) peak counts.

Estimating total escapements by year ( $\mathrm{ETE}_{\mathrm{j}}$ ) incorporated the sum of peak aerial and foot survey
counts for $n$ streams in year $j\left(\mathrm{PC}_{\mathrm{j}}\right)$, index stream peak counts ( $\mathrm{PC}_{\mathrm{j}}{ }^{*}$ ), sum of all weir counts ( $\mathrm{WC}_{\mathrm{j}}$; assumed to be total escapement); and expansion factors $\mathrm{PCXF}_{89}$ (expands $\mathrm{PC}_{\mathrm{j}}$ 's) and EDXF (which allows for estimating pink salmon escapements into streams not surveyed in year j) where;

$$
\begin{align*}
& P C_{. j}=\sum_{i=1}^{n_{j}} P C_{i j},  \tag{3}\\
& P C X F_{89}=\frac{E T E_{89}}{P C_{89}},
\end{align*}
$$

and

$$
\begin{equation*}
E D X F=\frac{\sum_{j=0}^{9} P C_{1969+2 j}}{\sum_{j=0}^{9} P C_{1969+2 j}^{*}} \tag{5}
\end{equation*}
$$

The KMA escapement distribution expansion factor (EDXF) used aerial and foot stream survey coverage data from odd-years 1969-1987. The total pink salmon escapement for a given year within an area was estimated by

$$
\begin{equation*}
E T E_{j}=P C_{. j} * P C X F_{89} * E D X F+W C_{. j} . \tag{6}
\end{equation*}
$$

The KMA peak count $\left(\mathrm{PC}_{\mathrm{j}}\right)$ to estimated total escapement $\left(\mathrm{ETE}_{\mathrm{j}}\right)$ expansion factor ( $\mathrm{PCXF}_{89}$ ) was estimated to be 1.84 (Appendix A.1; Barrett et al. 1990); and the EDXF expansion factor was 1.06 (Appendix A.2).

The CMA PCXF $_{89}$ was estimated to be 1.20 (Appendix B.1; Barrett 1990); and the EDXF using data from 1969-1983 was 1.17 (Appendix B.2). Within the Chignik area for 1985-1991, estimated total escapements were derived using the AUC model (Johnson and Barrett 1988) and a 15 d stream life value (Barrett et al. 1990); adjustment for escapement distribution was also performed.

Total available spawning habitat estimates for KMA and CMA pink salmon index streams employed the following equations from Cochran (1977) incorporating variance estimator modifications for systematic sampling by Wolter (1984), provided by B. Alan Johnson (Alaska Department of Fish and Game, Commercial Fisheries memorandum 9 February, 1990). Estimates for each stream's total available spawning habitat were derived by:

$$
\begin{equation*}
\hat{Y}=\frac{N}{n} \sum_{i=1}^{n} M_{i} \bar{Y}_{i} ; \tag{7}
\end{equation*}
$$

where the sample mean for the $i^{\text {th }}$ primary unit is equal to

$$
\begin{equation*}
\bar{Y}_{i}=\sum_{i=1}^{m_{i}} \frac{y_{i j}}{m_{i}} ; \tag{8}
\end{equation*}
$$

and the overall sample mean per subunit is

$$
\begin{equation*}
\bar{Y}=\sum_{i=1}^{n} \frac{\bar{Y}_{i}}{n} . \tag{9}
\end{equation*}
$$

For these equations n represents number of primary units sampled, N depicts total number of possible primary units, $m_{i}$ number of subunits sampled within the $i^{\text {th }}$ primary unit, $\mathrm{M}_{\mathrm{i}}$ the total number of subunits within the $i^{\text {th }}$ primary unit, and $y_{i j}$ the measurement for the $j^{\text {th }}$ subunit within the $\mathrm{i}^{\text {th }}$ primary unit. The variance estimator for total available spawning habitat is:

$$
\begin{equation*}
\operatorname{var} \hat{y}=N^{2} \frac{\left(1-f_{1}\right)}{n} S_{1}^{2}+\frac{N}{n} \sum_{i=1}^{n} \frac{M_{1}^{2}\left(1-f_{2 i}\right) S_{2 i}^{2}}{m_{i}} \tag{10}
\end{equation*}
$$

which includes

$$
\begin{equation*}
s_{1}^{2}=M_{i}^{2} \sum_{i=2}^{n} \frac{\left(\bar{Y}_{i}-\bar{Y}_{i-1}\right)^{2}}{2(n-1)} \tag{11}
\end{equation*}
$$

and

$$
\begin{equation*}
s_{2 i}^{2}=\frac{1}{n m_{i}} \sum_{j=2}^{m_{i}} \frac{\left(y_{i j}-y_{i, j-1}\right)^{2}}{2\left(m_{i-1}\right)}, \tag{12}
\end{equation*}
$$

where $f$ is the sampling fraction $n / N$ for primary units, and $f_{2 i}$ equals $m_{i} / M_{i}$ which is the sampling fraction of subunits within the $\mathrm{i}^{\text {th }}$ primary unit. Several Kodiak index streams (Russian River, Seal Bay, Geographic, and Alinchak Creeks) did not have complete surveys conducted, while two systems had all available subunits measured (Bauman's and Big Waterfall Creeks).

Total available spawning habitat estimates for each index stream were employed to derive spawner densities by dividing the 1989 estimated total escapement by the estimated spawning habitat for that stream, expressed as number of fish per $\mathrm{m}^{2}$ of spawning habitat. Spawner density was used as an independent variable for analyses incorporating egg retention and preemergent fry data.

## Fecundity and Length

We modeled the dependence of fecundity ( F ) on length $(\mathrm{L})$ for the fecundity-length data collected from Litnik, E. Paramanof, Akalura, and Lake Bay as

$$
\begin{equation*}
F=\beta_{o}+\beta_{1} L+\beta_{2} Z_{1}+\beta_{3} Z_{2}+\beta_{4} Z_{3}+\beta_{5} Z_{1} L+\beta_{6} Z_{2} L+\beta_{7} Z_{3} L, \tag{13}
\end{equation*}
$$

where

$$
\begin{gathered}
Z_{1}=\left\{\begin{array}{l}
1, \text { if Litnik sample } \\
0, \text { otherwise, }
\end{array}\right. \\
Z_{2}=\left\{\begin{array}{l}
1, \text { if Paramanof sample } \\
0, \text { otherwise, }
\end{array}\right.
\end{gathered}
$$

and,

$$
Z_{3}=\left\{\begin{array}{l}
1, \text { if Akalura sample } \\
0, \text { otherwise } .
\end{array}\right.
$$

The model given by equation (13) allows the fecundity-length data from each of the four streams to be fit by a unique linear relationship. We therefore refer to equation (13) as the "full model" for fecundity on length. In contrast is the "reduced model", which allows for only one linear equation to describe the fecundity-length relationship for data from all populations as

$$
\begin{equation*}
F=\beta_{0}+\beta_{1} L \tag{14}
\end{equation*}
$$

Our interest was not with the exact values of the parameters $6_{2}$ through $6_{7}$ but whether these parameters differ significantly from zero. If $b_{2}$ through $b_{7}$ are zero, then the full model in
equation (13) reduces to equation (14); i.e., zero values for these parameters indicate that a single linear equation is sufficient to describe the fecundity-length relationship within each of these four populations. However, if any of the $b_{2}$ through $b_{7}$ parameters differ from zero, then a single linear equation would erroneously describe the fecundity-length relationship within each of the four streams. We tested the adequacy of the reduced model relative to the full model by analysis of variance (Kleinbaum and Kupper 1978). The null hypothesis for the test was,

$$
H_{0}: \beta_{2}=\beta_{3}=\ldots=\beta_{7}=0,
$$

versus the alternative hypothesis,

$$
H_{a}: \beta_{i} \neq 0 \text {, for at least one } i=2,3, \ldots 7
$$

The null hypothesis states that the fecundity-length relationship for each of the four samples falls on the same line. Deriving a single fecundity-length equation would allow for estimating pre-spawning fecundity for egg retention samples, potential egg deposition by index stream, and subsequently egg to preemergent fry survival.

## Egg Retention

Conceptually for the egg retention data, we hypothesized that there should be a positive effect of spawner density (number of fish per $\mathrm{m}^{2}$ of spawning habitat) on egg retention by individuals of a given population. To test this hypothesis we fit the data to a logistic model (Cox and Snell 1989),

$$
\begin{equation*}
P\left(E R_{i}=0\right)=\frac{\exp \beta_{0}+\beta_{1} \ln \left(D_{i}\right)}{1+\exp \left(\beta_{0}+\beta_{1} \ln \left(D_{i}\right)\right)}, \tag{15}
\end{equation*}
$$

where $\mathrm{P}\left(\mathrm{ER}_{\mathrm{i}}=0\right)$ is the probability that a female from stream i retains no eggs after spawning, and $D_{i}$ is the density of spawners in stream i. Negative values for the $b_{1}$ parameter in the above model indicate that the probability of a female retaining no eggs, decreases with increasing spawner density. Therefore, negative values for $b_{1}$ are consistent with a positive effect of spawner density on egg retention.

## Spawner Success Of 1989 Brood Year

Linear regression analysis of preemergent fry sampling data prior to 1990 was performed for each stream to test dependence of fry density on several independent variables. Live fry density
will be referred to as fry $/ \mathrm{m}^{2}$ : live fry in all digs from a stream-year divided by total $\mathrm{m}^{2}$ sampled. The independent variables were:
(1) total progeny density $\left(\mathrm{TP} / \mathrm{m}^{2}\right)$ : live and dead fry and eggs divided by $\mathrm{m}^{2}$ sampled;
(2) egg to fry survival: live fry divided by total progeny;
(3) proportion of digs with progeny: number of digs with progeny divided by total number of digs;
(4) spawning activity index: total progeny multiplied by proportion of digs with progeny; spawning density index: total progeny divided by proportion of digs with progeny.

The fry $/ \mathrm{m}^{2}$ versus $\mathrm{TP} / \mathrm{m}^{2}$ relationship was consistently the most linear fit to the data, and for some streams linearity was further improved by logarithmic or square root transformations of one or both variables. Fry $/ \mathrm{m}^{2}$ on $\mathrm{TP} / \mathrm{m}^{2}$ models, with or without transformations, were used to test for depression in 1989 spawner success (i.e., low fry $/ \mathrm{m}^{2}$ relative to $\mathrm{TP} / \mathrm{m}^{2}$ ). Observed $1990 \mathrm{TP} / \mathrm{m}^{2}$ values were entered into these models to generate corresponding predicted $\mathrm{fry} / \mathrm{m}^{2}$ values and $\alpha=0.10$ critical values of $\mathrm{fry} / \mathrm{m}^{2}$ for depression (one-tailed) to which observed 1990 $\mathrm{fry} / \mathrm{m}^{2}$ values were compared. Probabilities of $\mathrm{fry} / \mathrm{m}^{2}$ values less than or equal to those observed in 1990 were also calculated employing the $\mathrm{TP} / \mathrm{m}^{2}$ values and regression models. There were 47 streams (KMA and CMA combined) each with a database containing greater than 3 years of data for model construction. Employing data from all years, rather than only those associated with the odd-year brood line, improved model fit and allowed more streams to be analyzed. We hypothesized that survival of eggs to live fry after deposition is likely to depend on environmental factors and progeny density rather than on brood year.

The 1989 brood year spawner success was assessed by comparing $1990 \mathrm{fry} / \mathrm{m}^{2}$ for the index streams to odd-year brood line historical fry $/ \mathrm{m}^{2}$ averages, standard deviations of the averages, and $90 \%$ confidence intervals of the averages (Appendix A.3).

## Relationship Between Spawner and Live Fry Densities

The nonparametric Mann-Whitney "U" test compares rank sums for two groups of data that have been combined and rank ordered from lowest to highest (Zar 1984). This test was employed to make several comparisons of fry yield and spawner abundance using two indices of spawner abundance:
(1) spawner density: 1989 estimated total escapement for stream i ( $\mathrm{ETE}_{689}$ ) divided by the estimated total available spawning habitat for stream $\mathrm{i}\left(\mathrm{ESH}_{\mathrm{i}}\right)$,

$$
\begin{equation*}
D_{i}=\frac{E T E_{i 89}}{E S H_{i}} \text {; } \tag{16}
\end{equation*}
$$

(2) the difference or standardized residuals $\left(\mathrm{E}_{\mathrm{d} \mathrm{i}}\right)$, in standard deviations of ETE $\mathrm{E}_{\mathrm{i} 99}$ stream i
minus the odd-year average ( $\overline{\mathrm{X}}_{\text {edd }}$ ) divided by the standard deviation of average odd-year brood line escapement (\#SD ${ }_{\text {eod }}$ ):

$$
\begin{equation*}
E_{d i}=\frac{E T E_{i 89}-\bar{x}_{e o d}}{S D_{e o d}} . \tag{17}
\end{equation*}
$$

This represents the 1989 escapement relative to the historical odd-year average escapement in terms of standard deviations of the historical odd-year average escapement. Our hypothesis regarding this study component is:
$\mathrm{H}_{0}$ : fry production increases as escapement increases for stream i versus the alternative hypothesis, $H_{3}$ : fry production decreases when escapement is above some level $\epsilon^{\prime}$.

Alternatively stated, fry production should tend to increase with increasing escapement level unless overescapement with associated density dependent mortality of eggs and fry occurs, resulting in reduction of fry produced. Therefore, index streams with very high escapements or spawner densities, might show either an increase or decrease in fry yield relative to streams with less extreme escapements.

To test whether streams that showed depressed spawner success (i.e. lower fry $/ \mathrm{m}^{2}$ relative to $\mathrm{TP} / \mathrm{m}^{2}$ ) by 1989 spawners had significantly higher spawner abundance than streams that did not show such depression, the ranks of the two indices were compared between the two stream groups in a one-tailed "U" test at $\alpha=0.10$.

Additionally, two relationships were tested for comparing fry yields between index streams with moderate spawner density and those with high spawner density: (1) $\mathrm{D}_{\mathrm{i}}$ vs. fry/m², and (2) \#SD $\mathrm{csc}_{\mathrm{cs}}$ vs. $\# \mathrm{SD}_{\mathrm{fry}} / \mathrm{m}^{2}$, which are the differences in standard deviations that 1989 escapements and 1990 fry $/ \mathrm{m}^{2}$ were from the historical odd-year brood line average escapement, and $\mathrm{fry} / \mathrm{m}^{2}$, respectively. To stratify streams into two groups for the first of these two-tailed "U" tests at $\alpha=0.10$, individual streams were ranked in order of increasing spawner density with cumulative spawner density calculated stepwise along this ascending order (Table 3). Moderate spawner density streams were those with cumulative spawner density less than the spawner density cut off point of 1.4 fish $/ \mathrm{m}^{2}$, derived from the midpoint escapement goal and spawning habitat estimates (Table 3; Appendix A.4). The moderate strata included Kodiak streams with individual spawner densities ranging from 0.09 to 4.35 fish $/ \mathrm{m}^{2}$. All other streams were designated as having high spawner density (range 5.68 to 60.9 fish $/ \mathrm{m}^{2}$ ). For the second of these tests, moderate escapement streams were those for which \#SD ${ }_{\text {esc }}$ was less than 2 (range for Kodiak streams was -0.94 to 1.82 ), and high escapement streams were those for which 2 was less than \#SD csc ( 2.08 to 18.40 for Kodiak). All Mann-Whitney tests were conducted separately for Kodiak and Chignik streams because these areas experience considerably different abiotic conditions. Comparisons of fry yield between streams with moderate and high spawner abundances could not be made for Chignik, owing to a lack of streams with high spawner abundance.

Table 3. Moderate and high spawner density index streams of the preemergent fry sampling program, 1989 brood year.


Moderate Spawner Density Streams (Kodiak Management Area)

| Narrows | $257-401$ | 0.095 | 264.76 | 21,917 | 2,097 | 0.095 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| Karluk | $255-101$ | 0.132 | 0.08 | 853,844 | 111,977 | 0.131 |
| Kukak | $262-271$ | 0.178 | 6.73 | 875,101 | 115,779 | 0.132 |
| Marka | $252-343$ | 0.347 | 10.32 | 935,000 | 136,573 | 0.146 |
| E. Paramanof | $251-404$ | 0.480 | 278.27 | 976,775 | 154,991 | 0.159 |
| Zachar | $254-301$ | 0.889 | 12.06 | $1,094,540$ | 258,974 | 0.237 |
| Sheratin | $259-371$ | 1.714 | 39.85 | $1,164,507$ | 378,917 | 0.325 |
| Barling | $258-522$ | 1.900 | 51.70 | $1,242,067$ | 526,340 | 0.424 |
| Dakavak | $262-551$ | 1.902 | 0.36 | $1,276,702$ | 592,250 | 0.464 |
| Dog Salmon | $257-403$ | 2.290 | 179.86 | $1,414,457$ | 907,809 | 0.642 |
| Kashvik | $262-604$ | 2.348 | 0.81 | $1,534,814$ | $1,190,409$ | 0.776 |
| Sidold's | $259-242$ | 2.370 | 116.41 | $1,593,752$ | $1,330,077$ | 0.835 |
| Kinak | $262-451$ | 2.415 | 161.66 | $1,616,239$ | $1,384,404$ | 0.857 |
| Kiliuda | $258-207$ | 3.068 | 77.01 | $1,638,977$ | $1,454,166$ | 0.887 |
| Little R. | $253-115$ | 3.097 | 5.62 | $1,701,348$ | $1,647,333$ | 0.968 |
| Perenosa | $251-830$ | 3.274 | 72.29 | $1,730,571$ | $1,743,033$ | 1.007 |
| Miam | $259-412$ | 3.655 | 118.56 | $1,784,208$ | $1,939,120$ | 1.087 |
| American | $259-231$ | 3.910 | 255.97 | $1,877,741$ | $2,304,503$ | 1.227 |
| Saltery | $259-415$ | 3.937 | 521.49 | $1,932,226$ | $2,519,044$ | 1.304 |
| Jute | $262-801$ | 4.257 | 14.54 | $1,937,699$ | $2,542,347$ | 1.312 |
| Hurst | $259-414$ | 4.351 | 102.65 | $1,970,451$ | $2,684,882$ | 1.363 |

High Spawner Density Streams (Kodiak Management Area)

|  |  |  |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| Buskin | $259-211$ | 5.689 | 89.75 | $2,018,396$ | $2,957,667$ | 1.465 |
| Terror | $253-331$ | 6.625 | 0.75 | $2,080,416$ | $3,368,580$ | 1.619 |
| Uyak-202 | $254-202$ | 6.479 | $1,070.00$ | $2,179,998$ | $4,013,737$ | 1.841 |
| Missak | $262-402$ | 6.369 | 131.75 | $2,185,651$ | $4,049,857$ | 1.853 |
| Kanatak | $262-802$ | 6.883 | 186.68 | $2,208,656$ | $4,208,207$ | 1.905 |
| Bauman's | $253-332$ | 7.674 | 240.71 | $2,254,698$ | $4,561,540$ | 2.023 |
| Deadman | $257-502$ | 9.679 | 546.94 | $2,367,674$ | $5,655,083$ | 2.388 |
| Oil | $262-751$ | 10.300 | 225.90 | $2,378,890$ | $5,770,695$ | 2.426 |
| Humpy | $257-701$ | 10.380 | 464.79 | $2,550,629$ | $7,553,891$ | 2.962 |
| Kaiugnak | $258-542$ | 14.400 | 736.78 | $2,558,844$ | $7,672,214$ | 2.998 |
| Uganik | $253-122$ | 16.530 | 9.42 | $2,606,981$ | $8,468,113$ | 3.248 |
| Big Creek | $262-851$ | 26.460 | 131.75 | $2,685,530$ | $10,546,546$ | 3.927 |
| Portage | $262-702$ | 29.640 | 2.33 | $2,690,215$ | $10,685,426$ | 3.972 |
| Seal | $251-901$ | 38.730 | 869.79 | $2,690,370$ | $10,691,426$ | 3.974 |
| Danger | $252-332$ | 46.690 | 292.54 | $2,691,210$ | $10,730,633$ | 3.987 |
| Seven Rivers | $258-701$ | 60.880 | 383.53 | $2,711,190$ | $11,923,316$ | 4.398 |
| L Waterfall | $251-822$ | 63.530 | 70.00 | $2,712,864$ | $12,029,663$ | 4.434 |
| Alinchak a | $262-651$ | 160.900 | 27.28 | $2,716,230$ | $12,571,490$ | 4.628 |

Moderate Spawner Density Streams (Chignik Management Area)

| Portage | $272-842$ | 0.062 | 64.22 | 19,107 | 1,200 | 0.063 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| Spoon | $273-823$ | 0.071 | 0.00 | 42,759 | 2,900 | 0.068 |
| Ivan R. | $273-722$ | 0.169 | 13.68 | 231,759 | 34,900 | 0.151 |
| North Fk. | $272-702$ | 0.197 | 0.00 | 360,596 | 60,397 | 0.167 |
| Amber Ck. | $272-703$ | 0.264 | 0.00 | 560,637 | 113,397 | 0.202 |
| Hook Bay | $272-302$ | 0.293 | 0.00 | 713,881 | 158,397 | 0.222 |
| Humpback | $275-502$ | 0.462 | 65.59 | 824,175 | 209.397 | 0.254 |

Table 3. (page 2 of 2 )

| Stream |  | Stream Specific Density |  | Cummulative |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Spawning Habjtat ( $\mathrm{m}^{2}$ ) | Escapement (No. of fish) | Spawner Density (fish/m ${ }^{2}$ ) |
|  |  | Spawner |  |  | Live Fry |
| Name | No. |  |  |  | $\left(\mathrm{fish} / \mathrm{m}^{2}\right)$ | (fry/m²) |
| Foot Bay | 273-802 |  | 0.575 | 227.79 | 842,955 | 220,197 | 0.261 |
| Ivanof | 275-406 | 0.655 | 200.65 | 1,099,678 | 388,600 | 0.353 |
| Agripina | 272-961 | 0.701 | 279.66 | 1,217,443 | 471,192 | 0.387 |
| Chiginagak 905 | 272-905 | 1.639 | 180.85 | 1,271,727 | 560,192 | 0.440 |
| Chiginagak 904 | 272-904 | 1.662 | 0.18 | 1,290,972 | 592,192 | 0.459 |
| Kumliun | 272-501 | 3.693 | 2.56 | 1,315,206 | 681,691 | 0.518 |

${ }^{\text {a }}$ Based upon incomplete spawning habitat survey.

Spawner-recruit curves were fit to odd brood year (1963 to 1989) pink salmon escapement and return data from the Kodiak and Chignik management areas, separately. We used the models of Ricker (1954) and Beverton and Holt (1957) as two competing models for the spawner-recruit curve. For the Ricker model, number of returns (recruits) $R_{i}$, resulting from number of spawners, $\mathrm{S}_{\mathrm{i}}$, in year i is modelled as,

$$
\begin{equation*}
R_{i}=\beta_{1} S_{i} \exp \left(-\beta_{2} S_{i}\right) \tag{18}
\end{equation*}
$$

while for the Beverton-Holt model the relationship is modelled as,

$$
\begin{equation*}
R_{i}=\frac{1}{\beta_{1}+\beta_{2} / S_{i}} . \tag{19}
\end{equation*}
$$

For both models the parameters $B_{1}$ and $B_{2}$ are constrained to be non-negative. Non-zero values of $B_{2}$ in both models define a negative effect of spawner numbers on subsequent returns per spawner. In both models, $\mathrm{R}_{\mathrm{i}}$ will be 0 when $\mathrm{S}_{\mathrm{i}}$ is 0 , but the two models differ in the density dependent response in $\mathrm{R}_{\mathrm{i}}$ to increasing values of $\mathrm{S}_{\mathrm{i}}$. For the Ricker model, $\mathrm{R}_{\mathrm{i}}$ has a maximum value of $S_{i}=\left(B_{2}\right)^{-1}$ which is the carrying capacity of the environment for recruits, and decreases to 0 as $S_{i}$ increases to $\infty$. The Ricker $b_{1}$ represents the estimated maximum return per spawner, commonly referred to as the productivity parameter. For the Beverton-Holt model $R_{i}$ increases asymptotically to $\left(B_{1}\right)^{-1}$ as $S_{i}$ increases to $\infty$.

We computed maximum likelihood estimates (MLE's) of the parameters for both models using the transform-both sides (TBS) methodology of Carrol and Ruppert (1988). Following Carrol and Ruppert, we used the modified power transformation of Box and Cox (1964) as the transforming function. MLE's were computed using the "pseudo-model" approach of Carrol and Ruppert and the nonlinear estimation routines (NONLIN) of SYSTAT (Wilkinson 1990).

## RESULTS

Stream Life Estimates

Data collection from the Akalura system in 1989 suffered from poor survey conditions and high stream flow, therefore stream life was not estimated. Pink salmon stream life for East Paramanof Cr. (escapement of 20,561 ) and the Litnik R. (escapement of 7,477 ) populations averaged 8.5 and 8.3 days, respectively (Table 4).

Table 4. Stream life estimates derived using weir escapement, cumulative carcass, adjusted and not adjusted tagged fish counts from Kodiak Management Area pink salmon populations during 1989-1990.

| Stream | 1989 ${ }^{\text {a }}$ | $1990^{\text {a }}$ | $c c^{\text {b }}$ | Stream Life Estimate (Days) |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | - Tagging Event |  |  |  |  |  |  |  |  |  |
|  |  |  |  | 1 |  | 2 |  | 3 |  | 4 |  | Avg. |  |
| Name Number |  |  |  | $\mathrm{Nadj}^{\text {a }}$ | Adj | $\mathrm{N}_{\text {adj }}$ | Adj | Nadj | . Adj | $\mathrm{Nadj}^{\text {a }}$ | Adj | $\mathrm{Nadj}^{\text {a }}$ | Adj. |
| Barling Cr. 258-522 |  | 7.9 | 15.4 | 10.4 | 15.4 | 7.3 | 15.1 | 7.7 | 14.8 |  |  | 8.5 | 15.1 |
| Litnik R, 252-342 | 8.5 | 6.7 |  | 4.0 | 8.1 | 2.6 | 6.3 |  |  |  |  | 3.3 | 7.2 |
| Pink Cr. |  | 6.5 | 6.8 |  |  |  |  |  |  |  |  |  |  |
| Akalura Cr. 257-302 | d | 3.5 | 5.9 | 7.5 | 9.7 | 4.3 | 10.0 |  |  |  |  |  |  |
| Pillar Cr. 259-101 |  |  | 16.4 |  |  |  |  |  |  |  |  |  |  |
| Saltery Cr. 259-415 |  | 13.8 | 15.8 | 19.1 | 26.1 | 13.8 | 18.7 | 10.5 | 14.6 | 11.2 | 13.2 | 13.7 | 18.1 |
| E.Paramanof 251-405 | 8.3 | 14.4 | 14.2 | 7.9 | 11.8 | 4.0 |  | 4.2 | 11.5 | 3.8 | 6.4 | 5.0 | 9.9 |
| Average | 8.4 | 8.8 | 12.4 | 9.8 | 14.2 | 6.4 | 12.5 | 7.5 | 13.6 | 7.5 | 9.8 | 7.6 | 12.6 |

${ }^{\text {a }}$ Area under the curve method employing foot survey and weir counts.
${ }^{b}$ Cumulative spawner days from foot survey counts divided by cumulative carcass counts.
${ }^{c}$ Stream life estimated using tagged fish survey data not adjusted for visibility.
${ }^{d}$ Data excluded due to high water conditions.

The 1990 stream life estimates derived from cumulative spawner days divided by total enumerated escapement were: 7.9 d Barling Cr., 6.7 d Afgonak R., 6.5 d Pink Cr., 3.5 d Akalura Cr., 13.8 d Saltery Cr., and 14.4 d East Paramanof Cr. (Table 4); the average over all populations was 8.8 days. Stream life estimated by cumulative spawner days divided by cumulative carcass counts were: $15.4,6.8,5.9,16.4,15.8$, and 14.4 days for the Barling, Pink, Akalura, Pillar, Saltery, and East Paramanof systems; the average over all populations was 12.4 days.

Estimates using live tagged fish were completed for the East Paramanof and Saltery Creek's populations using all four tag colors with about 7-10 days separating each tagging event. Escapements into these systems were 21,749 and 4,556 fish, respectively (Appendix A.5). For the other three populations only two to three tag colors were used. Adjusting estimates for tagged fish visibility (tagged fish detectable $P_{j k}$ ) increased the unadjusted estimates by 2.6 to 10.0 days. A trend of decreasing stream life over time for the earliest to latest tagging events was observed for the East Paramanof, Saltery, and to a lesser extent Barling populations. Averages over all populations by tagging event, not adjusted were: $9.8 \mathrm{~d}, 6.4 \mathrm{~d}, 7.5 \mathrm{~d}$, and 5.5 d . Adjusted estimates by event were: $14.2 \mathrm{~d}, 12.5 \mathrm{~d}, 13.6 \mathrm{~d}$ and 9.8 d (Table 4). A three day moving average of weir escapement counts for Barling, Saltery, and East Paramanof Creeks depicts tagging events by population, relative to escapement distribution (Figure 5).

Pink salmon stream life estimates generated from tagged fish carcass recovery data averaged over all tag colors for a population were: $17.9 \mathrm{~d}, 14.6 \mathrm{~d}, 15.6 \mathrm{~d}$, and 11.4 d (Table 5). A decreasing trend in average stream life by date was observed.

## Estimated Total Escapement and Commercial Catch

Kodiak Management Area odd-year brood line ETE's (1963-1991) averaged 4.9 million (range 1.21 to 21.0; Table 6). Commercial harvests excluding 1989, have ranged from 0.18 million to 15.25 million. The total return from the 1989 escapement was 21.63 million or about 1.02 returns per spawner; odd-year escapement goals range from 2.3 to 7.0 million with a midpoint of 4.67 (Appendix A.4; Barrett et al. 1990). The midpoint goal has been reached four times during 1963-1987. Escapement and run data (1969-1989) by fishing district depict the 1991 run was greater than average (1969-1987) for all districts except the Northeast District, which was slightly above average (Appendix A.6; Figure 6).

Within the CMA estimated escapements ranged from 0.21 million to 1.13 million. The 1989 escapement of 1.45 million was 2 X both the escapement goal and 1963-1987 average escapement. Commercial harvests have averaged 0.7 million (range 25,000 to 1.87 million). The return from the 1989 escapement was 1.96 million ( 1.34 returns per spawner), whereas the odd-year average (1963-1987) is 1.41 million (Table 7). The escapement goal of 0.7 million has been exceeded five times during 1963-1987.


Figure 5. Three day moving average of daily escapement counts by system depicting tagging events in relation to escapement distribution.

Table 5. Stream life estimates derived for selected Kodiak pink salmon populations using color coded tagged fish carcass recoveries, 1990.

| Stream |  | Stream Life Estimate by Tagging Event (Days) |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Name | Number | I | 2 | 3 | 4 |  |
| Barling Cr. | 258-522 | 19.9 | 17.6 | 17.3 |  |  |
| Litnik R. | 252-342 | 10.1 | 8.4 |  |  |  |
| Akalura Cr . | 257-302 | 13.4 | 8.0 |  |  |  |
| Saltery Cr. | 259-415 | 25.1 | 18.9 | 14.8 | 13.4 |  |
| E.Paramanof Cr . | 251-405 | 21.4 | 20.2 | 14.8 | 9.5 |  |
| Average |  | 17.9 | 14.6 | 15.6 | 11.4 |  |

Table 6. Kodiak Management Area odd-year pink salmon escapements, commercial catch, average weight, return, and return per spawner, 1963-1989.

| Brood Year | Estimated Total Escapement | Commercial Catch |  |  | Return/ <br> Spawnera |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Numbers | Mean Weight (kg) | Return |  |
| 1963 | 2,554,969 | 5,480,000 | 1.5 | 8,034,969 | 1.92 |
| 1965 | 2,026,981 | 2,886,831 | 1.7 | 4,913,812 | 1.19 |
| 1967 | 2,231,335 | 187,813 | 1.9 | 2,419,148 | 6.99 |
| 1969 | 3,110,525 | 12,492,576 | 1.8 | 15,603,101 | 2.13 |
| 1971 | 2,293,556 | 4,332,994 | 1.7 | 6,626,550 | 0.75 |
| 1973 | 1,213,496 | 511,708 | 1.8 | 1,725,204 | 4.30 |
| 1975 | 2,284,854 | 2,942,801 | 1.9 | 5,227,655 | 4.63 |
| 1977 | 4,334,716 | 6,250,667 | 1.9 | 10,585,383 | 3.88 |
| 1979 | 5,709,497 | 11,121,333 | 1.7 | 16,830,830 | 2.62 |
| 1981 | 5,802,258 | 9,183,467 | 1.7 | 14,985,725 | 1.44 |
| 1983 | 3,902,959 | 4,474,760 | 1.6 | 8,377,719 | 2.45 |
| 1985 | 5,710,970 | 3,886,501 | 1.6 | 9,597,471 | 1.55 |
| 1987 | 4,869,933 | 4,013,509 | 1.6 | 8,883,442 | 4.36 |
| 1989 | 21,084,539 | 183,235 | 1.3 | 21,267,774 | 1.02 |
| 1991 | 6,381,344 | 15,252,123 | 1.3 | 21,633,467 |  |
| Average | 4,900,795 | 5,546,687 | $1.7{ }^{\text {b }}$ | $10,447,483$ | $2.94{ }^{\text {c }}$ |

${ }^{\text {a }}$ Return is from brood year +2 .
b Average weight excluding 1989 and 1991.
c Excluding 1989.


Figure 6. Kodiak Management Area odd year pink salmon estimated escapements, catch, and run numbers by fishing district, 1969-1991.

Table 7. Chignik Management Area odd-year pink salmon escapements, commercial catch, return, and return per spawner, 1963-1989.

| Brood <br> Year | Estimated <br> Total <br> Escapement | Commercial <br> Catch | Return | Return/ <br> Spawner |
| :--- | ---: | ---: | ---: | ---: |
|  |  |  |  |  |
| 1963 | 930,032 | $1,662,400$ | $2,592,432$ | 1.73 |
| 1965 | 494,437 | $1,117,100$ | $1,611,537$ | 1.12 |
| 1967 | 447,574 | 108,300 | 555,874 | 6.23 |
| 1969 | $1,010,858$ | $1,779,800$ | $2,790,658$ | 1.14 |
| 1971 | 549,067 | 612,300 | $1,161,367$ | 0.42 |
| 1973 | 209,438 | 25,500 | 234,938 | 1.81 |
| 1975 | 313,432 | 66,200 | 379,632 | 5.07 |
| 1977 | 987,032 | 604,700 | $1,591,732$ | 3.04 |
| 1979 | $1,130,518$ | $1,876,600$ | $3,007,118$ | 1.72 |
| 1981 | 787,072 | $1,162,600$ | $1,949,672$ | 0.67 |
| 1983 | 209,043 | 321,100 | 530,143 | 3.36 |
| 1985 | 528,989 | 175,000 | 703,989 | 1.20 |
| 1987 | 390,309 | 246,800 | 637,109 | 3.79 |
| 1989 | $1,453,450$ | 27,712 | $1,481,162$ | 1.34 |
| 1991 | 788,861 | $1,169,248$ | $1,958,109$ |  |
|  |  |  |  |  |

[^0]Within the KMA 46 of 51 aerial survey index streams had total available spawning habitat data collected. The Karluk and Ayakulik Rivers are the largest within the KMA and provide the greatest amount of spawning habitat $829,515 \mathrm{~m}^{2}$ and $364,644 \mathrm{~m}^{2}$, respectively (Table 8). However, both are even brood year dominant systems (Donnelly 1983). For odd-year dominant systems Humpy, Deadman, Kashvik, and Zachar Creeks along with the Dog Salmon River, all had greater than $100,000 \mathrm{~m}^{2}$ of estimated spawning habitat. Spawner densities for these systems varied from 0.89 fish $/ \mathrm{m}^{2}$ (Zachar Cr.) to 10.38 fish $/ \mathrm{m}^{2}$ (Humpy Cr.). Spawner densities for all index streams ranged from 0.01 to 60.8 fish $/ \mathrm{m}^{2}$. Coefficients of variation (CV) for most streams were less than $10 \%$ (range $0.1 \%$ to $277.0 \%$ ).

Within the CMA, 14 index streams had spawning habitat estimates generated, ranging from $17,885 \mathrm{~m}^{2}$ to $256,723 \mathrm{~m}^{2}$; coefficients of variation were all less than $10 \%$ (Table 9). Spawner densities between systems were variable with most substantially lower than estimates from the KMA. The Amber and Foot Bay systems both had spawning habitat estimated with incomplete surveys.

## Fecundity

Of the four populations sampled for fecundity and length, East Paramanof Cr. pink salmon were the smallest in length ( $X_{1}=465 \mathrm{~mm}$ ) and least fecund ( $\mathrm{X}_{\mathrm{f}}=1,312$ eggs/female). Alternatively, Akalura Cr. fish were the largest and most fecund ( $X_{1}=483 \mathrm{~mm} ; \mathrm{X}_{\mathrm{f}}=1,627 \mathrm{eggs} /$ female $)$. The two northern populations (East Paramanof and Litnik) were the smallest as compared to the Akalura and Lake Bay populations located about 162 and 405 km respectively, to the south (Table 10).
Testing for differences between each population's linear relationship of fecundity on length provided an F statistic value of 9.38 with 6 and 293 df ; we therefore rejected the null hypothesis (no significant difference of fecundity on length between populations) with P less than $10^{-8}$ (Table 11; Figure 7). We determined the fecundity-length relationship differs between pink salmon from the four populations sampled. Fecundity, according to our analysis does vary positively with body length within a given locality, however no single fecundity-length relationship was applicable to all localities. Therefore, no attempt was made to apply a single fecundity-length relationship to estimate fecundity for populations sampled for egg retention or preemergent fry.

## Egg Retention

There were 4,743 carcasses examined from 34 spawning populations, 128 fish overall were designated unspawned. Egg retention from the Kodiak and Chignik pink salmon populations varied from zero to intact egg skeins. Mean egg retention for all populations ranged from 1.5 to $146.0 \mathrm{eggs} /$ female with corresponding spawner densities of 0.2 to $60.8 \mathrm{fish} / \mathrm{m}^{2}$ (Table 12).

Table 8. Estimates of total available spawning habitat and 1989 spawner densities for Kodiak Management Area aerial survey escapement index streams.

| Stream |  |  | Spawning Habitat |  |  | $\frac{\left(m^{2}\right)}{m_{i}^{a}}$ | $\begin{aligned} & 1989 \\ & \text { Spawner Density } \\ & \text { (No. Fish/m2) } \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Name | Number | Length (km) | Est. | SD | CV |  |  |
| Afognak District |  |  |  |  |  |  |  |
| Malina Cr. | 251-105 | 4.8 | 5,524 | 330.8 | 5.9 | 60 | 0.73 |
| E. Paramanof Cr . | 251-404 | 5.6 | 41,346 | 917.3 | 2.2 | 60 | 0.48 |
| Big Waterfall Cr. | 251-821 | 0.5 | 1,488 | 411.2 | 27.6 | 7 | 3.09 |
| Portage Cr. | 251-830 | 2.4 | 29,223 | 1,120.1 | 3.8 | 60 | -b, c |
| Seal Bay Cr. | 251-902 | 1.2 | 103 | 39.2 | 38.1 | 16 | _D, c |
| Danger Cr. | 252-332 | 6.4 | 839 | 138.9 | 16.6 | 60 | 46.73 |
| Afognak R. | 252-342 | 2.4 | 27,864 | 481.9 | 1.7 | 60 | 1.49 |
| Marka Cr. | 252-343 | 6.2 | 58,724 | 1,416.2 | 2.4 | 60 | 0.35 |
| Northwest Kodiak District |  |  |  |  |  |  |  |
| Sheratin Cr. | 252-371 | 6.4 | 67,847 | 1,287.6 | 1.9 | 60 | -b, c |
| Little R. | 253-115 | 14.4 | 62,370 | 1,223.2 | 1.9 | 60 | 3.10 |
| Uganik R. | 253-122 | 7.6 | 46,903 | 3,536.9 | 7.5 | 60 | 8.76 |
| Terror R. | 253-331 | 14.8 | 61,192 | 2,032.5 | 3.3 | 60 | 6.64 |
| Baumans Cr. | 253-332 | . 6 | 44,646 | 1,733.1 | 3.9 | 48 | $7.91{ }^{\text {c }}$ |
| Browns Cr. | 254-204 | 5.1 | 40,177 | 1,642.4 | 4.1 | 60 | 1.59 |
| Uyak Cr. | 254-202 | 5.6 | 97,818 | 1,112.2 | 1.1 | 60 | 6.59 |
| Zacher Cr. | 254-301 | 12.8 | 116,852 | 2,391.7 | 2.0 | 60 | 0.89 |
| Southwest Kodiak District |  |  |  |  |  |  |  |
| Karluk R. | 255-101 | 34.4 | 829,515 | 5,437.2 | $<0.1$ | 60 | 0.13 |
| Ayakulik R. | 256-201 | 38.4 | 364,644 | 3,460.6 | $<0.1$ | 60 | 0.12 |
| Sturgeon R. | 256-401 | 22.4 | 178,278 | 1,658.2 | $<0.1$ | 60 | 0.01 |
| Alitak Bay District |  |  |  |  |  |  |  |
| Narrows Cr. | 257-401 | 5.6 | 21,532 | 803.8 | 3.7 | 60 | 0.10 |
| Dog Salmon R. | 257-403 | 6.8 | 137,754 | 2,059.0 | 1.5 | 60 | 2.29 |
| Deadman Cr . | 257-502 | 8.8 | 110,465 | 2,080.5 | 1.9 | 60 | 9.90 |
| Humpy Cr. | 257-701 | 14.4 | 171,739 | 2,041.9 | 1.1 | 60 | 10.38 |
| Eastside Kodiak District |  |  |  |  |  |  |  |
| Kiliuda Cr. | 258-207 | 3.6 | 22,737 | 1,406.0 | 6.2 | 60 | 3.06 |
| Barling R. | 258-522 | 6.4 | 75,210 | 1,470.8 | 1.9 | 60 | 1.96 |
| Kaiugnak Cr. | 258-542 | 2.0 | 7,824 | 448.5 | 5.7 | 60 | 15.12 |
| Seven R. | 258-701 | 10.0 | 19,588 | 742.6 | 3.8 | 60 | 60.88 |
| Miam Cr. | 259-412 | 4.8 | 53,636 | 1,206.7 | 2.2 | 60 | 3.65 |
| Hurst Cr. | 259-414 | 11.6 | 32,471 | 914.6 | 2.8 | 60 | 4.39 |
| Saltry Cr. | 259-415 | 6.8 | 53,695 | 922.3 | 1.7 | 60 | 4.68 |
| Northeast Kodiak District |  |  |  |  |  |  |  |
| Pillar Cr. | 259-102 | 1. 6 | 2,541 | 196.2 | 7.7 | 60 | 16.90 |
| Buskin R. | 259-211 | 7.2 | 47,945 | 1,688.9 | 3.5 | 60 | 5.69 |
| Sid Old's R. | 259-231 | 8.2 | 93,533 | 1,869.1 | 1.9 | 60 | 3.90 |
| Russian R. | 259-242 | 3.8 | 58,938 | 1,406.3 | 2.4 | 48 | -b,c |
| American R. | 259-231 | 9.1 | 93,383 | 1,625.2 | 1.7 | 60 | 3.91 |

Table 8. (page 2 of 2 )

| Stream |  |  | Spawning Habitat |  |  | $\frac{\left(m^{2}\right)}{m_{i}^{2}}$ | $\begin{aligned} & 1989 \\ & \text { Spawner Density } \\ & \text { (No. Fish } / \mathrm{m}^{2} \text { ) } \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Name | Number | Length (km) | Est. | SD | CV |  |  |
| Mainland District |  |  |  |  |  |  |  |
| Kukak Cr. | 262-271 | 2.4 | 21,257 | 654.1 | 3.1 | 60 | 0.17 |
| Missak Cr. | 262-402 | 2.8 | 5,292 | 713.7 | 13.5 | 53 | 6.82 |
| Kinak Cr. | 262-451 | 6.0 | 21,738 | 980.8 | 4.5 | 60 | 2.50 |
| Geographic Cr. | 262-501 | 1.2 | 63 | 174.7 | 277.2 | 12 | - b |
| Dakavak Cr. | 262-551 | 8.4 | 34,635 | 2,802.4 | 8.1 | 60 | 1.90 |
| Kashvik Cr. | 262-604 | 8.8 | 117,682 | 1,688.6 | 1.4 | 60 | 2.40 |
| Alinchak Cr. | 262-651 | 6.8 | 3,317 | 329.2 | 9.9 | 48 | - ${ }^{\text {d }}$ |
| Portage Cr. | 262-702 | 3.6 | 4,543 | 618.8 | 13.6 | 60 | 30.57 |
| Oil Cr. | 262-751 | 4.4 | 10,966 | 673.7 | 6.1 | 60 | 10.54 |
| Jute Cr. | 262-801 | 3.6 | 5,473 | 468.7 | 8.5 | 60 | 4.25 |
| Kanatak Cr. | 262-802 | 2.4 | 23,005 | 1,083.7 | 4.7 | 60 | 6.88 |
| Big Cr. | 262-851 | 9.2 | 77,704 | 1,424.2 | 1.8 | 60 | 26.74 |

${ }^{\text {a }}$ Number of secondary units sampled.
b Streams that were not escapement surveyed in 1989.
${ }^{\text {c }}$ Entire stream reach available for spawning was measured.
${ }^{\mathrm{d}}$ Incomplete spawning surveys, therefore estimates are approximate.

Table 9. Estimates of total available spawning habitat and 1989 spawner densities for Chignik Management Area aerial survey escapement index streams.

| Stream |  |  | Spawning Habitat ( $m^{2}$ ) |  |  |  | 1989 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Name | Number | $\begin{gathered} \text { Length } \\ (\mathrm{km}) \end{gathered}$ | Est. | SD | CV | $m_{i}{ }^{\text {a }}$ | Spawner Density <br> (No. Fish/m ${ }^{2}$ ) |
| Eastern District |  |  |  |  |  |  |  |
| Ocean Bch. Cr. | 272-801 | 9.2 | 80,013 | 2,146.6 | 2.7 | 60 | 0.17 |
| Chiginagak 904 Cr . | 272-904 | 1.2 | 19,245 | 1,304.1 | 6.8 | 60 | 1.66 |
| Chiginagak 905 Cr . | 272-905 | 7.5 | 54,283 | 971.6 | 1.8 | 60 | 1. 64 |
| Agripina Cr. | 272-961 | 11.2 | 115,699 | 1,863.3 | 1.6 | 60 | 2.33 |
| Amber Cr. | 272-702 | 21.6 | 200,041 | 2,399.3 | 1.2 | 48 | 0.26 |
| Central District |  |  |  |  |  |  |  |
| Hook Bay Cr. | 272-302 | 18.3 | 153,244 | 3,022.7 | 1.9 | 60 | 0.30 |
| Kumliun Cr. | 272-501 | 9.9 | 24,234 | 551.8 | 2.3 | 60 | 3.69 |
| North Fork Cr. | 272-514 | 30.6 | 128,837 | 1,925.4 | 1.5 | 60 | 0.20 |
| Perryville District |  |  |  |  |  |  |  |
| Ivanof Cr . | 275-406 | 28.2 | 256,723 | 1,971.7 | $<.1$ | 50 | 0.66 |
| Humpback Cr. | 275-502 | 15.6 | 110,294 | 1,564.7 | 1.4 | 60 | 0.46 |
| Western District |  |  |  |  |  |  |  |
| Ivan Cr. | 273-722 | 24.0 | 189,000 | 2,960.0 | 1.6 | 60 | 0.17 |
| Foot Bay Cr. | 273-802 | 2.0 | 17,885 | 891.5 | 5.0 | 57 | 0.60 |
| Spoon Cr. | 273-823 | 4.5 | 23,652 | 466.5 | 2.0 | 60 | 0.07 |
| Portage Cr. | 273-842 | 6.4 | 19,106 | 593.1 | 3.1 | 60 | 0.06 |

${ }^{a}$ Number of secondary units sampled.

Table 10. Pink salmon length and fecundity statistics from populations within the Kodiak and Chignik Management Areas, 1989.

| Stream |  | Length (mm) |  |  |  |  | Fecundity (No. of eqgs) |  |  |  |  | N |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Name | Area | Mean | Median | SD | Range |  | Mean | Median | SD | Range |  |  |
|  |  |  |  |  | Min. | Max. |  |  |  | Min. | Max. |  |
| Litnik R. | Kodiak | 472 | 470 | 26.6 | 420 | 530 | 1,445 | 1,457 | 255.3 | 930 | 2,127 | 50 |
| E. Paramanof Cr. | Kodiak | 465 | 465 | 19.4 | 416 | 501 | 1,312 | 1,304 | 232.5 | 862 | 1,960 | 50 |
| Akalura Cr. | Kodiak | 483 | 485 | 26.5 | 426 | 564 | 1,627 | 1,625 | 324.2 | 892 | 2,448 | 100 |
| Lake Bay Cr. | Chignik | 478 | 485 | 38.2 | 376 | 581 | 1,587 | 1,631 | 301.5 | 708 | 2,279 | 101 |

Table 11. Analysis of variance for the full model regression of fecundity on length and partitioning of sums of squares (SS) for the full model into two components: (1) SS explained by the reduced model alone; and, (2) the additional SS explained by the full model ("Full!'Reduced").

| Source | df | SS | MS | F |
| :---: | :---: | :---: | :---: | :---: |
| Full Model |  |  |  |  |
| Regression <br> Residual | 7 293 | $\begin{aligned} & 134,373 \times 10^{2} \\ & 159,028 \times 10^{2} \end{aligned}$ | $\begin{array}{r} 1,919,613.0 \\ 54,275.61 \end{array}$ | 35.36 |
| Reduced Model |  |  |  |  |
| Regression Residual | 1 299 | $103,802 \times 10^{2}$ $189,599 \times 10^{2}$ | $\begin{gathered} 103,802 \times 10^{2} \\ 63,410.946 \end{gathered}$ | 163.70 |
| Full\|Reduced |  |  |  |  |
| Regression <br> Residual | 6 293 | $30,571 \times 10^{2}$ $159,028 \times 10^{2}$ | $\begin{aligned} & 5,095 \times 10^{2} \\ & 54,275.61 \end{aligned}$ | 9.38 |



Fish Length (millimeters)

Figure 7. Pink salmon length-fecundity relationships from four populations studied, 1989.

Table 12. Egg retention statistics and spawner densities observed for Kodiak and Chignik area pink salmon populations, 1989.

| Stream |  | ```1989 Spawner Density (No. of Fish/m}\mp@subsup{}{}{2}\mathrm{ )``` | Egg Retention |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Mean |  |  | Range |  | Percent |  |
| Name | Number |  | Medi | an SD | Min. | Max. | zero eggs | N |
| Kodiak Area Streams |  |  |  |  |  |  |  |  |  |
| Narrows Cr . | 257-401 | 0.10 | 15.7 | 0.0 | 44.1 | 0.0 | 267 | 57.0 | 86 |
| Karluk R. | 255-101 | 0.13 | 75.1 | 1.0 | 177.8 | 0.0 | 1, 122 | 49.0 | 147 |
| Marka Cr. | 252-343 | 0.35 | 2.0 | 0.0 | 11.3 | 0.0 | 131 | 77.3 | 150 |
| E. Paramanof Cr. | 251-404 | 0.48 | 40.4 | 3.0 | 197.3 | 0.0 | 1.858 | 21.7 | 157 |
| Browns Cr. | 254-204 | 1.59 | 14.5 | 0.0 | 84.9 | 0.0 | 943 | 66.0 | 150 |
| Barling Cr. | 258-522 | 1.96 | 11.9 | 0.0 | 48.9 | 0.0 | 359 | 70.0 | 150 |
| Dog Salmon R. | 257-403 | 2.29 | 7.2 | 2.0 | 18.1 | 0.0 | 137 | 30.0 | 150 |
| Portage Cr. | 251-830 | 3.27 | 26.3 | 2.5 | 57.2 | 0.0 | 378 | 34.0 | 150 |
| Miam Cr. | 259-412 | 3.65 | 11.0 | 1.0 | 43.3 | 0.0 | 396 | 43.7 | 126 |
| American R. | 259-231 | 3.91 | 22.5 | 0.0 | 90.7 | 0.0 | 721 | 62.7 | 150 |
| Hurst Cr. | 259-414 | 4.35 | 71.3 | 2.0 | 211.5 | 0.0 | 1.249 | 28.0 | 150 |
| Saltry Cr. | 259-415 | 4.68 | 50.3 | 2.0 | 152.1 | 0.0 | 1,589 | 38.0 | 150 |
| Buskin R. | 259-211 | 5.69 | 24.0 | 0.0 | 94.6 | 0.0 | 765 | 58.9 | 146 |
| Terror R. | 253-331 | 6.64 | 32.4 | 1.0 | 117.9 | 0.0 | 1,097 | 45.4 | 152 |
| Kanatak Cr. | 262-802 | 6.88 | 18.3 | 0.0 | 59.2 | 0.0 | 402 | 65.3 | 150 |
| Baumans Cr. | 253-332 | 7.91 | 140.0 | 35.5 | 249.8 | 0.0 | 1,473 | 20.0 | 150 |
| Uganik R. | 253-122 | 8.76 | 19.3 | 0.0 | 54.8 | 0.0 | 385 | 59.3 | 150 |
| Deadman $R$. | 257-502 | 9.90 | 23.2 | 1.5 | 64.4 | 0.0 | 520 | 38.2 | 144 |
| Humpy Cr. ${ }^{\text {a }}$ | 257-701 | 10.38 | 132.0 | 6.0 | 221.2 | 0.0 | 1,167 | 39.1 | $64^{\text {a }}$ |
| Kaiugnak Cr. | 258-542 | 15.12 | 19.5 | 0.0 | 63.2 | 0.0 | 429 | 54.1 | 148 |
| Pillar Cr. | 259-102 | 16.90 | 22.7 | 0.0 | 93.8 | 0.0 | 987 | 52.4 | 147 |
| Danger Cr. | 252-332 | 46.73 | 9.4 | 0.0 | 24.9 | 0.0 | 243 | 38.8 | 49 |
| Seven R. | 258-701 | 60.88 | 146.0 | 55.0 | 193.8 | 0.0 | 920 | 22.3 | 148 |
| Chignik Area Streams |  |  |  |  |  |  |  |  |  |
| Ocean Beach Cr. | 272-801 | 0.17 | 1.9 | 0.0 | 7.4 | 0.0 | 59 | 72.2 | 108 |
| Ivan R. | 273-722 | 0.17 | 2.3 | 0.0 | 6.3 | 0.0 | 47 | 57.7 | 78 |
| Amber Bay Cr. | 272-702 | 0.26 | 6.7 | 0.0 | 54.1 | 0.0 | 655 | 57.3 | 150 |
| Hook Bay Cr. | 272-302 | 0.29 | 1.5 | 0.0 | 7.2 | 0.0 | 87 | 66.9 | 157 |
| Humpback Cr. | 275-502 | 0.46 | 23.4 | 1.0 | 146.4 | 0.0 | 1,432 | 43.6 | 156 |
| Foot Bay Cr . | 273-802 | 0.60 | 15.3 | 2.0 | 60.4 | 0.0 | 397 | 40.0 | 45 |
| Ivanof Cr. | 275-406 | 0.66 | 13.6 | 0.5 | 71.9 | 0.0 | 840 | 50.0 | 162 |
| Chigniagak 905 Cr . | 272-905 | 1. 64 | 16.6 | 0.0 | 70.8 | 0.0 | 690 | 64.4 | 146 |
| Chiginagak 904 Cr. | 272-904 | 1. 66 | 6.7 | 1.0 | 20.7 | 0.0 | 176 | 45.0 | 151 |
| Agripina Cr. | 272-961 | 2.33 | 60.0 | 2.0 | 159.9 | 0.0 | 872 | 32.2 | 146 |
| Kumulin Cr. | 272-501. | 3.69 | 8.3 | 0.0 | 39.6 | 0.0 | 343 | 58.6 | 152 |

${ }^{\text {a }}$ There were 84 individuals designated as being unspawned; these fish were excluded from analysis.

Kodiak populations had the largest range of egg retention and spawner densities. The Chignik egg retention data were more variable between populations.

Maximum likelihood parameter estimates (MLE) for the logistic model of spawner density on percent of sample with zero eggs was $6_{0}-0.042$ ( $95 \%$ CI -0.023 to 0.107 ). The estimated $b_{1}$ value was -0.118 ( $95 \%$ CI of -0.155 to -0.082 ; Figure 8). These parameter estimates were consistent with the hypothesized positive effect of spawner density on egg retention (i.e., as spawner density increases the probability of a fish retaining zero eggs decreases).

## Success Of 1989 Spawners and Brood Year

Observed 1990 preemergent $\mathrm{fry} / \mathrm{m}^{2}$ was significantly below that predicted by the $\mathrm{fry} / \mathrm{m}^{2}$ vs. $\mathrm{TP} / \mathrm{m}^{2}$ linear regression models for 23 Kodiak and 7 Chignik streams (Tables 13 and 14), indicating that spawner success was poor. Thirteen streams in Kodiak and two in Chignik did not show significant depression in fry $/ \mathrm{m}^{2}$ but eight of these had lower than predicted $\mathrm{fry} / \mathrm{m}^{2}$. An alternative method for both KMA \& CMA index streams of employing escapement to predict preemergent fry $/ \mathrm{m}^{2}$ was discarded (Figure 9).

During 1990, preemergent fry $/ \mathrm{m}^{2}$ was greater than the odd-year brood line historical average in 23 streams in Kodiak (Table 13). For 13 of these, fry/m² was greater than the upper limit of the $90 \%$ confidence intervals of the average $\mathrm{fry} / \mathrm{m}^{2}$. In view of \#SDfry $/ \mathrm{m}^{2}$ values (standardized residuals), which compare 1990 fry yield to past yield and variability in that yield, 1990 fry production was extremely high in the Paramanof, Uyak-202, and Saltery systems (Table 13; Figure 10). Fry yield was also very high in Seal Bay, Danger, and Narrows Creeks, even though the first of these three showed significantly depressed fry $/ \mathrm{m}^{2}$. Of the 13 other streams that had positive $\# \mathrm{SD}_{\mathrm{fry}} / \mathrm{m}^{2}$ values, 11 had significantly depressed fry $/ \mathrm{m}^{2}$ as compared to predicted values. The $1990 \mathrm{fry} / \mathrm{m}^{2}$ was less than the odd-year brood line average in 13 Kodiak streams, and was significantly low in 11 of these. The most negative $\# \mathrm{SD}_{\mathrm{fry}} / \mathrm{m}^{2}$ value was only 1.42 for Little Waterfall Creek. The average $1990 \mathrm{fry} / \mathrm{m}^{2}$ for index streams stratified by fishing district was greater than historical averages in all Kodiak districts except the Mainland, where the average of the historical averages was slightly higher. Average $1990 \mathrm{fry} / \mathrm{m}^{2}$ values were substantially greater for the Northwest Kodiak, Alitak, and Eastside Kodiak Districts, but only slightly higher for the Afognak and Northeast Kodiak Districts.

For the CMA populations, $1990 \mathrm{fry} / \mathrm{m}^{2}$ was greater than the odd-year average in only Ivanof River and Portage Creek. The Ivanof value was greater than the $90 \%$ CI of the average, but the fry $/ \mathrm{m}^{2}$ was significantly depressed as compared to the predicted value (Table 14). The 1990 fry $/ \mathrm{m}^{2}$ was less than average in the other seven streams, was significantly depressed in six of these, and was below the $90 \%$ CI of the average for the Ivan and Spoon Rivers. The deviations of $1990 \mathrm{fry} / \mathrm{m}^{2}$ from average fry $/ \mathrm{m}^{2}$ were low for Chignik streams; the greatest $\# \mathrm{SD}_{\mathrm{fry}} / \mathrm{m}^{2}$ value being that of Spoon Creek -1.56 (Table 14; Figure 10). The average of the $1990 \mathrm{fry} / \mathrm{m}^{2}$ values for Chignik streams were substantially lower than the average of the historical averages.


Figure 8. Pink salmon egg retention represented as percent of females sampled that retained no eggs as a function of spawner density.

Table 13. Results of linear regression analyses of 1990 preemergent fry sampling data ( 1989 brood year), and historical odd-year brood line live fry density data for the Kodiak Management Area.

|  |  | Linear Regression Results |  |  |  |  |  | Historical Odd-year Brood Line Fry/m ${ }^{2}$ Data |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Model $\mathrm{No}^{\text {a }}$. | Number <br> Sampling <br> Years | $r^{2}$ | 1990 Fry/m ${ }^{2}$ |  |  | Number <br> Years <br> Averaged | $\mathrm{Fry} / \mathrm{m}^{2}$ |  |  |  |
| Name Stream | Number |  |  |  | Predicted | Observed | $\begin{gathered} \mathrm{p}^{\mathrm{t}} \\ \text { Pred. }<\text { Obser. } \end{gathered}$ |  | Mean | SD | $\begin{aligned} & \text { Standardized } \\ & \text { Residuals } \end{aligned}$ | d 90\% Cl |
| Afognak District |  |  |  |  |  |  |  |  |  |  |  |  |
| E. Paramanof | 251-404 | 3 | 13 | 0.798 | 252.9 | 278.3 | 0.6046 | 11 | 21.16 | 25.2 | 10.20 | 7.39-34.92 |
| L. Waterfall | 251-822 | 2 | 8 | 0.682 | 532.2 | 70.0 | 0.0076 | 4 | 484.25 | 291.5 | -1.42 1 | 141.24-827.26 |
| Perenosa | 251-830 | 4 | 14 | 0.765 | 289.7 | 72.3 | 0.0016 | 12 | 219.13 | 126.3 | -1.16 1 | 153.64-284.61 |
| Seal Bay | 251-901 | 1 | 13 | 0.859 | 701.3 | 459.2 | 0.0054 | 6 | 315.04 | 236.7 | 2.341 | 120.29-509.79 |
| Danger | 252-332 | 1 | 21 | 0.849 | 259.9 | 292.5 | 0.7179 | 11 | 89.74 | 81.7 | 2.48 | 42.39-137.09 |
| Marka | 252-343 | 4 | 9 | 0.954 | 10.9 | 10.3 | 0.2896 | 0 |  |  |  |  |
| Average |  |  |  |  |  | $234.5{ }^{\text {d }}$ |  |  | $174.43{ }^{\text {e }}$ |  |  |  |
| Northwest Rodiak District |  |  |  |  |  |  |  |  |  |  |  |  |
| Sheratin | 259-371 | 3 | 22 | 0.578 | 153.3 | 39.8 | 0.0529 | 11 | 40.63 | 33.3 | -0.02 | 21.54-59.72 |
| Uganik | 253-122 | 1 | 22 | 0.720 | 132.4 | 9.4 | 0.0193 | 12 | 66.13 | 68.7 | -0.83 | 30.51-101.74 |
| Terror | 253-331 | 3 | 22 | 0.387 | 56.5 | 0.7 | 0.0036 | 12 | 11.45 | 12.8 | -0.83 | 4.78-18.11 |
| Bauman's | 253-332 | 1 | 23 | 0.915 | 843.6 | 240.7 | 0.0062 | 12 | 313.38 | 225.3 | -0.32 19 | 196.55-430.20 |
| Uyak-202 | 254-202 | 1 | 22 | 0.978 | 943.6 | 1,070.0 | >0.9995 | 12 | 257.64 | 157.7 | 5.151 | 175.88-339.39f |
| Uyuk-203 | 254-203 | 1 | 5 | 0.999 | 252.7 | 238.0 | 0.0024 | 6 | 207.59 | 271.9 | 0.11 | 0.00-431.29 ${ }^{\text {f }}$ |
| Zachar | 254-301 | 1 | 22 | 0.746 | 92.0 | 12.1 | 0.0007 | 12 | 10.16 | 17.9 | 0.11 | 0.91-19.42 |
| Average |  |  |  |  | 230.1 |  |  | $124.64{ }^{\text {e }}$ |  |  |  |  |
| Alitak Bay District |  |  |  |  |  |  |  |  |  |  |  |  |
| Narrows | 257-401 | 4 | 13 | 0.810 | 309.3 | 264.8 | 0.3671 | 12 | 91.77 | 69.1 | 2.50 | 55.92-127.62 |
| Dog Salmon | 257-403 | 4 | 22 | 0.982 | 159.5 | 179.9 | 0.7241 | 12 | 104.60 | 113.0 | 0.67 | 46.00-163.20 |
| Deadman | 257-502 | 4 | 22 | 0.985 | 664.4 | 546.9 | 0.0241 | 12 | 432.65 | 196.3 | 0.58330 | 330.86-534.44 |
| Humpy | 257-701 | 1 | 23 | 0.876 | 742.6 | 464.8 | 0.0008 | 12 | 220.05 | 173.5 | 1.41 | 126.11-313.99 |
| Average |  |  |  |  |  | 364.1 |  |  | $212.27{ }^{\text {e }}$ |  |  |  |
| Sastside Kodiak District |  |  |  |  |  |  |  |  |  |  |  |  |
| Kiliuda | 258-207 | 1 | 22 | 0.960 | 82.28 | 77.0 | 0.3925 | 10 | 34.62 | 40.4 | 1.05 | 11.47-57.76 |
| Barling | 258-522 | 1 | 20 | 0.782 | 132.25 | 51.7 | 0.0223 | 12 | 99.48 | 93.9 | -0.51 | 48.22-150.74 |
| Kaiugnak | 258-542 | 1 | 22 | 0.899 | 814.35 | 736.8 | 0.2127 | 12 | 241.35 | 308.2 | 1.61 | 81.57-401.13 |
| Seven Rivers Beaver | 258-701. | 1 | 22 | 0.668 | 556.18 | 383.5 | 0.0534 | 12 | 284.73 | 158.1 | 0.62 | 202.76-366.70 f |
| Beaver | 259-365a | 1 | 9 | 0.995 | 573.48 | 549.9 | 0.2402 | 4 | 473.61 | 403.5 | 0.19 | 0.00-948.33 ${ }^{\text {f }}$ |
| Miam | 259-412 | 1 | 12 | 0.667 | 118.26 | 118.6 | 0.5025 | 0 |  |  |  |  |
| Hurst | 259-414 | 1 | 21 | 0.944 | 287.7 | 102.8 | $<0.0005$ | 10 | 49.36 | 56.4 | 0.95 | 16.65-82.06 |
| Saltery | 259-415 | 1 | 22 | 0.644 | 371.9 | 521.5 | 0.9935 | 12 | 94.73 | 67.5 | 6.32 | 59.73-129.72 |
| Average |  |  |  |  |  | 317.739 |  |  | $158.02{ }^{\text {e }}$ |  |  |  |

[^1]Table 13. (page 2 of 2 )

|  |  | Linear Regression Results |  |  |  |  |  | _Historical Odd-year Brood, Line Fry/m ${ }^{2}$ Data |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Model $\mathrm{NO}^{\text {a }}$ | NumberSamplingYears | $\mathrm{r}^{2}$ | $1990 \mathrm{Fry} / \mathrm{m}^{2}$ |  |  | Number Years Averaged | Fry/m ${ }^{2}$ |  |  |  |
| ${ }_{\text {Name }}$ Streatn | Number |  |  |  | Predicted | Observed | $\frac{\mathrm{p}^{2}}{\text { Pred. }<\text { Obser. }}$ |  | Mean | SD | $\begin{aligned} & \text { Standardized } \\ & \text { Residuals }{ }^{c} . \end{aligned}$ | 90\% CI |
| Northeast Kodiak District |  |  |  |  |  |  |  |  |  |  |  |  |
| Monashka | 259-101 | 4 | 9 | 0.576 | 75.4 | 2.7 | 0.0085 | 5 | 39.65 | 48.8 | -0.76 | 22.87-56.43 |
| Buskin | 259-211 | 3 | 22 | 0.734 | 268.3 | 89.7 | 0.0326 | 12 | 143.01 | 122.1 | -0.44 | 79.72-206.30 |
| American | 259-231 | 1 | 23 | 0.978 | 325.5 | 256.0 | 0.0021 | 12 | 142.93 | 124.4 | 0.91 | 78.43-207.42 |
| Sid Old's | 259-242 | 1 | 23 | 0.907 | 104.4 | 116.4 | 0.7024 | 12 | 93.02 | 70.2 | 0.33 | 56.63-129.41 |
| Average |  |  |  |  |  | 116.2 |  |  | $115.75{ }^{\text {e }}$ |  |  |  |
| Mainland District |  |  |  |  |  |  |  |  |  |  |  |  |
| Kukak | 262-271 | 1 | 6 | 0.806 | 49.4 | 6.7 | 0.0320 | 3 | 52.09 | 47.6 | -0.95 | 0.00-132.34 ${ }^{\text {f }}$ |
| Missak | 262-271 | 1 | 12 | 0.939 | 226.1 | 131.7 | 0.0008 | 6 | 99.51 | 73.5 | 0.44 | 39.06-159.96 |
| Kinak | 262-451 | 1 | 12 | 0.886 | 238.0 | 161.7 | 0.0030 | 6 | 142.62 | 80.9 | 0.24 | 76.05-209.17 |
| Dakavak | 262-551 | 4 | 14 | 0.983 | 18.5 | 0.4 | $<0.0005$ | 7 | 59.84 | 51.6 | -1.07 | 21.94-97.74 |
| Kashvik | 262-604 | 2 | 14 | 0.844 | 7.1 | 0.8 | 0.1497 | 7 | 44.50 | 59.5 | -0.73 | 0.79-88.21 |
| Alinchak | 262-651 | 1 | 10 | 0.843 | 137.5 | 27.3 | 0.0025 | 6 | 18.68 | 28.0 | 0.31 | 4.21-33.15 ${ }_{\text {f }}$ |
| Portage | 262-702 | 1 | 10 | 0.928 | 3.4 | 2.3 | 0.4890 | 4 | 19.52 | 35.9 | -0.48 | 0.00-61.74 |
| Kanatak | 262-802 | 1 | 9 | 0.964 | 176.9 | 186.7 | 0.6421 | 5 | 173.78 | 161.5 | 0.08 | 19.76-327.80 ${ }_{\text {f }}$ |
| Big Creek | 262-851 | 1 | 9 | 0.998 | 155.9 | 131.7 | 0.0013 | 5 | 115.03 | 166.2 | 0.10 | 0.00-273.51 ${ }^{\text {f }}$ |
| Average |  |  |  |  |  | 72.1 |  |  | $81.09{ }^{\text {e }}$ |  |  |  |

${ }^{\mathrm{a}}$ Linear regression model was live fry density ( $\mathrm{fry} / \mathrm{m}^{2}$ ) vs. total progeny density ( $\mathrm{TP} / \mathrm{m}^{2}$ ) with or without square root or logarithm transformation of either or both variables. The model numbers represent (dependent variable first): (1) fry/m ms . $\mathrm{TP} / \mathrm{m}^{2}$; (2) square root $\mathrm{fry} / \mathrm{m}^{2}$ vs. $\mathrm{TP} / \mathrm{m}^{2}$; (3) square root $\mathrm{fry} / \mathrm{m}^{2}$ vs. square root $\mathrm{TP} / \mathrm{m}^{2}$; and (4) logarithm fry $/ \mathrm{m}^{2}$ vs. logarithm $\mathrm{TP} / \mathrm{m}^{2}$.
b The probability, according to $1990 \mathrm{TP} / \mathrm{m}^{2}$ and the linear regression relationship, that $\mathrm{fry} / \mathrm{m}^{2}$ in 1990 would be $\leq$ the observed 1990 $\mathrm{fry} / \mathrm{m}^{2}$. Probabilities $<0.10$ indicate significantly low $1990 \mathrm{fry} / \mathrm{m}^{2}$.
${ }^{c}$ Standarized Residuals $=\left(1990 \mathrm{fry} / \mathrm{m}^{2}\right.$ - historical mean fry $/ \mathrm{m}^{2}$ ) $/$ standard deviation of historical mean $\mathrm{fry} / \mathrm{m}^{2}$.
d Average calculated excluding Marka Creek.
e Weighted average.
${ }^{\mathrm{f}}$ Zero was assigned to lower confidence limits that were negative numbers.
g Average calculated excluding Miam Creek.

Table 14. Results of linear regression analyses of 1990 preemergent fry sampling data (1989 brood year), and historical odd-year brood line live fry density data for the Chignik Management Area.

| Name Stream | Number | Linear Reqression Results |  |  |  |  |  | Historical Odd-year Brood Line Fry $/ \mathrm{m}^{2}$ Data |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Number |  |  | 1990 Fry/m² |  |  | $\begin{aligned} & \text { Number } \\ & \text { Years } \\ & \text { Averaged } \end{aligned}$ | $\text { Fry } / \mathrm{m}^{2}$ |  |  |  |
|  |  | Model No ${ }^{\text {a }}$. | $\begin{aligned} & \text { Sampling } \\ & \text { Years } \end{aligned}$ | $\mathrm{r}^{2}$ | Predicted | Observed | $\begin{gathered} \text { Prob }{ }^{\square} \\ \text { Pred. }<\text { Obser. } \end{gathered}$ |  | Mean | SD | andardized Residuals ${ }^{\text {c }}$ | . 90\% CI |
| Hook Bay | 272-302 | 1 | 14 | 0.820 | 69.2 | 0.0 | $<0.0005$ | 11 | 18.77 | 27.8 | -0.67 | 0.00-44.71 ${ }^{\text {a }}$ |
| Kumliun | 272-501 | 1. | 5 | 0.974 | 75.2 | 2.6 | 0.0009 | 4 | 45.93 | 40.3 | -1.08 | 0.00-120.95 ${ }^{\text {a }}$ |
| Chiginagak 905 | 272-905 | 1 | 6 | 0.995 | 261.1 | 180.8 | 0.0007 | 2 | 209.22 | 83.2 | -0.34 0 | 0.00-463.29 ${ }^{\text {a }}$ |
| Ivan | 273-722 | 3 | 17 | 0.958 | 33.8 | 13.7 | 0.0498 | 9 | 111.61 | 113.4 | -0.86 3 | 31.61-191.61 |
| Foot Bay | 273-802 | 1 | 11 | 0.977 | 304.3 | 227.8 | 0.0148 | 6 | 265.77 | 192.7 | -0.20 10 | 22.58-428.95 |
| Spoon | 273-823 | 3 | 9 | 0.932 | 8.1 | 0.0 | 0.0206 | 5 | 99.09 | 63.6 | -1.56 3 | 39.09-159.08 |
| Portage | 272-842 | 1 | 17 | 0.576 | 41.9 | 64.2 | 0.8035 | 6 | 40.34 | 40.0 | 0.60 | 14.49-66.19 |
| Ivanof | 275-406 | 3 | 17 | 0.983 | 267.2 | 200.6 | $<0.0005$ | 9 | 99.66 | 94.3 | 1.073 | 39.44-159.88 |
| Humpback | 275-502 | 3 | 17 | 0.979 | 79.0 | 65.6 | 0.2900 | 9 | 185.38 | 249.7 | -0.48 1 | 14.34-356.41 |
| Average |  |  |  |  |  | 68.7 |  |  | 119.52 |  |  |  |

${ }^{\mathrm{a}}$ Linear regression model was live fry density $\left(\mathrm{fry} / \mathrm{m}^{2}\right.$ ) as the dependent variable versus total progeny density (TP/m2) as the independent variable, with or without square root or logarithm transformation of either or both variables. The model numbers represent: (1) fry $/ \mathrm{m}^{2}$ vs. $\mathrm{TP} / \mathrm{m}^{2}$; (2)square root fry $/ \mathrm{m}^{2}$ vs TP/m ${ }^{2}$; (3) square root fry $/ \mathrm{m}^{2}$ vs square root $\mathrm{TP} / \mathrm{m}^{2}$; and (4) logarithm fry $/ \mathrm{m}^{2}$ vs logarithm $\mathrm{TP} / \mathrm{m}^{2}$.
b The probability, according to $1990 \mathrm{TP} / \mathrm{m}^{2}$ and the linear regression relationship, that $\mathrm{fry} / \mathrm{m}^{2}$ in 1990 would be $\leq$ the observed 1990 $\mathrm{fry} / \mathrm{m}^{2}$. Probabilities $<0.10$ indicate significantly low $1990 \mathrm{fry} / \mathrm{m}^{2}$.
c Zero assigned to lower confidence limits that were negative numbers.
d Weighted average.


Figure 9. Pink salmon preemergent fry/mas a function of escapement by index stream: (A) Kodiak; and (B) Chignik.


Figure l0. Preemergent fry standardized residuals as a function of 1989 escapements: (A) Kodiak; and (B) Chignik.

There were a wide range of spawner densities experienced within the KMA index streams during 1989 with most escapements higher than the historical average; differences ranged from small fractions to large numbers of standard deviations (Table 15). Few streams had escapements that were lower than average, all by less than one SD. For CMA streams, escapements were below the 1963-1989 escapement average in six of nine preemergent index streams with Chiginagak905 and Hook Bay Creeks receiving escapements greater than one SD above average (Table 16).

Of the Mann-Whitney comparisons of spawner abundance indices between streams that had depressed fry $/ \mathrm{m}^{2}$ (i.e. low fry $/ \mathrm{m}^{2}$ relative to $\mathrm{TP} / \mathrm{m}^{2}$ ) and streams that did not, only the comparison of spawner densities between these stream groups in Chignik yielded significant evidence ( $\mathrm{P}=0.087$ ) that depression was associated with increased spawner density (Table 17). However, the comparison of \#SD ${ }_{\text {esc }}$ values between these Chignik groups provided a test statistic far below the critical value. Comparisons of these indices between stream groups for Kodiak produced test statistics close to but below the critical values. Moderate and high spawner density stream groups in Kodiak had significantly different $1989 \mathrm{fry} / \mathrm{m}^{2}$ values ( $\mathrm{P}=0.025$ ). The average of the $1989 \mathrm{fry} / \mathrm{m}^{2}$ values found in moderate spawner density streams (91.796) was much lower than the average value for high spawner density streams (291.41). Comparison of \#SD ${ }_{\text {esc }}$ vs. \#SD $\mathrm{f}_{\mathrm{fry}} / \mathrm{m}^{2}$ values between moderate and high escapement stream groups in Kodiak yielded a test statistic well below the critical value.

## Spawner-Recruit

The Ricker and Beverton-Holt models fit to the Kodiak return per spawner data are essentially equivalent. What is important in the context of this study is that the Kodiak data indicates that for both models the $b_{2}$ or density dependent parameter is non-zero (Table 18; Figure 11). In contrast, parameter estimates for the Ricker and Beverton-Holt models based on Chignik data do not differ significantly from zero, suggesting that the dependence of return on number of spawners is weak over the range of observed data (Table 18; Figure 12). The predicted KMA 1991 pink salmon return from the odd-year midpoint escapement goal of 4.67 million (Ricker model) provided a return of 11.9 million fish or 2.5 returns per spawner ( $95 \%$ prediction interval of 3.6 to 24.4 million; 0.8 to 5.2 returns per spawner). The desired escapement goal of 7.0 million produced a predicted return of 15.7 million or 2.2 returns per spawner ( $95 \%$ prediction interval: 5.8 to 29.5 million; 0.8 to 4.2 returns per spawner). The maximum difference between replacement and the fitted curve occurs around 7.5 million (Figure 11). The CMA predicted return from the 0.7 million escapement goal was 1.3 million or 1.8 returns per spawner ( $95 \%$ prediction interval: 0.3 to 5.8 million; 0.4 to 8.3 returns per spawner).

Table 15. Preemergent fry sampling index streams classified as having moderate and high escapements within the Kodiak Management Area, 1989 brood year.

| Stream |  | Odd-Year Escapements (Numbers of Fish) |  |  | Escapement Years <br> Averaged ${ }^{\text {a }}$ (1963-1989) | SD |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Name | Number | Mean | SD | 1989 |  | Esc. ${ }^{\text {D }}$ | Fry/m $\mathrm{m}^{2 C}$ |
| Streams Classified as having Moderate Escapement in 1989 |  |  |  |  |  |  |  |
| Narrows Cr. | 257-401 | 9,540 | 7,926 | 2,097 | 11 | -0.939 | 2.50 |
| Seal Bay Cr. | 251-901 | 16,627 | 17,281 | 6,000 | 8 | -0.615 | 2.34 |
| Kinak Cr. | 262-451 | 70,124 | 51,702 | 54,327 | 9 | -0.306 | 0.24 |
| L. Waterfall | 251-822 | 125,736 | 161,451 | 106,347 | 4 | -0.120 | $-1.42$ |
| Kukak Cr | 262-271 | 4,767 | 8,113 | 3,802 | 11 | -0.119 | -0.95 |
| Dakavak Cr. | 262-551 | 40,966 | 46,037 | 46,037 | 14 | 0.542 | -1.07 |
| E.Paramanof Cr. | 251-404 | 12,029 | 14,287 | 20,056 | 12 | 0.562 | 10.20 |
| Danger Cr. | 252-332 | 20,047 | 21,711 | 39,207 | 14 | 0.883 | 2.48 |
| Sid Olds R. | 259-242 | 76,654 | 46,612 | 139,668 | 12 | 1. 352 | 0.33 |
| Perenosa Cr. | 251-830 | 43,115 | 37,596 | 95,700 | 12 | 1.399 | -1.16 |
| Monashka Cr. | 259-101 | 5,718 | 5,606 | 14,720 | 6 | 1.606 | -0.76 |
| Barling Cr. | 258-522 | 73,541 | 43,206 | 147,423 | 14 | 1.710 | -0.51 |
| Saltry Cr. | 259-415 | 106,916 | 58,975 | 214,541 | 14 | 1.825 | 6.32 |
| Streams Classified as having High Escapement in 1989 |  |  |  |  |  |  |  |
| Kiliuda Cr. | 258-207 | 19,501 | 24,151 | 69,762 | 12 | 2.081 | 1.05 |
| Zachar Cr. | 254-301 | 43,056 | 27,337 | 103,983 | 14 | 2.229 | 0.11 |
| Buskin R. | 259-211 | 82,622 | 79,360 | 272,785 | 14 | 2.396 | -0.44 |
| Dog Salmon R. | 257-403 | 117,882 | 68,953 | 315,559 | 14 | 2.867 | 0.67 |
| Missak Cr. | 262-402 | 7,765 | 8,520 | 36,120 | 10 | 3.328 | 0.44 |
| Sheratin R . | 259-371 | 46,123 | 20,818 | 119,943 | 6 | 3.546 | -0.02 |
| American R. | 259-231 | 83,179 | 75,003 | 365,383 | 14 | 3.763 | 0.91 |
| Kanatak Cr . | 262-802 | 32,785 | 28,689 | 158,350 | 11 | 4.377 | 0.08 |
| Uyak 202 R . | 254-202 | 170,582 | 97,247 | 645,157 | 14 | 4.880 | 5.15 |
| Uyak 203 R. | 254-203 | 59,559 | 48,415 | 303,600 | 12 | 5.041 | 0.11 |
| Kaiugnak Cr. | 258-542 | 25,997 | 16,997 | 118,323 | 14 | 5.432 | 1.61 |
| Hurst Cr. | 259-414 | 23,142 | 20,493 | 142,535 | 11 | 5.826 | 0.95 |
| Kashvik Cr. | 262-604 | 57,073 | 36,251 | 282,600 | 11 | 6.221 | -0.73 |
| Terror R. | 253-331 | 95,579 | 45,623 | 410,913 | 14 | 6.912 | -0.83 |
| Portage Cr. | 262-702 | 17,682 | 16,911 | 138,880 | 10 | 7.167 | -0.48 |
| Uganik R. | 253-122 | 125,317 | 92,153 | 795,899 | 14 | 7.277 | -0.83 |
| Humpy Cr. | 257-701 | 250,989 | 173,451 | 1,783,196 | 14 | 8.834 | 1.41 |
| Deadman R. | 257-502 | 187,443 | 101,643 | 1,093,543 | 14 | 8.915 | 0.58 |
| Bauman's Cr. | 253-332 | 21,903 | 22,980 | 353,333 | 13 | 14.423 | -0.32 |
| Seven River's | 258-701 | 140,099 | 70,374 | 1,192,683 | 14 | 14.957 | 0.62 |
| Alinchak R. | 262-651 | 26,444 | 29,894 | 541,827 | 13 | 17.241 | 0.31 |
| Big Creek | 262-851 | 99,027 | 107,563 | $2,078,433$ | 13 | 18.402 | 0.10 |

${ }^{\text {a }}$ Escapement years are odd-years only for which escapement counts were available.
b ((1989 escapement - mean escapement)/ standard deviation of mean escapement).
c $\left(\left(1990 \mathrm{fry} / \mathrm{m}^{2}-\right.\right.$ mean $\left.\mathrm{fry} / \mathrm{m}^{2}\right) /$ standard deviation of mean fry $\left./ \mathrm{m}^{2}\right)$.

Table 16. Preemergent fry sampling index streams classified as having moderate and high escapements within the Chignik Management Area, 1989 brood year.

| Stream |  | Odd-Year Escapements (Numbers of Fish) |  |  | $\begin{gathered} \text { Escapement } \\ \text { Years } \\ \text { Averaged } \\ (1963-1989)^{\text {a }} \end{gathered}$ | SD |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Name | Number | Mean | SD | 1989 |  | Esc. ${ }^{\circ}$ | Fry/m ${ }^{2 C}$ |
| Streams Classified as having Moderate Escapement in 1989 |  |  |  |  |  |  |  |
| Spoon Cr. | 273-823 | 7,237 | 5,685 | 1,700 | 10 | -0.974 | -1.56 |
| Ivan Cr. | 273-722 | 211,287 | 185,450 | 320,000 | 10 | -0.967 | -0.86 |
| Foot Bay Cr. | 273-802 | 20,505 | 15,015 | 10,800 | 10 | -0.646 | -0.20 |
| Portage Cr | 273-842 | 7,930 | 10,755 | 1,200 | 10 | -0.626 | 0.60 |
| Cape Kumliun Cr. | 272-501 | 124,223 | 103,381 | 89,499 | 8 | -0.336 | -1.08 |
| Humpback Cr. | 275-502 | 59,432 | 43,164 | 51,000 | 10 | -0.195 | -0.48 |
| Ivanof R. | 275-406 | 128,598 | 89,116 | 168,403 | 10 | 0.447 | 1.07 |
| Chiginagak 905 Cr . | 272-905 | 33,157 | 38,645 | 89,000 | 10 | 1.445 | -0.34 |
| Streams Classified as having High Escapement in 1989 |  |  |  |  |  |  |  |
| Hook Bay Cr. | 272-302 | 13,865 | 14,018 | 45,000 | 10 | 2.221 | -0.67 |

a Escapement years are odd-years only for which escapement counts were available.
b ((1989 escapement - mean escapement)/ standard deviation of mean escapement).
$\mathrm{c}\left(\left(1990 \mathrm{fry} / \mathrm{m}^{2}\right.\right.$ - mean fry $\left./ \mathrm{m}^{2}\right) /$ standard deviation of mean fry $\left./ \mathrm{m}^{2}\right)$.

Table 17. Results of Mann-Whitney "U" test comparisons for preemergent fry sampling program index stream groups, Kodiak and Chignik Management Areas.

| Stream Groups |  | Compared | Management Area | $\begin{aligned} & \text { Ranked } \\ & \text { Test } \\ & \text { Variable } \end{aligned}$ | $\begin{gathered} \text { Test }^{b} \\ \text { Statistic } \end{gathered}$ | $\begin{gathered} \text { Critical } \\ \text { Value } \\ (P=0.10) \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\begin{gathered} \text { Nondepressed } \\ \text { fry/m } \\ (n=14) \end{gathered}$ | vs. | $\begin{gathered} \text { Depressed } \\ \text { fry/m } \\ (n=23) \end{gathered}$ | Kodiak | \#SD ${ }_{\text {esc }}$ | 193 | 203 |
| $\begin{aligned} & \text { Nondepressed } \\ & \mathrm{fry} / \mathrm{m}_{2} \\ & (\mathrm{n}=2) \end{aligned}$ | vs. | $\begin{gathered} \text { Depressed } \\ \text { fry/m } \\ (n=7) \end{gathered}$ | Chignik | \#SD ${ }_{\text {esc }}$ | 7 | 10 |
| $\begin{gathered} \text { Nondepressed } \\ f r y / m^{2} \\ (n=14) \end{gathered}$ | vs. | $\begin{gathered} \text { Depressed } \\ \text { fry/m} \\ (n=21) \end{gathered}$ | Kodiak | Spawner Density | 183 | 185 |
| $\begin{gathered} \text { Nondepressed } \\ \text { fry/m2} \\ (n=14) \end{gathered}$ | vs. | $\begin{gathered} \text { Depressed } \\ \text { fry/m} \\ (n=21) \end{gathered}$ | Chignik | Spawner Density | $(p=0.0875)$ | 10 |
| Moderate <br> Spawner <br> Density $(n=18)$ |  | High <br> Spawner <br> Density $(n=21)$ | Kodiak | $\mathrm{fry} / \mathrm{m}^{2}$ | $\begin{gathered} 270 \\ (p=0.0246) \end{gathered}$ | 248 |
| Moderate Escapement ( $\mathrm{n}=13$ ) |  | High <br> Escapement $(\mathrm{n}=22)$ | Kodiak | $\# S D_{f r y} / m^{2}$ | 151 | 192 |

a $\#_{S D} \mathrm{esc}=(1989$ escapement- historical average escapement)/ standard deviation of historical escapements; spawner density $=1989$ escapement/ estimated total available spawning habitat: $\# \mathrm{SD}_{\mathrm{fry}} / \mathrm{m}^{2}=\left(1989 \mathrm{fry} / \mathrm{m}^{2}\right.$-historical average $\left.\mathrm{fry} / \mathrm{m}^{2}\right) /$ standard deviation of historical $\mathrm{fry} / \mathrm{m}^{2}$ values.
b
Test statistic value greater than the critical value indicates significant difference between the compared groups; significance level is stated in parenthesis.

Table 18. Ricker and Beverton-Holt model's parameter estimates for odd-year brood line (1963-1989) spawner-recruit data from the Kodiak and Chignik Management Areas.

| Management <br> Area | Model | Parameter |  | Number | Estimate | SE |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |



Figure 11. Spawner-recruit curves (A) Ricker model with $95 \%$ prediction intervals and (B) Beverton-Holt model for Kodiak Management Area odd-year broodline pink salmon, 1963-1989.


Figure l2. Spawner-recruit curves for Chignik Management Area odd-year broodine pink salmon, l963-1989.

## DISCUSSION

Stream life estimates generated from the various methods employed averaged 8.4 to 17.9 days and are similar to estimates generated from other studies (Perrin and Irvine 1990; Sharr and Sharp 1990). Pink salmon stream life appeared to decrease as the run progressed which is consistent with results from studies conducted in Southeast Alaska (Kingsbury 1977; Thomason and Jones 1984). Estimates generated using tagging data from Prince William Sound ranged from 7.2 to 21.9 days, whereas stream life reported by weekly strata had a grand mean of 13.9 days (range 9.9 d to 17.1 d ); estimates decreased over time within a system (Sharr and Sharp 1990).

Stream life estimate variability between methods both within and among streams made selection of a single value difficult. Tagging data estimates adjusted for proportion of detectable tags had the least amount of perceived bias when compared to other methods. Given the wide range of estimated values and the influence stream life has on escapement estimation using the AUC approach and aerial survey data, we felt using a 15 d value was prudent. Had we incorporated more streams of various sizes within our study, a suite of values tailored to run timing and stream size similar to the approach used by Sharr and Sharp (1990) would have been instituted. Regardless, Perrin and Irvine (1990) suggest caution be exercised when using stream life estimates for escapement estimation which transcend areas, streams, or years.

Total available spawning habitat estimates for KMA index streams ranged from less than 1,000 $\mathrm{m}^{2}$ (Danger Cr.) to $829,515 \mathrm{~m}^{2}$ (Karluk R.). For some streams, estimates were considered approximate owing to incomplete spawning ground surveys. Streams having CV values greater than $10 \%$ were those where habitat surveys were incomplete. Spawning habitat estimates generated for CMA streams varied from 17,000 to $250,000 \mathrm{~m}^{2}$; coefficients of variation were all less than $10 \%$. Spawner density ( $\mathrm{D}_{\mathrm{i}}$ ) estimates are precise as indicated by the spawning habitat CV values. However, our inability to accurately stratify populations which have bimodal run timing into early and late run components suggests that Di estimates for some spawning stocks might be overestimates. Alternatively, some spawning reaches had densities that were probably greater than those expressed in terms of the entire stream, as within stream direct measures for specific spawning areas were not estimated. In light of the escapement and total available spawning habitat estimates, the $\mathrm{D}_{\mathrm{i}}$ 's are not without error. We believe this error is relatively minor and imparts negligible influence on the egg retention and preemergent live fry analyses.

Fecundity estimates and the variability encountered between the four populations corresponds well with information on pink salmon presented by Heard (1991). Heard states that fecundity values deviate considerably between stocks, years, and regions. Therefore, our results are not atypical. Fecundity-length relationships for the four populations investigated were significantly different from one another. The lack of a single fecundity-length relationship severely limited our analysis of egg retention and preemergent fry survival.

For the egg retention data collected (Kodiak and Chignik populations combined), we found a highly variable relationship between mean egg retention and spawner density. A trend of increasing mean egg retention as a function of spawner density for most populations (Kodiak and

Chignik combined) was suggested by the data. However, the number of eggs retained was highly variable within samples and a known, but unspecifiable, relationship between fecundity and length further reduced clarity of this perceived trend. Interpretability and model fit was improved by transforming egg retention into a binary variable i.e., a value of zero when no eggs were retained and a value of one when eggs were retained. The index of egg retention for each population was expressed as the percent of each sample having zero eggs retained. Mean egg retention for the Kodiak pink salmon populations ranged from 1.5 to 146.0 eggs per female. Obtaining a statistically significant parameter estimate indicated the existence of a positive effect of spawner density on egg retention. Invariably, analyses were biased by omitting fish classified as "unspawned". However, ascertaining whether a fish had attempted to spawn prior to death or had died from kidney failure, oxygen depravation, or an epizootic proved unrealistic.

Few investigators have reported a relationship between spawner density (number of fish $/ \mathrm{m}^{2}$ ), egg retention, and subsequent fry survival with pink salmon. Hanavan and Skud (1954) showed that for in-situ pen enclosures with density controlled at 1.68 fish $/ \mathrm{m}^{2}$, mean egg retention was 409.4 eggs/female, while for densities of $3.7 \mathrm{fish} / \mathrm{m}^{2}$ egg retention averaged $624.2 \mathrm{eggs} /$ female. Corresponding egg to fry survival was reported as $10.1 \%$ and $7.7 \%$, respectively. Hanavan and Scud (1954) cautioned against using this information for calculating optimum spawner densities, but felt that crowding was influential on subsequent fry survival. No causal mechanism other than mechanical disruption of previously deposited fertilized eggs was discussed. McNeil (1964), citing in-situ experiments conducted within two Southeast Alaska streams, reported that spawning success (measured as number of females spawning successfully within a given area) was greatest at 330 females per $100 \mathrm{~m}^{2}$ of spawning habitat. Schroder (1973), relating density to spawning success with Chum salmon reported that spawning pairs were free from aggressive behavior only $8 \%$ of the time at spawner densities of 0.9 females $/ \mathrm{m}^{2}$, while mating activity decreased or ceased at densities above this level. For sockeye salmon data is limited, but Manzer and Miki (1986) have suggested that increased spawner density may cause increased egg retention.

Heard (1991) reports that pink salmon egg retention as a percentage of fecundity ranges from less than $1 \%$ to $40 \%$. A value intermediate between these two would limit egg deposition and reproductive potential for a population. We found that mean egg retention for the streams with the highest spawner densities (140, 132, and $146 \mathrm{eggs} /$ individual with spawner densities of 7.9 , 10.3 and 60.9 fish $/ \mathrm{m}^{2}$, respectively) when evaluated using a fecundity value of 1,400 eggs/female, represent less than $11 \%$ of each populations total egg capacity. This would represent an insignificant impact on reproductive potential in light of superimposition and other avenues of egg loss. However, if populations with the highest levels of egg retention also had substantial losses of eggs due to superimposition, then combined they would represent a substantial loss to a populations potential egg deposition. Results from the preemergent fry study component discount that this occurred for KMA spawning populations during 1989.

The foregone salmon harvests of 1989 resulted in the largest escapement on record for the Kodiak Management Area (L. Malloy, ADF\&G Kodiak, pers. comm); only five Kodiak index streams had less than average escapement. Brood year spawner success (predicted live fry $/ \mathrm{m}^{2}$ from $\mathrm{TP} / \mathrm{m}^{2}$, relative to observed 1990 live $\mathrm{fry} / \mathrm{m}^{2}$ ) by fishing district was apparently very good in the Northwest Kodiak, Alitak, and Eastside Kodiak Districts, and near average within the

Afognak, Northeast Kodiak, and Mainland Districts (Table 13). The 1989 brood year seems to have had poor success in Chignik but escapement for streams in this area were generally low (Mike Thompson, Alaska Department of Fish and Game Anchorage, personal communication). Brood year success was apparently good to exceptionally good in most Kodiak index streams (when evaluated independently) but was good in only two Chignik index streams. Some Kodiak streams and most Chignik streams showed only fair to low brood year success but none had particularly poor success. Significant depression in $1990 \mathrm{fry} / \mathrm{m}^{2}$ relative to $\mathrm{TP} / \mathrm{m}^{2}$ indicated poor spawner success in most Kodiak and Chignik streams. Evidently, only the Uyak-202 River and Saltery Creek had exceptionally good spawner success.

Of the 23 Kodiak streams that had depressed spawner success, 12 had good brood year success (better than average fry $/ \mathrm{m}^{2}$ ) and 11 had poor brood year success (less than average $\mathrm{fry} / \mathrm{m}^{2}$ ), suggesting that overescapement with density dependent mortality occurred in some streams. High spawner density streams had significantly higher fry/ $\mathrm{m}^{2}$ than moderate spawner density streams because the moderate group contained streams with very low spawner densities and because the cumulative spawner density goal ( $1.4 \mathrm{fish} / \mathrm{m}^{2}$ ) used to stratify the streams into the two groups may have been too low. Had the cut-off point been higher, perhaps 3.0, the two stream groups would have been much more similar in fry yields, suggesting that the Kodiak midpoint escapement goal is too low to obtain maximum fry yield. Moderate and high escapement stream groups were stratified at a level that was probably more extreme, and these groups did not differ significantly in their deviations from average fry $/ \mathrm{m}^{2}$. A well defined predictive relationship between parent year escapement and preemergent fry $/ \mathrm{m}^{2}$ produced would have undoubtedly simplified the analyses.

Although Kodiak pink salmon escapements were very large in 1989 and the spawner abundance indices tended to be somewhat higher for streams that showed significant depression than for streams that did not, there was no conclusive evidence that depression in fry $/ \mathrm{m}^{2}$ was directly caused by spawner density. Most Chignik index streams had low spawner abundance and depressed spawner success, although environmental conditions in this area are generally more extreme than in Kodiak and may have resulted in depressed 1989 spawner success. However, the spawner abundance indices quantify spawner abundance in entire streams and may not reflect spawner abundance in the sampled spawning riffles. Only 4 Kodiak streams had predicted 1990 fry $/ \mathrm{m}^{2}$ lower than the historical average $\mathrm{fry} / \mathrm{m}^{2}$, suggesting that egg deposition was greater than usual in most sampled spawning riffles.

Donnelly (1983) and Eggers et al. (1991) have reported significant density dependent relationships for Kodiak pink salmon populations. Both the egg retention and preemergent fry study components suggest that a density dependent response occurred between spawning and emergent fry for some Kodiak systems. This is partially substantiated with existence of a statistically significant density dependent parameter estimate for both the Ricker and BevertonHolt models as applied to the Kodiak data.

The KMA and CMA catch data are assumed to be precise. Escapements estimated using a 15 d stream life value either directly (AUC approach) or indirectly (expansion factors) coupled with error associated with aerial survey data are of potential concern. Donnelly (1983) used an expansion factor of 2.5 to convert aerial survey peak counts to estimated total escapement; expansion factors used herein, when combined were similar to this value. Sharr and Sharp
(1990) reported from preliminary analysis that escapements not corrected for stream life and aerial survey bias were $38 \%$ below corrected estimates. Walters and Ludwig (1981) consider escapement measurement errors less than or equal to $30 \%$ to be "accurate". This level of error is probably realistic relative to evaluating the Kodiak and Chignik escapement data. Return per spawner analyses for both areas did not account for this error and probably led to the narrow confidence intervals for the parameter estimates. Significant parameter estimates from Kodiak, and non-significant estimates from Chignik could be explained by larger measurement error within the Chignik data than for Kodiak. Had we been able to quantify and correct the escapement estimate bias for both Kodiak and Chignik escapement data, a more concrete return per spawner relationship for both areas would have resulted.

Overall, no conclusive evidence of reduced production of pink salmon adults from the 1989 escapement was found in our study.

## CONCLUSIONS

It is apparent that both egg retention and preemergent fry analyses demonstrated impact caused by spawner densities experienced within some streams during 1989. Evaluation of the overall 1989 brood year success using spawner-recruit analyses also demonstrated a density dependent response to the 1989 escapement. However, the variability in the Kodiak pink salmon return per spawner data made numerical estimation of the damage caused by this escapement event tenuous. The major question remaining unanswered is whether the 1989 odd-year escapement event has caused depressed production from the even-year brood line; since 1989, even-year brood line production has been well below forecast. This forecast is one of the most accurate for any salmon species in Alaska. We feel after evaluating even-year preemergent fry indices and escapement numbers that the causal mechanism for this depressed production is marine derived.

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APPENDIX

Appendix A.1. Pink salmon peak and estimated total escapement counts by fishing district for streams assessed using aerial and foot surveys within the Kodiak Management Area, 1989.

|  | Numbers of Fish |  |
| :--- | :---: | :---: |
|  | Estimated Total <br> Escapement <br> (ETE) | ETE/PC |

Afognak District

| 251-101 | 0 | 0 |  |
| :---: | :---: | :---: | :---: |
| 251-102 | 409 | 778 | 1.90 |
| 251-103 | 4,000 | 7,604 | 1.90 |
| 251-104 | 300 | 570 | 1. 90 |
| 251-105 | 1,200 | 4, 022 | 3.35 |
| 251-109 | 500 | 951 | 1. 90 |
| 251-201 | 774 | 1,471 | 1. 90 |
| 251-202 | 651 | 1,238 | 1.90 |
| 251-206 | 9 | 17 | 1. 90 |
| 251-207 | 200 | 380 | 1.90 |
| 251-301 | 0 | 0 |  |
| 251-403 | 200 | 1,351 | 6.76 |
| 251-405 | 75 | 143 | 1.90 |
| 251-406 | 27,500 | 27,500 | 1.00 |
| 251-407 | 8,425 | 8,425 | 1.00 |
| 251-502 | 225 | 225 | 1.00 |
| 25ı-503 | 0 | 0 |  |
| 251-504 | 615 | 615 | 1.00 |
| 251-505 | 325 | 325 | 1.00 |
| 251-505A | 424 | 424 | 1. 00 |
| 251-506 | 274 | 274 | 1.00 |
| 251-507 | 287 | 287 | 1.00 |
| 251-510 | 2,000 | 2,000 | 1.00 |
| 251-811 | 41 | 41 | 1.00 |
| 251-812 | 2 | 2 | 1.00 |
| 251-813 | 2,150 | 2,150 | 1.00 |
| 251-821 | 0 | 0 |  |
| 251-826 | 0 | 0 |  |
| 251-829 | 0 | 0 |  |
| 251-832 | 75 | 143 | 1.90 |
| 251-901 | 1,500 | 6,000 | 4.00 |
| 252-103 | 1,150 | 2,186 | 1. 90 |
| 252-104 | 325 | 618 | 1.90 |
| 252-301 | 1,500 | 1,500 | 1.00 |
| 252-302 | 50,000 | 95,052 | 1.90 |
| 252-303 | 0 | 0 |  |

-Continued-

Appendix A.1. (page 2 of 8 )

| Stream No. | Numbers of Fish |  |  |
| :---: | :---: | :---: | :---: |
|  | $\begin{gathered} \text { Peak Count } \\ (\mathrm{PC}) \end{gathered}$ | Estimated Total Escapement (ETE) | ETE/PC ${ }^{\text {a }}$ |
| 252-305 | 3,000 | 3,000 | 1.00 |
| 252-306 | 150,000 | 285,157 | 1.90 |
| 252-307 | 60,000 | 114,063 | 1.90 |
| 252-308 | 3 | 6 | 1.90 |
| 252-309 | 60,000 | 114,063 | 1.90 |
| 252-317 | 1 | 8 | 8.00 |
| 252-318 | 150,000 | 285,157 | 1.90 |
| 252-319 | 0 | 0 |  |
| 252-323 | 150,000 | 285,157 | 1.90 |
| 252-331 | 2,356 | 2,356 | 1.00 |
| 252-331A | 1,429 | 2,717 | 1.90 |
| 252-331B | 1,156 | 2,198 | 1.90 |
| 252-332 | 30,000 | 39,207 | 1.31 |
| 252-333 | 15,000 | 28,516 | 1.90 |
| 252-335 | 189 | 359 | 1. 90 |
| 252-337 | 3 | 6 | 2.00 |
| 252-338 | 2 | 4 | 1.90 |
| 252-339 | 7 | 13 | 1.90 |
| 252-343 | 10,000 | 20,794 | 2.08 |
| Subtotal | 738,282 | I, 349,069 | 1.83 |

## Northwest Kodiak District

| $253-114$ | 2,500 | 3,707 | 1.48 |
| :--- | ---: | ---: | ---: |
| $253-115$ | 84,000 | 193,167 | 2.30 |
| $253-121$ | 8,000 | 13,600 | 1.70 |
| $253-122$ | 520,000 | 795,899 | 1.53 |
| $253-141$ | 1,800 | 3,422 | 1.90 |
| $253-142$ | 900 | 4,400 | 4.89 |
| $253-311$ | 1,500 | 2,852 | 1.90 |
| $253-313$ | 12,000 | 12,000 | 1.00 |
| $253-321$ | 6,000 | 11,406 | 1.90 |
| $253-331$ | 367,000 | 410,913 | 1.12 |
| $253-332$ | 325,000 | 353,333 | 1.09 |
| $253-333$ | 8,800 | 2,740 | 3.43 |
| $254-103$ | 1,400 | 3,802 | 1.90 |
| $254-105$ | 60,000 | 1,400 | 1.00 |
| $254-201$ | 300,000 | 60,000 | 1.00 |
| $254-202$ | 165,000 | 251,957 | 2.15 |
| $254-203$ | 34,000 | 63,820 | 1.53 |
| $254-204$ | 500 | 1,500 | 1.88 |
| $254-205$ | 1,000 | 1,000 | 3.00 |
| $254-206$ | 1,000 | 2,200 | 1.00 |
| $254-207$ | 1,000 | 2,000 | 2.20 |
| $254-208$ | 1,500 | 1,500 | 2.00 |
| $254-209$ | 1,500 | 1,500 | 1.00 |
| $254-210$ | 5,000 | 5,000 | 1.00 |
| $254-211$ | 2,500 | 2,500 | 1.00 |
| $254-212$ |  |  |  |

-Continued-

Appendix A.1. (page 3 of 8 )

| Stream No. | Numbers of Fish |  |  |
| :---: | :---: | :---: | :---: |
|  | Peak Count (PC) | Estimated Total Escapement (ETE) | ETE/PC ${ }^{\text {a }}$ |
| 254-301 | 95,000 | 103,983 | 1.09 |
| 254-302 | 3,000 | 3,000 | 1.00 |
| 254-401 | 48,000 | 91,250 | 1.90 |
| 254-402 | 2,500 | 4,753 | 1.90 |
| 254-403 | 100 | 190 | 1.90 |
| 254-404 | 1,500 | 4,833 | 3.22 |
| 254-405 | 400 | 1,920 | 4.80 |
| 259-363 | 0 | 0 |  |
| 259-365 | 236,000 | 432,100 | 1.83 |
| 259-366 | 8,700 | 16,539 | 1.90 |
| 259-368 | 7,000 | 13,307 | 1.90 |
| 259-371 | 39,000 | 219,943 | 3.08 |
| 259-372 | 4,500 | 14,700 | 3.27 |
| 259-373 | 1,800 | 1,800 | 1.00 |
| 259-381 | 0 | 0 |  |
| 259-382 | 54,000 | 126,000 | 2.33 |
| 259-383 | 0 | 0 |  |
| 259-391 | 5,400 | 5,477 | 1.01 |
| 259-392 | 4,000 | 7,604 | 1.90 |
| 259-394 | 3,200 | 3,200 | 1.00 |
| 259-395 | 286 | 544 | 1.90 |
| 259-397 | 300 | 1,500 | 5.00 |
| 259-398A | 36 | 68 | 1.90 |
| 259-398B | 5 | 10 | 1.90 |
| 259-399 | 3 | 6 | 1.90 |
| Subtotal | 2,420,630 | 3,803,498 | 1.57 |
| Southwest Kodiak District |  |  |  |
| 256-101 | 0 | 0 |  |
| 256-102 | 0 | 0 |  |
| 256-202 | 0 | 0 |  |
| 256-301 | 1,500 | 2,852 | 1.90 |
| 256-302 | 0 | 0 |  |
| 256-303 | 0 | 0 |  |
| 256-303A | 0 | 0 |  |
| 256-401 | 1,250 | 1,692 | 1.35 |
| 256-402 | 3,000 | 5,800 | 1.93 |
| Subtotal | 5,750 | 10,344 | 1.80 |

Alitak Bay District

| $257-101$ | 0 | 0 |  |
| ---: | ---: | ---: | ---: |
| $257-102$ | 1,000 | 1,000 | 1.00 |
| $257-301$ | 0 | 0 |  |
| $257-303$ | 600 | 1,400 | 2.33 |
| $257-401$ | 550 | 2,097 | 3.81 |
| $257-402$ | 300 | 570 | 1.90 |

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Appendix A.1. (page 4 of 8 )

| Stream No. | Numbers of Fish |  |  |
| :---: | :---: | :---: | :---: |
|  | Peak Count (PC) | Estimated Total Escapement (ETE) | ETE/PC ${ }^{\text {a }}$ |
| 257-405 | 6,000 | 6,000 | 1.00 |
| 257-502 | 410,000 | 1,093,543 | 2.67 |
| 257-503 | 87,000 | 139,417 | 1.60 |
| 257-504 | 32,000 | 67,133 | 2.10 |
| 257-601 | 12,000 | 24,800 | 2.07 |
| 257-602 | 35,000 | 72,333 | 2.07 |
| 257-603 | 55,000 | 62,962 | 1.14 |
| 257-604 | 123,000 | 212,733 | 1.73 |
| 257-605 | 7,000 | 14,260 | 2.04 |
| 257-701 | 690,000 | 1,783,196 | 2.58 |
| 257-702 | 94,000 | 103,400 | 1.10 |
| 257-702A | 65,000 | 134,333 | 2.07 |
| 257-703 | 0 | 0 |  |
| Subtotal | 1,618,450 | 3,719,177 | 2.30 |

## Eastside Kodiak District

| $258-101$ | 2,600 | 12,880 | 4.95 |
| :--- | ---: | ---: | ---: |
| $258-201$ | 4,700 | 28,420 | 6.05 |
| $258-202$ | 200 | 271 | 1.36 |
| $258-203$ | 150 | 150 | 1.00 |
| $258-204$ | 200 | 227 | 1.14 |
| $258-205$ | 300 | 340 | 1.13 |
| $258-206$ | 4,100 | 7,653 | 1.87 |
| $258-207$ | 60,500 | 69,762 | 1.15 |
| $258-208$ | 900 | 1,647 | 1.83 |
| $258-209$ | 1,600 | 1,600 | 1.00 |
| $258-210$ | 0 | 0 |  |
| $258-211$ | 400 | 647 | 1.62 |
| $258-212$ | 4,000 | 10,800 | 2.70 |
| $258-213$ | 28,500 | 43,700 | 1.53 |
| $258-301$ | 2,800 | 2,800 | 1.00 |
| $258-302$ | 3,500 | 3,500 | 1.00 |
| $258-303$ | 2,000 | 6 | 3.00 |
| $258-304$ | 575 | 10,247 | 3.42 |
| $258-305$ | 150 | 1,648 | 2.87 |
| $258-306$ | 1,100 | 430 | 2.87 |
| $258-307$ | 0 | 3,520 | 3.20 |
| $258-401$ | 0 | 0 |  |
| $258-511$ | 11,000 | 0 | 0 |
| $258-512$ | 0,80 | 11,000 |  |
| $258-513$ | 10,800 | 0 |  |
| $258-514$ | 15,000 | 35,800 |  |
| $258-521$ | 53,300 | 0 | 147,423 |

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Appendix A.1. (page 5 of 8 )

| Stream No. | Numbers of Fish |  | $E T E / P C^{\text {a }}$ |
| :---: | :---: | :---: | :---: |
|  | Peak Count (PC) | Estimated Total Escapement (ETE) |  |
| 258-541A | 0 | 0 |  |
| 258-542 | 78,000 | 118,323 | 1.52 |
| 258-544 | 17,000 | 19,257 | 1.13 |
| 258-551 | 2,600 | 10,450 | 4.02 |
| 258-544 | 1,000 | 1,533 | 1.53 |
| 258-555 | 300 | 620 | 2.07 |
| 258-601 | 4,000 | 8,267 | 2.07 |
| 258-602 | 50,400 | 81,256 | 1.61 |
| 258-603 | 42,000 | 81,200 | 1.93 |
| 258-701 | 450,000 | 1,192,683 | 2.65 |
| 258-702 | 68,000 | 146,467 | 2.15 |
| 258-703 | 47,000 | 65,033 | 1.38 |
| 258-704 | 0 | 0 |  |
| 258-705 | 19,000 | 25,367 | 1.34 |
| 258-706 | 0 | 0 |  |
| 258-707 | 100 | 190 | 1.90 |
| 258-801 | 0 | 0 |  |
| 258-803 | 0 | 0 |  |
| 258-804 | 0 | 0 |  |
| 258-805 | 0 | 0 |  |
| 258-807 | 0 | 0 |  |
| 258-808 | 100 | 200 | 2.00 |
| 258-809 | 0 | 0 |  |
| 258-810 | 0 | 0 |  |
| 258-851 | 0 | 0 |  |
| 258-852 | 100 | 100 | 1.00 |
| 258-901 | 1,800 | 3,720 | 2.07 |
| 257-902 | 17,000 | 35,133 | 2.07 |
| 258-903 | 3,000 | 3,000 | 1.00 |
| 258-904 | 2,000 | 2,000 | 1.00 |
| 259-401 | 3,600 | 6,844 | 1.90 |
| 259-403 | 100 | 190 | 1.90 |
| 259-411 | 2,000 | 3,802 | 1.90 |
| 259-412 | 61,000 | 196,087 | 3.21 |
| 259-414 | 96,000 | 142,535 | 1.48 |
| 259-415 ${ }^{\text {b }}$ | 81,000 | 251,741 | 3.11 |
| 259-416 | 6,500 | 12,357 | 1.90 |
| 259-417 | 500 | 1,300 | 2.60 |
| 259-417A | 0 | 0 |  |
| 259-418 | 0 | 0 |  |
| 259-418A | 900 | 1,711 | 1.90 |
| 259-418B | 0 | 0 |  |
| 259-419 | 0 | 0 |  |
| 259-420 | 0 | 0 |  |
| 259-421 | 0 | 0 |  |
| 259-422 | 20,200 | 47,160 | 2.33 |
| 259-423 | 700 | 1,820 | 2.60 |
| 259-424 | 46,000 | 103,478 | 2.25 |
| 259-425 | 6,300 | 16,647 | 2.64 |
| 259-426 | 800 | 1,521 | 1.90 |
| 259-427 | 200 | 200 | 1.00 |
| 259-428 | 0 | 0 |  |
| Subtotal | 1,369,077 | 3,076,233 | 2.25 |

-Continued-

Appendix A.1. (page 6 of 8 )

|  | Numbers of Fish |  |  |  |
| :--- | :---: | :---: | :---: | :---: |
|  | Peak Count <br> Streamated Total <br> (PC) | Escapement <br> (ETE) |  |  |

Northeast Kodiak District

| $259-101$ | 8,000 | 11,519 | 1.44 |
| ---: | ---: | ---: | ---: |
| $259-102$ | 42,100 | 42,950 | 1.02 |
| $259-105$ | 4,000 | 7,298 | 1.82 |
| $259-221$ | 22,000 | 41,823 | 1.90 |
| $259-222$ | 36,500 | 108,498 | 2.97 |
| $259-223$ | 113,000 | 329,627 | 2.92 |
| $259-231$ | 152,000 | 365,383 | 2.40 |
| $259-233$ | 500 | 951 | 1.90 |
| $259-235$ | 3,700 | 5,427 | 1.47 |
| $259-242$ | 46,000 | 139,668 | 3.04 |
| $259-243$ | 47,700 | 90,680 | 1.90 |
| $259-244$ | 12,600 | 12,600 | 1.00 |
| $259-245$ | 10,200 | 19,391 | 1.90 |
| $259-246$ | 300 | 300 | 1.00 |
| $259-251$ | 39,400 | 74,901 | 1.90 |

Subtotal
$538,000 \quad 2.33$

Mainland District

| 262-101 | 100 | 190 | 1.90 |
| :---: | :---: | :---: | :---: |
| 262-102 | 12,000 | 21,693 | 1.81 |
| 262-103 | 600 | 1,141 | 1.90 |
| 262-104 | 0 | 0 |  |
| 262-105 | 0 | 0 |  |
| 262-106 | 0 | 0 |  |
| 262-107 | 0 | 0 |  |
| 262-108 | 3,600 | 6,480 | 1.80 |
| 262-109 | 0 | 0 |  |
| 262-110 | 0 | 0 |  |
| 262-111 | 0 | 0 |  |
| 262-113 | 3,800 | 7,224 | 1.90 |
| 262-151 | 0 | 0 |  |
| 262-152 | 15,000 | 28,516 | 1.90 |
| 262-153 | 48,000 | 91,250 | 1.90 |
| 262-154 | 0 | 0 |  |
| 262-155 | 800 | 1,521 | 1.90 |
| 262-201 | 3,000 | 12,000 | 4.00 |
| 262-202 | 0 | 0 |  |
| 262-203 | 0 | 0 |  |
| 262-204 | 0 | 0 |  |
| 262-205 | 42,000 | 79,844 | 1.90 |
| 262-206 | 6,000 | 11,406 | 1.90 |
| 262-207 | 0 | 0 |  |
| 262-254 | 100 | 190 | 1.90 |
| 262-271 | 2,000 | 3,802 | 1.90 |
| 262-272 | 0 | 0 |  |
| 262-273 | 0 | 0 |  |
| 262-301 | 0 | 0 |  |
| 262-351 | 9,000 | 17,109 | 1.90 |
| 262-352 | 3,800 | 7,224 | 1.90 |

-Continued-

Appendix A.1. (page 7 of 8 )

| Stream No. | Numbers of Fish |  |  |
| :---: | :---: | :---: | :---: |
|  | Peak Count (PC) | Estimated Total Escapement (ETE) | ETE/PC ${ }^{\text {a }}$ |
| 262-401 | 3,800 | 7,224 | 1.90 |
| 262-402 | 19,000 | 36,120 | 1.90 |
| 262-451 | 37,000 | 54,327 | 1.47 |
| 262-501 | 200,000 | 380,209 | 1.90 |
| 262-502 | 0 | 0 |  |
| 262-503 | 1,000 | 1,901 | 1.90 |
| 262-504 | 10,000 | 10,000 | 1.00 |
| 262-505 | 10,000 | 10,000 | 1.00 |
| 262-551 | 38,000 | 65,910 | 1.73 |
| 262-552 | 0 | 0 |  |
| 262-601 | 0 | 0 |  |
| 262-602 | 0 | 0 |  |
| 262-603 | 0 | 0 |  |
| 262-604 | 80,000 | 282,600 | 3.53 |
| 262-605 | 10,000 | 22,667 | 2.27 |
| 262-606 | 20 | 38 | 1.90 |
| 262-651 | 445,000 | 541,827 | 1.22 |
| 262-652 | 162,000 | 201,400 | 1.24 |
| 262-653 | 230,000 | 230,000 | 1.00 |
| 262-654 | 0 | 0 |  |
| 262-655 | 300 | 620 | 2.07 |
| 262-656 | 13,600 | 33,320 | 2.45 |
| 262-701 | 44,000 | 92,000 | 2.09 |
| 262-702 | 67,000 | 138,880 | 2.07 |
| 262-703 | 42,000 | 42,000 | 1.00 |
| 262-704 | 20,000 | 27,253 | 1.36 |
| 262-705 | 12,000 | 19,867 | 1. 66 |
| 262-706 | 0 | 0 |  |
| 262-751 | 73,000 | 115,612 | 1.58 |
| 262-752 | 10,000 | 11,560 | 1.16 |
| 262-753 | 3,500 | 6,654 | 1.90 |
| 262-801 | 22,000 | 23,303 | 1.06 |
| 262-802 | 76,000 | 158,350 | 2.08 |
| 262-803 | 500 | 500 | 1.00 |
| 262-804 | 200 | 200 | 1.00 |
| 262-850 | 5,000 | 5,000 | 1.00 |
| 262-851 | 1,320,000 | 2,078,433 | 1.57 |
| 262-852 | 91,000 | 110,580 | 1.22 |
| 262-853 | 66,000 | 124,200 | 1.88 |
| 262-854 | 10,000 | 12,000 | 1.20 |
| 262-856 | 100 | 190 | 1.90 |
| 262-858 | 1,900 | 2,489 | 1.31 |
| 262-859 | 0 | 0 |  |
| 262-860 | 0 | 0 |  |
| 262-861 | 9,000 | 18,600 | 2.07 |
| 262-862 | 200 | 200 | 1.00 |
| 262-863 | 0 | 0 |  |
| 262-864 | 0 | 0 |  |
| 262-865 | 50 | 95 | 1.90 |
| 262-866 | 800 | 1,521 | 1.90 |
| 262-868 | 100 | 190 | 1.90 |

-Continued-

Appendix A.1. (page 8 of 8 )

| Stream No. | Numbers of Fish |  |  |
| :---: | :---: | :---: | :---: |
|  | Peak Count (PC) | Estimated Total Escapement (ETE) | ETE/PC ${ }^{\text {a }}$ |
| 262-951 | 24,000 | 45,625 | 1.90 |
| 262-952 | 28,000 | 53,229 | 1.90 |
| Subtotal | 3,335,870 | 5,256,284 | 1.58 |
| Total | 10,026,059 | 18,465,619 | 1.84 |

${ }^{\text {a }}$ Estimated total escapements derived by the AUC method included fractions of whole numbers which the ETE/PC ratio reflects in several instances.
b Estimate includes estimates of fish spawning downstream of the counting weir.

Appendix A.2. Pink salmon aerial survey odd-year peak counts, Kodiak Management Area, 1969-1987.

| Stream No. | Peak Count |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1969 | 1971 | 1973 | 1975 | 1977 | 1979 | 1981 | 1983 | 1985 | 1987 | $\begin{array}{r} \text { Mean } \\ (1969-87) \end{array}$ | 1989 S | Index <br> Stream |
| Afognak District |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 251-1.01 | 0 | 3,200 | 200 | 0 | 0 | 5,000 | 450 | 600 | 1,200 | 7,500 | 1,815 | 0 | ) |
| 251-102 | 25 | 0 |  | 0 | 60 | 600 |  | 0 |  |  | 114 | 409 |  |
| 251-103 | 0 | 50 |  |  | 30 | 675 |  | 170 | 0 | 0 | 132 | 4,000 |  |
| 251-104 |  | 10 |  |  | 0 |  |  | 0 |  | 0 | 3 | 300 |  |
| 251-105 | 1,000 | 100 | 3,800 | 900 | 10,000 | 19,000 | 16,500 | 13,500 | 3,500 | 25,000 | 9,330 | 1,200 | * |
| 251-106 |  | 0 | 0 |  |  | 0 | 0 |  |  |  | 0 |  |  |
| 251-201. | 0 | 100 |  | 0 | 20 | 1,124 | 0 | 7,000 | 400 |  | 1,081 | 774 |  |
| 251-201 ${ }^{\text {a }}$ | 0 | 0 |  |  | 0 | 87 | 0 | 400 |  | 0 | 70 |  |  |
| 251-202 | 0 | 100 | 0 |  | 600 | 450 |  | 3,300 | 20 |  | 639 | 651 |  |
| 251-203 |  |  |  |  | 1,200 | 600 | 7,500 |  |  |  | 3,100 |  |  |
| 251-204 |  |  |  |  |  | 68 | 0 |  |  |  | 34 |  |  |
| 251-301 | 1,600 | 7,200 |  |  | 0 | 8,000 | 28,300 | 1,500 | 7,599 | 5,500 | 7,462 | 0 |  |
| 251-302 | 25 | 300 | 700 | 0 | 0 | 15,000 | 11,000 | 0 | 0 | 0 | 2,703 |  |  |
| 251-403 | 8,200 | 1,500 | 800 | 1,500 | 500 | 1,500 | 8,000 | 7,000 | 400 | 0 | 2,940 | 200 |  |
| 251-404 | 2,900 | 1,300 | 3,000 | 200 | 1,300 | 13,000 | 26,000 | 13,000 | 9,900 | 6,500 | 7,710 |  | * |
| 251-405 | 0 | 0 |  | 0 |  |  |  | 0 | 10 |  | 2 | 75 |  |
| 251-406 |  | 0 |  | 0 | 0 | 15 |  |  |  |  | 4 | 27,500 |  |
| 251-407 |  | 0 |  |  |  | 3,499 |  |  |  |  | 1,750 | 8,425 |  |
| 251-501 |  |  |  |  | 0 |  |  |  | 0 |  | 0 |  |  |
| 251-502 |  |  |  |  | 0 | 5 |  |  |  |  | 3 | 225 |  |
| 251-503 |  |  |  | 0 | 0 | 0 |  |  |  |  | 0 | 0 |  |
| 251-504 |  |  |  | 0 | 0 | 0 |  |  |  |  | 0 | 615 |  |
| 251-505 |  |  |  | 0 | 0 | 0 |  |  |  |  | 0 | 305 |  |
| 251-601 | 0 | 0 |  | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 |  |  |
| 251-701 | 0 | 0 |  |  | 0 |  | 0 | 0 | 0 |  | 0 |  |  |
| 251-702 | 0 | 0 |  |  | 0 | 0 | 100 | 0 | 87 |  | 27 |  |  |
| 251-703 |  |  |  |  | 0 | 0 | 0 | 0 | 0 |  | 0 |  |  |
| 251-706 |  |  |  |  |  |  |  |  | 3 |  | 3 |  |  |
| 251-821 | 0 |  |  | 500 | 15 |  | 1,800 | 2,000 | 0 | 0 | 616 | 0 |  |
| $251-822^{\text {c }}$ |  |  |  |  |  |  | 61,193 | 21,670 | 199.211 | 29,093 | 77,092 | 106,347 | 7 * |
| 251-823 |  |  |  |  | 0 |  | 0 | 0 | 0 | 0 | 0 |  |  |
| 251-824 | 0 |  |  |  | 0 |  | 0 | 0 | 0 |  | 0 |  |  |
| 251-825 |  |  |  |  |  | 0 | 138 | 0 |  |  | 46 |  |  |
| 251-826 |  |  |  | 0 | 10 |  | 73 | 0 | 0 |  | 17 | 0 | O |
| 251-827 |  |  |  |  | 0 | 12 | 452 | 800 | 5,000 | 0 | 1,044 |  |  |

-Continued-

Appendix A.2. (page 2 of 11)

| Stream No. | Peak Count |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1969 | 1971 | 1973 | 1975 | 1977 | 1979 | 1981 | 1983 | 1985 | 1987 | $\begin{array}{r} \text { Mean } \\ (1969-87) \end{array}$ | 1989 S | Index <br> Stream |
| 251-828 | 100 |  |  | 0 | 0 | 0 | 581 |  | 375 |  | 176 |  |  |
| 251-829 |  |  |  |  |  | 0 | 1,205 | 0 | 0 |  | 301 | 0 | 0 |
| 251-830 | 25,000 | 25,000 | 11,000 | 21,000 | 21,600 | 71,000 | 56,000 | 18,000 | 33,145 | 12,094 | 29,384 |  | * |
| 251-831 |  |  |  |  |  |  |  | 533 | 26,904 | 202 | 9,213 |  | - |
| 251-901 | 0 | 3,323 |  | 1,688 | 14,314 | 10,000 | 29,000 | 5,500 | 15,000 |  | 9,853 | 1,500 | 0 |
| 251-901A |  |  |  |  |  | 0 |  |  |  |  | 0 |  |  |
| 251-902 |  |  |  |  |  |  | 500 | 0 | 3,000 |  | 1,167 |  |  |
| 251-903A |  |  |  |  |  |  |  | 0 |  | 0 | 0 |  |  |
| 252-101 |  | 300 |  | 90 |  | 300 | 4,400 | 0 | 9,000 |  | 2,348 |  |  |
| 252-102 |  | 50 |  | 0 | 0 | 700 | 8,500 | 0 | 11,000 |  | 2,893 |  |  |
| 252-301 |  |  |  |  |  |  |  |  | 10,000 |  | 10,000 | 1,500 |  |
| 252-302 |  |  |  |  |  |  |  | 100 |  | 0 | 50 | 50,000 |  |
| 252-305 |  |  |  |  |  |  |  | 85 |  | 0 | 43 | 3,000 |  |
| 252-306 |  |  |  |  |  |  |  | 2,550 | 15,000 | 0 | 5,850 | 150,000 |  |
| 252-307 |  |  |  |  |  |  |  | 213 | 5,000 | 0 | 1,738 | 60,000 |  |
| 252-308 |  |  |  |  |  |  |  |  |  | 0 | 0 | 3 | 3 |
| 252-309 |  |  |  |  |  |  | 6,000 |  |  | 800 | 3,400 | 60,000 |  |
| 252-311 | 0 |  |  | 25 |  | 0 | 500 |  |  |  | 131 |  |  |
| 252-312 | 50 |  |  | 90 | 563 | 284 | 1,200 |  |  |  | 437 |  |  |
| 252-314A |  |  |  |  | 500 |  |  |  |  |  | 500 |  |  |
| 252-315 | 25 |  |  | 200 |  | 90 | 1,200 |  |  |  | 379 |  |  |
| 252-316 | 250 | 300 |  | 438 | 718 | 1.062 |  |  |  |  | 554 |  |  |
| 252-317 | 0 |  |  | 0 | 1,409 | 1,012 | 100 |  |  | 0 | 420 | 1 | 1 |
| 252-318 |  |  |  | 0 | 208 | 4 | 0 |  | 0 |  | 42 | 150,000 |  |
| 252-319 | 0 |  |  | 0 | 25 | 0 |  | 0 |  | 0 | 4 | 0 | 0 |
| 252-321 | 0 |  |  | 18 |  | 850 | 0 |  |  |  | 217 |  |  |
| 252-323 |  |  |  |  | 0 | 300 |  | 0 | 10,000 | 0 | 2,060 | 150,000 |  |
| 252-324 |  | 3.800 |  | 8,643 | 6,661 | 15,988 | 40,834 | 0 | 25,000 | 153,482 | 31,801 |  |  |
| 252-331 | 600 |  | 200 | 0 | 158 | 3,200 | 18,500 | 1,096 | 200 | 1,200 | 2,795 | 2,356 |  |
| 252-332 | 7,600 | 5,800 | 3,000 | 2,100 | 21,000 | 24,000 | 44,029 | 11,000 | 13,000 | 2,700 | 13,423 | 30,000 | 0 * |
| 252-333 | 0 |  |  | 127 |  | 512 | 0 | 0 |  | 900 | 257 | 15,000 |  |
| 252-335 |  |  | 100 | 0 | 0 |  | 0 | 0 |  | 0 | 17 | 189 |  |
| 252-341b | 0 |  |  | 0 | 3 | 799 | 1,000 | 0 |  |  | 300 |  |  |
| 252-342 ${ }^{\text {b }}$ | 48,000 | 13,600 | 38,000 | 2,800 | 84,000 | 19.020 | 17,508 | 84,956 | 8,860 | 8,780 | 32,552 | 41,611 | 1 |
| 252-343 | 12,000 | 12,500 | $0$ | 0 | 29,000 | 41,500 | 38,700 | 34,500 | 9,800 | 7,100 | 18,510 | 10,000 | 0 * |

-Continued-

Appendix A.2. (page 3 of 11)

|  | Peak Count |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Stream No. | 1969 | 1971 | 1973 | 1975 | 1977 | 1979 | 1981 | 1983 | 1985 | 1987 | $\begin{array}{r} \text { Mean } \\ (1969-87) \end{array}$ | 1989 | Index Stream |

## Northwest Kodiak District

|  | 253-114 | 100 |  | 0 |  | 0 | 4 | 250 | 50 |  |  | 67 | 2,500 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 253-115 | 0 |  | 2,000 |  | 11,500 | 36,000 | 82,000 | 32,000 | 18,300 | 41,000 | 27,850 | 84,000 | * |
|  | 253-121 | 500 | 1,500 | 1,000 | 0 | 50 | 1,600 | 6,000 | 1,600 |  | 16,000 | 3,139 | 8,000 |  |
|  | 253-122 | 50,000 | 37,000 | 53,000 | 77,500 | 44,000 | 99,000 | 75,000 | 128,000 | 114,000 | 190,000 | 86,750 | 520,000 | * |
|  | 253-123 | 0 |  |  |  | 0 |  | 0 | 0 |  |  | 0 |  |  |
|  | 253-141 |  |  |  |  |  |  |  | 50 |  |  | 50 | 1,800 |  |
|  | 253-201 |  |  |  |  |  |  |  |  |  | 0 | 0 |  |  |
|  | 253-203 |  |  |  |  |  |  |  |  |  | 0 | 0 |  |  |
|  | 253-311 |  |  |  |  |  |  | 100 | 350 |  |  | 225 | 1,500 |  |
|  | 253-313 |  |  | 0 |  | 362 | 60 | 0 | 150 | 1,200 | 0 | 253 | 12,000 |  |
|  | 253-321 | 1,500 | 2,500 | 3,000 |  | 325 | 4,800 | 11,700 | 1,400 |  | 2,600 | 3,478 | 6,000 |  |
| シ | 253-322 |  |  |  | 20 | 2 | 375 | 0 | 0 |  |  | 79 |  |  |
|  | 253-323 |  |  | 0 |  | 362 | 60 | 0 | 0 |  | 0 | 70 |  |  |
|  | 253-331 | 46,000 | 40.000 | 22,000 | 43,500 | 56,000 | 80,000 | 92,000 | 42,250 | 80,000 | 72,000 | 57,375 | 367,000 | - |
|  | 253-332 | 23,000 | 6,000 | 3,000 | 550 | 5,900 | 18,100 | 44,500 | 8,100 | 21,000 | 15,000 | 14,515 | 325,000 | $\bullet$ |
|  | 253-333 |  |  |  |  | 35 | 0 | 0 | 1,200 | 4,100 | 100 | 906 | 800 |  |
|  | 253-351 | 0 |  |  |  |  | 0 |  | 0 |  |  | 0 |  |  |
|  | 253-352 | 200 |  |  |  |  | 80 |  | 0 |  |  | 93 |  |  |
|  | 254-103 |  |  | 0 |  | 300 | 1,800 |  | 190 | 80 |  | 474 | 2,000 |  |
|  | 254-104 |  |  |  |  | 12,075 |  | 600 |  |  |  | 6,338 |  |  |
|  | 254-105 |  |  |  |  |  |  | 3,000 | 700 | 30 |  | 1,243 | 1,400 |  |
|  | 254-201 | 400 |  | 10 | 13,500 | 13,500 | 19,000 | 15,000 | 6,000 | 4,500 | 1,600 | 8,168 | 60,000 |  |
|  | 254-202 | 103,000 | 65,000 | 50,000 | 72,000 | 111,000 | 147,000 | 136,000 | 180,000 | 175,500 | 96,500 | 113,600 | 300,000 | - |
|  | 254-203 | 13,000 | 30,000 | 4,000 | 24,000 | 24,130 | 58,000 | 58,000 | 62,500 | 80,000 | 30,000 | 38,363 | 165,000 | * |
|  | 254-204 | 2,600 | 400 | 800 |  | 4,500 | 5,200 | 32,000 | 4,500 | 7,100 | 11,000 | 7,567 | 34,000 | - |
|  | 254-205 |  |  | 0 | 1,200 | 1,950 | 400 | 0 | 0 | 0 | 0 | 444 | 500 |  |
|  | 254-206 |  |  |  | 500 | 1,350 | 300 | 0 | 0 | 0 | 0 | 307 | 1,000 |  |
|  | 254-207 |  |  | 0 |  | 8,500 | 1,500 | 0 | 0 | 40 | 300 | 1,477 | 1,000 |  |
|  | 254-208 |  |  |  |  | 1,480 | 350 | 300 | 60 | 30 | 500 | 453 | 1,000 |  |
|  | 254-209 |  |  |  |  | 640 |  | 0 | 0 | 0 |  | 160 | 1,500 |  |
|  | 254-210 |  |  |  |  | 1,900 |  | 0 | 0 | 0 |  | 475 | 1,500 |  |
|  | 254-211 |  |  |  |  | 560 |  | 0 | 0 |  |  | 187 | 5,000 |  |
|  | 254-212 |  |  |  |  | 390 |  | 0 | 600 | 2,300 |  | 823 | 2,500 |  |
|  | 254-213 |  |  |  |  |  |  |  | 220 | 1,500 |  | 860 |  |  |
|  | 254-301 | 18,000 | 14,000 | 11,000 | 35,500 | 35,500 | 43,000 | 50,000 | 30,000 | 4,000 | 30,000 | 27,100 | 95,000 | - |

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| Stream No. | Peak Count |  |  |  |  |  |  |  |  |  |  |  | Index <br> Stream |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1969 | 1971 | 1973 | 1975 | 1977 | 1979 | 1981 | 1983 | 1985 | 1987 | $\begin{array}{r} \text { Mean } \\ (1969-87) \end{array}$ | 1989 |  |
| 254-302 | 0 |  | 5,000 | 0 | 3,200 | 6,000 | 300 | 0 |  | 0 | 1,813 | 3,0 |  |
| 254-401 | 2,000 | 20,000 | 0 |  | 45,000 | 18,000 | 0 | 30,000 | 0 | 0 | 12,778 | 48,0 |  |
| 254-402 |  |  |  |  |  | 400 | 100 | 40 |  |  | 180 | 2,5 |  |
| 254-403 |  |  |  |  | 400 | 150 | 50 | 20 |  |  | 155 |  |  |
| 254-404 |  |  |  |  |  |  | 2,800 | 7,100 | 750 | 450 | 2,775 | 1,5 |  |
| 254-405 |  |  |  |  |  |  | 2,300 | 100 |  |  | 1,200 |  |  |
| 255-102 |  |  | 80 |  | 7,770 | 129 |  | 0 |  |  | 1,995 |  |  |
| 259-362 |  |  |  |  | 125 |  |  |  |  |  | -125 |  |  |
| 259-363 | 0 |  | 0 | 0 | 0 | 0 | 200 | 200 | 0 | 0 | 44 |  | 0 |
| 259-364 | 1,400 |  | 50 |  | 400 | 12,000 | 4,500 | 400 | 2,500 | 2,100 | 2,919 |  |  |
| 259-365 | 8,300 | 4.000 | 8,000 | 11,000 | 19,300 | 29,600 | 58,250 | 18,000 | 35,000 | 29,000 | 22,045 | 236,0000 |  |
| 259-366 | 1,500 |  | 300 |  | 150 | 16,500 | 4,300 | 600 | 1,600 | 4,700 | 3,706 | 8,7 |  |
| 259-367 |  |  |  |  | 4 |  | 120 |  |  | 1,400 | 508 |  |  |
| 259-368 |  |  |  |  | 73 | 2,700 | 20 |  |  | 1.50 | 711 | 7,0 |  |
| 259-369 |  |  |  |  | 250 |  | 0 |  |  | 200 | 150 |  |  |
| 259-371 | 11,200 | 6,400 | 4,000 | 12,000 | 18,000 | 38,000 | 20,500 | 18,000 | 40,900 | 15,000 | 18,400 | 39,0 | 0 * |
| 259-372 |  |  |  |  |  | 60 | 0 | 200 |  | 2,000 | 18,56 | 4,5 |  |
| 259-373 |  |  |  |  |  |  | 20 |  |  |  | 20 | 1,8 |  |
| 259-381 |  |  |  |  |  | 0 | 6,000 | 0 |  | 0 | 1,500 |  | 0 |
| 259-382 | 4,700 | 2,300 | 1,800 | 2,650 | 3,500 | 16,200 | 4,000 | 2,800 | 6,600 | 3,780 | 4,833 | 54,0 |  |
| 259-383 | 1,700 | 200 | 500 | 0 | 500 | 3,800 | , 0 | 2,900 |  | +800 | 1,156 |  | 0 |
| 259-391 | 1,500 |  | 500 | 400 | 300 | 8,600 | 20,000 | 5,500 | 7,400 | 2,000 | 5,133 | 5,4 |  |
| 259-392 | 2.500 |  | 200 | 900 | 2,500 | 1,300 | 6,500 | 0 | 3,000 |  | 2,113 | 4,0 |  |
| 259-393 | 200 |  | 0 | 150 | 200 | 20 | 250 | 600 | 300 | 0 | 191 |  |  |
| 259-394 |  |  |  |  |  |  |  | 700 | 5,400 | 1,900 | 2,667 | 3,2 |  |
| 259-395 |  |  |  |  |  |  |  | 0 |  | 1,900 | 2,450 | , 2 |  |
| 259-397 |  |  |  |  |  |  |  |  | 4,500 |  | 4,500 |  |  |
| Southwest Kodiak District |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 255-101 ${ }^{\text {a }}$ | 0 |  | 57 | 210 | 52,319 | 82,973 | 51,248 | 38,902 | 41,232 | 24,222 | 32,351 | 109,8 | 0 - |
| 256-201 ${ }^{\text {a }}$ | 2,100 | 0 | 839 | 2,568 | 3,716 | 10,878 | 4,358 | 17,702 | 3,000 | 7,819 | 5,298 | 109, |  |
| 256-202 |  |  | 0 | 0 | 0 | 10 |  | 1,400 |  |  | 282 |  | 0 |
| 256-301 |  |  | 0 | 0 | 3,500 | 3,000 | 13,000 | 3,000 | 2,600 | 3.440 | 3,568 | 1,5 |  |
| 256-302 |  |  | 0 | 0 | 0 | 0 | 50 | 800 |  | 0 | 121 |  | 0 |
| 256-303 |  |  | 0 |  | 0 | 0 | 0 | 1,100 | 10,500 | 0 | 1,657 |  | 0 |
| 256-401 | 0 | $4,000$ | 0 | 0 | 2,000 | 45,000 | 0 | 1, 0 | 4,500 | 0 | 5,500 | 1,2 | 0 - |
| 256-402 |  | 0 |  | 0 | 5,000 | $400$ |  | 4,500 | 5,000 |  | 2,483 | 3,0 |  |

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Alitak Bay District

|  | 257-101 |  | 0 |  |  | 0 | 0 |  | 0 |  | 0 | 0 | 0 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 257-102 | 0 | 0 |  |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1,000 |
|  | 257-301 | 0 |  |  |  |  | 1,500 | 0 | 0 |  |  | 375 | 0 |
|  | 257-302 | 14,000 |  | 742 | 9,494 | 13,380 | 100 | 0 | 9,000 | 2,500 | 22,791 | 8,001 | 49,608 |
|  | 257-303 | 0 |  |  | 0 | 2,500 | 0 | 0 | 600 | 0 | 0 | 388 | 600 |
|  | 257-304 | 35,000 | 15,000 | 2,197 | 22,000 | 18,000 | 5,277 | 1,506 | 424 | 5,235 | 1,010 | 10,565 |  |
|  | 257-305 |  |  |  |  |  |  |  | 10 |  |  | 10 |  |
|  | 257-401 | 6,000 | 5,000 | 1,000 | 200 | 2,500 | 3,830 | 16,000 | 3,000 | 8,000 |  | 5,059 | 550 |
|  | 257-402 | 25 | 400 | 2,777 | 2,000 | 7,000 | 6,000 | 3,000 | 1,000 | 0 | 2,200 | 2,440 | 300 |
|  | 257-403 | 60,000 | 63,000 | 22,000 | 38,000 | 98,000 | 85,178 | 157,000 | 72,000 | 141,000 | 55,993 | 79,303 | 315,559 |
|  | 257-404 | 19,000 | 19,000 | 8,000 | 3,500 |  | 17,000 | 50 | 5,000 |  |  | 10,221 |  |
| $\omega$ | 257-405 | 5,000 |  |  | 4,500 | 4,000 | 2,400 | 4,000 | 4,500 |  |  | 4,067 | 6,000 |
|  | 257-406 |  |  |  |  | 300 |  | 3,500 | 700 |  |  | 1,500 |  |
|  | 257-407 |  |  |  |  | 400 |  | 0 | 24 |  |  | 141 |  |
|  | 257-409 |  |  |  |  |  |  | 0 | 0 |  |  | 0 |  |
|  | 257-501 |  |  | 0 | 10 | 100 |  |  | 0 | 800 |  | 182 |  |
|  | 257-502 | 140,000 | 100,000 | 40,000 | 76,700 | 141,000 | 97,500 | 180,000 | 148,000 | 166,500 | 130,700 | 122,040 | 410,000 |
|  | 257-503 | 15,000 | 4,000 | 500 | 340 | 8,000 | 7,000 | 13,000 | 8,200 | 9,000 | 6,100 | 7,114 | 87,000 |
|  | 257-504 |  |  | 0 | 0 |  | 3,500 |  | 2,000 | 350 |  | 1,170 | 32,000 |
|  | 257-505 |  |  |  |  |  |  |  | 0 |  |  | 0 |  |
|  | 257-601 | 1,500 | 0 | 450 | 0 | 2,800 | 4,500 | 50 | 2,000 | 40,000 | 0 | 5,130 | 12,000 |
|  | 257-602 | 1,700 | 4,000 | 400 | 900 | 3,000 | 6,000 | 3,800 | 8,700 | 35,000 | 25,000 | 8,850 | 35,000 |
|  | 257-603 | 6,000 | 1,000 | 4,000 | 200 | 3,000 | 10,500 | 2,500 | 10,100 | 48,300 | 24,500 | 11,010 | 55,000 |
|  | 257-604 | 2,000 | 400 | 1,000 | 2,400 | 4.500 | 6,000 | 800 | 15,500 | 58,000 | 0 | 9,060 | 123,000 |
|  | 257-605 | 0 | 0 | 0 | 0 | 0 |  | 0 | 1,000 | 2,000 |  | 375 | 7,000 |
|  | 257-701 | 44,000 | 122,000 | 45,000 | 71,300 | 92,000 | 306,000 | 240,000 | 142,000 | 273,000 | 208,500 | 154,380 | 690,000 |
|  | 257-702 |  |  | 500 | 4,200 | 10,500 | 900 |  | 1,600 | 8,000 | 0 | 3,671 | $94,000$ |
|  | 257-703 | 0 |  |  | 0 | 600 | 6,000 |  | 5,000 | 90 | 0 | 1,670 | $0$ |
|  | Eastside Kodiak District |  |  |  |  |  |  |  |  |  |  |  |  |
|  | 258-101 | 1,100 | 200 | 0 |  |  | 6,300 | 9,000 | 200 | 1,100 |  | 2,557 | 2,600 |
|  | 258-201 | 2,000 |  | 0 | 0 |  | 3,500 | 2,000 | 800 | 5,200 | 100 | 1,700 | 4,700 |
|  | 258-202 | 15 | 100 | 250 | 0 |  | 1, 000 | 0 | 20 | - 70 | 400 | 206 | 200 |
|  | 258-203 | 0 | 0 |  | 20 |  | 0 | 0 | 0 | 0 | 0 | 3 | 150 |

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| Stream No. | Peak Count |  |  |  |  |  |  |  |  |  |  |  | Index <br> Stream |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1969 | 1971 | 1973 | 1975 | 1977 | 1979 | 1981 | 1983 | 1985 | 1987 | Mean $(1969-87)$ | 1989 |  |
| 258-204 | 0 | 200 | 0 | 100 |  | 0 | 1,000 | 200 | 0 | 0 | 167 | 200 |  |
| 258-205 |  | 0 | 300 | 0 |  | 0 | 0 | 0 | 100 | 3,300 | 463 | 300 |  |
| 258-206 | 2,500 | 3,000 | 700 | 200 | 1,350 | 4.400 | 9,000 | 6,000 | 13,400 | 6,080 | 4,663 | 4,100 |  |
| 258-207 | 500 | 2,500 | 1,500 | 3,700 | 2,700 | 22,200 | 47,000 | 8,800 | 12,600 | 11,180 | 11,268 | 60,500 |  |
| 258-208 |  |  | 0 | 0 | 200 | 0 | 0 | 0 | 100 | - 0 | 38 | 900 |  |
| 258-209 | 1,000 | 50 | 500 | 0 |  | 1,200 | 3,100 | 2,500 | 1,200 | 100 | 1,072 | 1,600 |  |
| 258-210 | 0 | 0 | 0 | 0 |  | 0 | 5,000 | 0 |  | 0 | 625 | 0 | 0 |
| 258-211 | 0 | 1,000 | 10 | 0 | 400 | 0 | 0 | 0 | 50 | 0 | 146 | 400 |  |
| 258-212 | 0 |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 4,000 |  |
| 258-213 |  |  |  |  |  | 150 | 0 | 0 | 0 | 0 | 30 | 28,500 |  |
| 258-301 | 0 |  |  |  | 0 | 0 | 0 | 600 | 500 |  | 183 | 2,800 |  |
| 258-302 | 0 | 200 | 10 |  | 2.600 | 0 | 0 | 300 | 400 | 0 | 390 | 3,500 |  |
| 258-303 | 0 |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 | , 2 | 2 |
| 258-304 | 1,100 |  | 20 |  | 13,000 | 11,000 | 6,300 | 4,200 | 2,800 | 200 | 4,828 | 3,000 |  |
| 258-305 | 0 |  | 0 | 0 | 200 | 700 | 75 | 0 | 0 |  | 122 | 575 |  |
| 258-306 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  | 0 | 150 |  |
| 258-307 |  |  |  |  |  | 0 | 0 | 0 | 0 | 0 | 0 | 1,100 |  |
| 258-401 | 500 | 0 | 1,000 | 0 | 500 | 0 | 0 | 0 | 400 | 0 | 240 | 0 | 0 |
| 258-402 |  |  |  |  | 580 | 2,000 |  | 150 | 20 |  | 688 |  |  |
| 258-403 |  |  |  |  |  |  |  |  | 800 |  | 800 |  |  |
| 258-404 |  |  |  |  |  |  |  |  | 900 | 0 | 450 |  |  |
| 258-511 | 2,200 | 5,400 | 300 | 0 | 600 | 3,500 | 8,600 | 3.000 | 3,600 | 0 | 2,720 | 0 | 0 |
| 258-512 | 0 |  | 100 |  | 400 | 3,000 | 0 | 0 | 300 | 0 | 475 | 11,000 |  |
| 258-513 | 0 |  | 50 |  | 2,760 | 0 | 0 | 100 | 700 | 0 | 451 | 0 | 0 |
| 258-514 |  |  |  |  |  |  | 0 | 500 | 1,200 | 0 | 425 | 10,800 |  |
| 258-515 |  |  |  |  |  |  | 8,000 | 0 | 0 | 0 | 2,000 |  |  |
| 258-516 |  |  |  |  |  |  |  |  | 0 | 0 | 0 |  |  |
| 258-521 | 1,900 | 1,900 | 5.000 | 13,500 | 28,000 | 6,000 | 22,000 | 22,000 | 15,000 | 3,400 | 11,870 | 15,000 |  |
| 258-522 | 35,000 | 23,000 | 23,000 | 39,400 | 48,500 | 67,500 | 52,000 | 78,000 | 77,000 | 14,100 | 45,750 | 53,300 | * |
| 258-523 | 1,500 | 0 | 0 | 0 | 30 | 100 | 0 | 700 | 50 |  | 264 |  |  |
| 258-531 | 0 |  |  | 0 | 60 | 0 | 5,000 | 400 | 1,100 |  | 937 | 0 | 0 |
| 258-532 | 0 |  |  | 0 | 0 | 80 | 0 | 200 | 25 |  | 44 | 0 | 0 |
| 258-533 | 0 |  |  |  | 0 | 0 | 0 | 25 |  | 0 | 4 | 0 | 0 |
| 258-541 | 4,000 | 3,000 | 150 | 250 | 7,000 | 13,000 | 11,000 | 6,000 | 3,200 | 500 | 4,810 | 30,500 |  |
| 258-542 | 25,000 | 1,500 | 1,100 | 1,600 | 13,000 | 17,000 | 22,000 | 23,100 | 21,000 | 11,000 | 13,630 | 78,000 | * |
| 258-543 | 0 | 0 |  | 0 |  | 0 | 4,000 | 0 | 10 |  | 573 |  |  |
| 258-544 |  |  |  |  |  | 0 | 1,200 | 500 | 3,800 | 2,200 | 1,540 | 17.000 |  |

-Continued-

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| Stream No. | Peak Count |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1969 | 1971 | 1973 | 1975 | 1977 | 1979 | 1981 | 1983 | 1985 | 1987 | $\begin{array}{r} \text { Mean } \\ (1969-87) \end{array}$ | 1989 | Index <br> Stream |
| 258-551 | 7,000 | 0 | 3,000 | 0 | 400 | 500 | 34,000 | 2,000 | 1,800 |  | 5,411 | 2,600 |  |
| 258-552 | 3,800 |  | 3,000 |  | 14,000 | 11,500 | 17,000 | 4,000 | 0 | 5,360 | 7,333 |  |  |
| 258-553 | 8,000 | 0 | 200 |  | 15,000 | 9,500 | 16,000 | 6,000 |  | 850 | 6,944 |  |  |
| 258-554 | 100 |  | 0 | 0 | 600 | 300 | 2,500 | 500 |  |  | 571 | 1,000 |  |
| 258-555 | 25 |  | 0 |  | 50 | 0 | 0 | 200 |  |  | 46 | 300 |  |
| 258-601 | 400 |  |  | 0 | 200 | 200 | 0 | 40 | 3,000 |  | 549 | 4,000 |  |
| 258-602 | 22,000 | 10,000 | 1,500 | 2,500 | 9,000 | 33,000 | 2,500 | 30,800 | 10,000 | 8,800 | 13,010 | 50,400 |  |
| 258-603 | 5,200 | 0 | 13,600 | 0 | 10,000 | 9,000 | 0 | 10,000 | 12,000 | 480 | 6,028 | 42,000 |  |
| 258-701 | 39,000 | 54,000 | 24,000 | 86,000 | 53,000 | 100,700 | 128,000 | 86,000 | 60,000 | 164,400 | 79,570 | 450,000 | - |
| 258-702 | 8,000 |  | 0 | 0 | 9,000 | 8,000 | 10,600 | 4,000 | 5,200 | 8,580 | 5,931 | 68,000 |  |
| 258-703 |  |  |  |  | 6,000 | 10,000 | 5,600 | 18,000 | 400 |  | 8,000 | 47,000 |  |
| 258-704 |  |  |  |  | 0 |  |  | 800 | 0 |  | 267 | 0 |  |
| 258-705 |  |  |  |  |  |  | 0 |  | 6 |  | 3 | 19,000 |  |
| 258-706 |  |  |  |  |  |  | 10,000 |  |  |  | 10,000 | 19,000 | 0 |
| 258-801 | 17,000 |  | 600 | 0 | 4,500 | 1. 200 | $800$ | 0 |  |  | $3,443$ | 0 | 0 |
| 258-801A |  |  |  | 20 |  | 0 | 0 |  |  |  | $7$ |  |  |
| 258-801B |  |  |  |  |  |  | 3,400 |  |  |  | 3,400 |  |  |
| 258-802 |  |  |  | 0 | 20 | 0 |  |  | 200 |  | , 55 |  |  |
| 258-803 |  |  |  | 0 | 0 | 2 |  | 0 | 0 |  | 0 | 0 | 0 |
| 258-803A |  | 0 |  | 0 | 0 | 15 |  |  |  |  | 4 | 0 | 0 |
| 258-804 |  | 0 |  | 0 | 90 | 0 |  | 0 | 0 |  | 15 | 0 | 0 |
| 258-804A |  | 0 |  | 0 | 0 | 0 |  |  |  |  | 0 |  |  |
| 258-805 |  | 0 |  | 0 | 0 | 0 |  | 0 | 0 |  | 0 | 0 | 0 |
| 258-806 |  |  |  | 0 | 0 | 0 |  | 0 | 0 |  | 0 |  |  |
| 258-807 |  |  |  |  |  |  |  |  | 13 |  | 13 | 0 | 0 |
| 258-808 |  |  |  | 0 | 100 | 100 | 0 | 450 |  |  | 130 | 100 |  |
| 258-809 |  |  |  | 0 |  | 20 |  | 200 |  |  | 73 | 0 | 0 |
| 258-810 |  |  |  | 0 |  | 700 |  |  |  |  | 350 | 0 | 0 |
| 258-811 |  |  |  | 0 |  |  |  |  |  |  | 0 |  |  |
| 258-851 |  |  |  |  |  |  |  | 1,500 | 500 |  | 1,000 | 0 | 0 |
| 258-852 |  |  |  |  |  |  |  |  | 40 |  | 40 | 100 |  |
| 258-853 |  |  |  |  |  |  |  |  | 0 | 0 | 0 |  |  |
| 258-901 |  |  |  |  |  |  |  | 200 | 1,000 |  | 600 | 1,000 |  |
| 258-902 |  |  |  |  |  |  |  | 6,000 | 8,000 |  | 7.000 | 17,000 |  |
| 258-903 |  |  |  |  |  |  |  | 150 | 3,000 | 0 | 1,050 | 3,000 |  |
| 258-904 |  |  |  |  |  |  |  | 0 | $0$ |  | $0$ | $2,000$ |  |
| 259-401 | 20,000 | 21,000 | 15,000 | 14,000 | 0 | 23,000 | 3,800 | 5,400 | 11,800 | 150 | 11,415 | 3,600 |  |

-Continued-

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| Stream No. | Peak Count |  |  |  |  |  |  |  |  |  |  |  | Index <br> Stream |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1969 | 1971 | 1973 | 1975 | 1977 | 1979 | 1981 | 1983 | 1985 | 1987 | $\begin{array}{r} \text { Mean } \\ (1969-87) \end{array}$ | 1989 S |  |
| 259-402 | 10 | 0 | 0 | 0 |  | 0 |  | 120 | 0 | 0 | 16 |  |  |
| 259-403 | 0 | 60 | 100 | 100 |  | 0 |  | 20 | 2,400 | 0 | 335 | 100 |  |
| 259-404 | 0 | 130 | 0 | 50 |  | 200 |  | 0 | 10 | 0 | 49 |  |  |
| 259-405 | 0 |  | 0 |  |  | 0 |  | 0 | 0 | 0 | 0 |  |  |
| 259-410 | 2,100 | 7,000 | 2,900 | 12,000 | 100 | 14,000 | 4,100 | 3.800 | 6,600 | 12,000 | 6,460 |  |  |
| 259-411 | 10,200 | 1,500 | 0 | 260 | 100 | 5,000 | 2,000 | 400 | 14,500 | 2,000 | 3,596 | 2,000 |  |
| 259-412 | 4,000 | 3,000 | 1,100 | 12,000 | 11,900 | 71,000 | 15,380 | 17,500 | 40,000 | 22,500 | 19,838 | 61,000 | 0 |
| 259-413 | 0 | 0 |  | 300 |  | 0 |  |  | 10 |  | 62 |  |  |
| 259-414 | 20,000 | 30,000 | 8,300 | 7,700 | 33,000 | 17,000 | 6,050 | 3,500 | 1,500 | 11,100 | 13,815 | 96,000 | 0 |
| 259-414A |  |  |  |  |  |  |  |  | 1,50 | 11,100 | 13, 50 | 96,000 |  |
| 259-415 ${ }^{\text {d }}$ | 85,000 | 57,000 | 25,000 | 46,000 | 144,000 | 68,000 | 57,000 | 39,000 | 28,400 | 39,687 | 58,909 | 81,000 | 0 |
| 259-416 | 10,000 | 3.000 | 1,000 | 6,600 |  | 8,600 | 3,000 | 7,500 | 7,700 | - 0 | 5,267 | 6,500 |  |
| 259-417 |  |  | 0 | 20 |  | 10 | 0 | 500 | 400 | 0 | 133 | 500 |  |
| 259-418A | 100 | 0 | 100 | 0 | 0 | 0 | 1,000 | 3,000 | 9,000 | 0 | 1,320 | 900 |  |
| 259-418B | 0 | 0 | 0 | 0 | 0 | 0 | 1,100 | 1,000 | 20 | 0 | 212 | 0 | 0 |
| 259-419 | 0 |  | 0 | 0 | 0 | 0 | 0 | 0 | 80 | 200 | 31 | 0 | 0 |
| 259-419A |  |  |  |  |  | 445 | 0 | 0 | 250 | 0 | 139 |  |  |
| 259-420 |  |  |  | 0 |  |  | 0 |  |  |  | 0 | 0 | 0 |
| 259-421 | 0 |  |  | 0 |  | 2,400 | 0 | 10 | 30 |  | 349 | 0 | 0 |
| 259-422 | 5,000 | 5,000 | 2,000 | 4,000 |  | 7,800 | 5,500 | 9.000 | 7. 200 | 630 | 5,126 | 20,200 |  |
| 259-423 | 2,100 | 2,000 | 0 | 6,000 |  | 3,000 | 500 | 4,000 | 2,000 | 1,700 | 2,367 | 700 |  |
| 259-424 | 3,600 | 1,800 | 3,600 | 5,400 | $22,000$ | 57,300 | 14,400 | 6,400 | 4,550 | 11,000 | $13,005$ | 46,000 |  |
| 259-424A |  |  |  |  | $6,200$ |  | 14,400 |  |  |  | $6,200$ | , |  |
| 259-425 | 600 | 0 | 0 |  |  | 800 | 0 | 300 |  |  | 283 | 6,300 |  |
| 259-426 | 0 | 0 | 100 | 0 |  | 500 | 0 | 0 |  | 75 | 84 | 800 |  |
| 259-427 |  |  | 0 |  |  | 100 | 200 | 0 | 0 | 0 | 50 | 200 |  |
| 259-428 | 0 | 550 | 3,500 |  | 0 | 0 | 0 | 0 | 0 |  | 506 | 0 | 0 |
| Northeast | Kodiak Di | rict |  |  |  |  |  |  |  |  |  |  |  |
| 259-101 | 3,500 | 0 | 50 | 600 | 3,800 | 3,300 | 1,300 | 1.250 | $8,700$ | 295 | $2,280$ | 8,000 |  |
| 259-101A | 300 |  | 500 |  | 200 | 400 | 2,500 | 650 | 2,800 |  | 1,050 |  |  |
| 259-102 | 500 | 500 | 1,000 | 4,500 | 4,800 | 850 | 400 | 430 | 5,040 | 300 | 1,832 | 42,000 |  |
| 259-103 |  |  |  |  |  |  |  |  |  | 0 | 1.0 | , |  |
| 259-211 | 66,500 | 7,900 | 26,350 | 22,000 | 54,000 | 61,000 | 88,000 | 53,000 | 166,688 | 27,892 | 57,333 |  | * |
| 259-221 | 3,000 | 2,000 | 1,000 |  | . 300 | 1,400 | 1,400 | 300 | 4,000 | , 300 | 1.522 | 22,000 |  |
| 259-222 | 16,340 | 3,100 | 2,250 | 6,000 | 12,000 | 24,200 | 5,600 | 2,000 | 10,400 | 18,200 | 10,009 | 36,500 |  |

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| Stream No. | Peak Count |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1969 | 1971 | 1973 | 1975 | 1977 | 1979 | 1981 | 1983 | 1985 | 1987 | $\begin{array}{r} \text { Mean } \\ (1969-87) \end{array}$ | 1989 | Index <br> Stream |
| 259-223 | 15,100 | 4,400 | 1,500 | 11,250 | 6,000 | 19.550 | 10,000 | 5,500 | 20,400 | 10,000 | 10,370 | 113,00 |  |
| 259-224 |  |  |  |  |  |  |  | 0 | 3,000 |  | 1,500 |  |  |
| 259-231 | 52,000 | 18,900 | 11,300 | 0 | 51,900 | 45,300 | 54,000 | 71,000 | 141,000 | 112,800 | 55,820 | 152,0 | 0 |
| 259-232 |  |  |  | 0 |  |  |  | 0 |  |  | 0 |  |  |
| 259-233 |  | 1,000 |  | 20 | 200 | 0 | 0 | 0 | 0 | 1,000 | 278 |  |  |
| 259-234 |  |  |  |  |  |  | 5,500 |  |  |  | 5,500 |  |  |
| 259-235 |  |  |  |  |  |  |  |  | 300 | 200 | 250 | 3.7 |  |
| 259-242 | 37,000 | 40,000 | 13,000 | 16,000 | 41,000 | 80,000 | 73,000 | 30,020 | 69,400 | 66,100 | 46,552 | 46,0 | 0 |
| 259-243 | 5,000 | 300 |  | 400 |  | 1,100 | 0 | 0 |  | 0 | 971 | 47.7 |  |
| 259-244 |  |  |  |  | 0 |  |  | 0 | 0 | 0 | 0 | 12, 6 |  |
| 259-245 |  |  |  |  |  |  |  | 1,390 | 4,600 | 0 | 1,997 | 10, 2 |  |
| 259-246 |  |  |  |  |  |  |  |  | 0 | 0 | 0 |  |  |
| 259-251 | 13,000 | 4,000 | 600 | 5,600 | 4,200 | 11,500 | 1,500 | 2,800 | 7,800 | 12,000 | 6,300 | 39,4 |  |
| 259-252 | 300 |  | 50 | 0 | 500 | 500 | 0 | 260 | 60 | 14,000 | 1,841 |  |  |
| 259-253 |  |  |  | 300 |  |  | 0 | 0 | 300 | 10 | 122 |  |  |
| 259-254 | 4,000 | 1,500 | 200 | 1,300 | 2,700 | 4.000 | 650 | 3,000 | 10,500 | 9,950 | 3,780 |  |  |
| 259-255 | 1,700 | 4,000 | 300 | 3,200 | 11,000 | 60 | 2,600 |  | 0 | 0 | 2,540 |  |  |
| Mainland District |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 262-101 |  | 0 |  |  |  | 200 | 50 | 100 | 900 |  | 250 |  |  |
| 262-102 |  | 0 |  |  |  | 600 | 0 | 300 | 0 | 0 | 150 | 12,00 |  |
| 262-103 |  | 0 |  |  |  | 0 | 30 | 50 | 2,200 | 0 | 380 |  |  |
| 263-104 |  | 0 |  |  |  | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 |
| 262-105 |  | 0 |  |  |  | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 |
| 262-106 |  | 0 |  |  |  | 4,000 | 0 | 500 | 0 | 0 | 750 |  | 0 |
| 262-106A |  |  |  |  |  |  |  |  | 1,500 |  | 1,500 |  |  |
| 262-107 |  | 0 |  |  |  | 0 | 0 | 0 | 1,200 | 0 | 200 |  | 0 |
| 262-108 |  | 50 |  |  |  | 300 | 0 | 0 | 600 | 0 | 158 | 3,6 |  |
| 262-151 | 0 | 100 | 0 | 7,400 | 0 | 0 | 0 | 0 | 0 | 0 | 750 |  | 0 |
| 262-151A |  |  | 0 |  |  |  |  |  |  |  | 0 |  |  |
| 262-152 | 70,000 | 15,000 | 15,000 | 11,000 | 40,000 | 25,000 | 40,000 | 11,000 | 25,000 | 22,000 | 27,400 | 15,00 | 0 * |
| 262-152A |  |  | 0 | 0 | 500 |  |  |  |  |  | $167$ |  |  |
| 262-153 | 99,000 | 20,000 | 500 | 27,000 | 70,000 | 19,050 | 42,000 | 33,000 | 15,000 | 9.000 | 33,455 | 48,00 | 0 * |
| 262-153A |  |  | 0 | $0$ | 0 |  |  |  |  |  | 0 |  |  |
| 262-153B |  |  |  |  | 250 |  |  |  |  |  | 250 |  |  |
| 262-154 | 0 | 0 | 0 | 40 | 300 | 0 | 0 | 0 | 0 | 0 | 34 |  | 0 |

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| Stream No. | Peak Count |  |  |  |  |  |  |  |  |  |  |  | Index <br> Stream |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1969 | 1971 | 1973 | 1975 | 1977 | 1979 | 1981 | 1983 | 1985 | 1987 | $\begin{array}{r} \text { Mean } \\ (1969-87) \end{array}$ | 1989 S |  |
| 262-155 |  |  |  | 0 | 50 |  |  |  | 0 | 100 | 38 | 800 |  |
| 262-156 |  |  |  | 150 | 30 |  |  |  |  |  | 90 |  |  |
| 262-157 |  |  |  | 0 | 0 |  |  |  |  |  | 0 |  |  |
| 262-158 |  |  |  | 0 | 0 |  |  |  |  |  | 0 |  |  |
| 262-159 |  |  |  | 0 | 200 |  |  |  |  |  | 100 |  |  |
| 262-201 | 25 | 0 | 0 | 0 | 2,000 | 600 | 0 | 200 | 200 | 3,000 | 603 | 3,000 |  |
| 262-202 | 0 | 0 | 6,000 | 0 | 0 | 0 | 0 | 0 | 0 | 200 | 620 | 0 | 0 |
| 262-203 | 10,000 | 0 | 1,500 | 0 | 3,000 | 0 | 1,500 | 2,000 | 2,000 | 0 | 2,000 |  | 0 * |
| 262-204 | 500 | 0 | 0 | 0 | 600 | 100 | 1,100 | 300 | 0 | 0 | 260 | 0 | 0 |
| 262-205 | 2,000 | 1,000 | 500 | 2,500 | 3.500 | 6,000 | 11,000 | 450 | 7,700 | 3,400 | 3,805 | 42,000 | 0 * |
| 262-253 | 0 | 0 | 0 | 0 | 0 | 0 | 50 | 0 | 1,700 |  | 194 |  |  |
| 262-254 | 20 | 100 |  | 0 |  | 0 | 0 | 20 | 300 | 0 | 55 | 100 |  |
| 262-255 | 1,500 | 0 | 50 | 0 | 100 |  | 0 | 220 | 75 |  | 243 |  |  |
| 262-256 | 0 | 0 | 50 | 0 | 100 |  | 0 | 16 | 150 |  | 40 |  |  |
| 262-271 | 0 | 0 | 0 | 15,000 | 5,000 | 3,000 | 2,000 | 500 | 2,000 | 1,000 | 2,850 | 2,000 | 0 * |
| 262-272 | 0 | 0 | 0 | 0 | 6,000 | 300 | 500 |  | 900 | 0 | 856 | 0 | 0 |
| 262-301 | 0 | 0 | 0 | 0 | 3,600 | 0 | 0 | 0 | 0 | 0 | 360 | 0 | 0 |
| 262-302 |  |  |  |  |  |  | 0 |  |  |  | 0 |  |  |
| 262-351 | 0 | 0 | 0 | 0 | 5,000 | 3,000 | 6,000 | 2,600 | 2,500 | 600 | 1,970 | $9,000$ |  |
| 262-352 |  |  |  |  |  | 6,000 | 0 | 0 |  |  | 2,000 | $3,800$ |  |
| 262-401 | 0 | 0 | 0 | 0 | 1,500 | 1. 200 | 9,300 | 500 | 2,700 | 100 | 1,530 | 3,800 |  |
| 262-402 | 3,000 | 100 | 400 | 4,200 | 10,500 | 5,000 | 14,000 | 2,000 | 3,000 | 0 | 4,220 | 19,000 | 0 * |
| 262-451 | 13,000 |  | 500 | 34,000 | 55,000 | 55,000 | 89,000 | 45,000 | 45,000 | 6,500 | 38,111 | 37,000 | 0 |
| 262-452 |  | 200 | 800 |  | 0 | 3,000 | 200 | 0 | 0 |  | 600 |  |  |
| 262-453 |  |  |  |  |  |  | 0 | 1,100 | 0 |  | 367 |  |  |
| 262-501 | 3,000 | 500 | 1,000 | 0 | 15,000 | 21,000 | 51,200 | 4,500 | 16,500 | 14,000 | 12,670 | $200,000$ | 0 * |
| 262-502 |  |  |  |  | 200 | 0 | 6,000 | 200 |  | 700 | $1,420$ | $0$ | 0 |
| 262-551 | 18,000 | 1,000 | 3,000 | 29.000 | 76,000 | 77,000 | 31,000 | 8,000 | 17,000 | $18,000$ | $27,800$ | 38,000 | 0 * |
| 262-552 |  |  |  |  |  | 0 | $0$ | 0 | $0$ | $0$ | $0$ | $0$ | 0 |
| 262-601 | 0 | 0 |  |  | 0 | 0 | 0 | 0 | 0 |  | 0 |  | 0 |
| 262-602 | 2,000 | 5,000 | 2,000 | 0 | 3,000 | 0 | 0 | 0 |  | 2,000 | 1,556 | 0 | 0 |
| 262-603 | 200 | 0 | 6,000 | 0 | 200 | 0 | 0 | 0 | 1,900 | 0 | -830 |  | 0 |
| 262-604 | 14,000 | 18,000 | 9,500 | 16,000 | 60,000 | $38,700$ | 37,000 | 33,000 | $45,000$ | $63,000$ | 33,420 | $80,000$ | 0 * |
| 262-605 |  |  |  |  | 4,300 | 1,200 |  | 0 | 1,200 | $700$ | 1,480 | 10,000 |  |
| 262-606 |  |  |  |  |  |  | 0 |  |  | 0 | 0 | 20 |  |
| 262-651 | 7,000 | 6,000 | 6,000 | 5,200 | 28,000 | 17.000 | 11,000 | 12,000 | 42,000 | 51,000 | 18,520 | 445,000 | 0 * |
| 262-652 | 1,300 | 800 | 500 | 20 | 0 | 300 | 1.500 | 2,000 | 15,000 | 28,000 | 4,942 | 162,000 |  |

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| Stream No. | Peak Count |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1969 | 1971 | 1973 | 1975 | 1977 | 1979 | 1981 | 1983 | 1985 | 1987 | $\begin{array}{r} \text { Mean } \\ (1969-87) \end{array}$ | 1989 S | Index Stream |
| 262-653 | 1,200 | 2,500 | 500 | 10 |  | 250 | 5,100 | 2,800 | 7,600 | 18,000 | 4,218 | 230,000 |  |
| 262-654 | 3,300 | 5,000 | 2,500 | 2,000 | 9,000 | 0 | 0 | 0 | 2,000 | 1,100 | 2,490 | -0 | 0 |
| 262-655 |  | 6,000 |  |  |  | 700 | 0 |  |  | 1,400 | 2,025 | 300 |  |
| 262-656 |  |  |  |  |  |  | 0 |  |  | 600 | 300 | 13,600 |  |
| 262-701 | 250 | 500 | 300 | 450 | 2,000 | 500 | 2,000 | 2,200 | 7,000 | 5,400 | 2,060 | 44,000 |  |
| 262-702 | 4,300 | 2,000 | 3,000 |  | 7,000 | 12,000 | 14,000 | 6,800 | 16,000 | 31,000 | 10,678 | 67,000 | - |
| 262-702A | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |
| 262-703 | 0 | 0 | 0 | 120 |  | 0 | 0 | 0 | 0 | 0 | 13 | 42,000 |  |
| 262-704 | 50 | 0 | 0 | 50 | 300 | 0 | 5,000 | 1,600 | 0 | 2,000 | 900 | 20,000 |  |
| 262-705 |  | 0 |  | 0 |  | 300 | 250 | 1,200 | 2,000 | 4,200 | 1,136 | 12,000 |  |
| 262-750 |  | 0 |  | 0 |  | 300 | 250 |  |  |  | 138 |  |  |
| 262-751 | 4,000 | 2,500 | 5,000 | 2,200 |  | 5,700 | 15,000 | 3,100 | 31,000 | 13,000 | 9,056 | 73,000 | 0 |
| 262-752 | 1,500 | 0 | 0 | 3,500 |  | 400 | 0 | 200 | 1, 0 | 4,700 | 1,144 | 10,000 |  |
| 262-801 | 600 | 1,000 | 300 | 80 | 1,500 | 1,200 | 500 | 800 | 1,200 | 2,600 | 978 | 22,000 | O |
| 262-802 | 16,000 | 6,000 | 2,000 | 3,000 | 10,500 | 40,000 | 50,000 | 6,500 | 14,000 | 21,000 | 16,900 | 76,000 | - |
| 262-803 | 0 | 100 |  | 0 | 200 | 200 | 0 | 0 | 150 | 0 | 72 | 500 |  |
| 262-804 |  |  |  |  |  |  |  |  |  | 0 | 0 | 200 |  |
| 262-851 | 21,000 | 4.000 | 5,000 | 13,000 | 112,000 | 184,000 | 65,000 | 40,000 | 69,000 | 141,000 | 65,400 | 1,320,000 | * |
| 262-851A |  |  |  |  |  |  |  |  |  | 0 | - 0 |  |  |
| 262-852 | 2,000 | 600 | 0 | 2,600 | 4.000 | 4,500 | 16,000 | 5,000 | 11,000 | 12,500 | 5,820 | 91,000 |  |
| 262-853 | 3,800 | 0 | 0 | 200 | 600 | 500 | 2,500 | 6,000 | 4,200 | 8,200 | 2,600 | 66,000 |  |
| 262-854 | 500 | 0 | 0 | 0 | 300 | 100 | 20 | 50 | 2,800 | 8,600 | 1,237 | 10,000 |  |
| 262-854A |  |  |  |  | 25 |  |  |  |  |  | 1. 25 | 10,000 |  |
| 262-855 | 0 | 0 |  | 4 | 0 |  | 0 | 15 |  | 0 | 3 |  |  |
| 262-856 | 500 | 0 | 0 | 0 |  | 6,200 | 3,200 | 4,000 | 8,000 | 3,200 | 2,789 | 100 |  |
| 262-856A |  |  |  |  |  | 2,600 |  | 3,000 | 7,800 | 19,000 | 8,100 |  |  |
| 262-856B |  |  |  |  |  | 400 |  | 200 | 400 | 100 | 275 |  |  |
| 262-857 | 0 | 0 | 0 | 10 | 0 | 0 | 0 | 0 |  |  | 1 |  |  |
| 262-858 | 0 | 5,000 | 0 | 8,700 | 0 | 2,000 | 0 | 0 | 0 | 5,000 | 2,070 | 1,900 |  |
| 262-859 | 0 | 6,000 |  | 0 | 300 | 200 |  | 500 | 0 |  | 1,000 | 0 |  |
| 262-859A |  |  |  |  | 75 | 0 |  | 0 |  | 4,500 | 1,144 |  |  |
| 262-951 |  |  |  |  | 650 |  |  |  |  |  | 1,650 | 24,000 |  |

a Total escapement counted through fish weirs.
${ }^{\mathrm{b}}$ Afognak River fish weir was moved downstream in 1986. Counts preceeding were adjusted to account for fish spawning between the two weir sites.
${ }^{\text {c }}$ Little Waterfall (251-822) escapement counts pre-1981 were excluded owing to additional spawning habitat created by a fishpass.
${ }^{\mathrm{d}}$ Saltery Creek (259-415) escapement counts reflect estimates of fish spawning downstream of the weir.

Appendix A.3. Odd-year pink salmon preemergent fry densities (live fry $/ \mathrm{m}^{2}$ ) by stream and brood year, Kodiak Management Area, 1965-1989.

| Stream | 1965 | 1967 | 1969 | 1971 | 1973 | 1975 | 1977 | 1979 | 1981 | 1983 | 1985 | 1987 | $\begin{gathered} \text { Average } \\ 1965-1987 \end{gathered}$ | $\begin{gathered} 90 \% \mathrm{CI} \\ 1965-1987 \end{gathered}$ | 1989 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Alinchak |  |  |  |  | 0.0 |  | 60.2 | 0.0 | 2.0 | 49.0 | 0.9 |  | 18.7 | 4.2-50.4 | 27.3 |
| American | 8.8 | 85.4 | 150.4 | 93.2 | 83.6 | 219.1 | 361.4 | 33.3 | 174.6 | 0.0 | 126.6 | 378.8 | 142.9 | 78.4-207.4 | 256.0 |
| Barling |  | 13.4 | 36.0 | 0.0 | 20.1 | 63.3 | 107.3 | 39.0 | 168.2 | 289.9 | 147.7 | 210.3 | 99.6 | 48.2-150.9 | 51.7 |
| Bauman's | 119.4 | 353.8 | 107.0 | 113.0 | 108.1 | 317.2 | 375.9 | 401.5 | 188.1 | 252.9 | 560.4 | 863.2 | 313.4 | 196.5-430.2 | 240.7 |
| Beaver |  |  |  |  |  |  |  |  | 1,037.0 | 493.9 | 171.4 | 192.1 | 473.6 | $0-948.3^{\text {a }}$ | 549.4 |
| Big Creek |  |  |  |  | 4.7 |  | 138.9 | 36.5 |  | 0.0 |  | 395.0 | 115.0 | 0-273.5 ${ }^{\text {a }}$ | 131.8 |
| Buskin | 64.7 | 11.5 | 46.3 | 35.9 | 31.8 | 51.5 | 164.3 | 141.8 | 374.8 | 277.5 | 248.4 | 267.6 | 143.0 | 79.7-206.3 | 89.8 |
| Dakavak |  |  |  |  | 85.1 | 19.5 | 111.2 | 0.3 | 85.8 | 0.0 |  | 117.0 | 55.6 | 21.9-97.7 | 0.4 |
| Danger |  | 157.6 | 31.1 | 227.9 | 0.0 |  | 156.0 | 16.2 | 67.0 | 10.2 | 56.9 | 174.5 | 89.7 | 42.4-137.1 | 292.5 |
| Deadman | 283.8 | 403.5 | 351.1 | 725.4 | 218.7 | 532.2 | 691.5 | 238.3 | 755.2 | 242.4 | 349.6 | 400.3 | 432.7 | 330.9-534.4 | 546.9 |
| Dog Salmon | 73.8 | 254.7 | 46.7 | $8 \mathrm{8}$. | 0.0 | 83.0 | 67.2 | 21.2 | 49.6 | 120.4 | 50.1 | 400.3 | 104.6 | 46.0-163.2 | 179.9 |
| Humpy | 5.1 | 281.7 | 80.2 | 144.3 | 88.7 | 422.4 | 508.0 | 35.9 | 228.1 | 220.1 | 135.1 | 491.1 | 220.0 | 126.1-313.9 | $464 . \mathrm{B}$ |
| Hurse |  |  | 96.3 | 172.7 | 22.7 | 12.7 | 69.5 | 34.6 | 0.0 | 0.0 | 0.1 | 84.9 | 49.4 | 16.7-82.1 | 102.9 |
| Kaiugnak | 60.4 | 908.8 | 68.4 | 0.9 | 21.6 | 213.6 | 768.2 | 9.8 | 295.2 | 55.5 | 421.4 | 72.2 | 241.3 | 81.6-401.1 | 736.8 |
| Kanatak |  |  |  |  |  |  | 48.2 | 124.9 | 452.9 | 159.2 |  | 83.6 | 173.8 | 19.8-327.8 | 186.7 |
| Karluk |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 0.1 |
| Kashvik |  |  |  |  | 5.3 | 12.5 | 24.0 | 20.5 | 2.3 | 164.9 |  | 82.1 | 44.5 | 0.8-88.2 | 0.8 |
| Kiliuda |  |  | 81.6 | 14.1 | 3.6 | 19.3 | 0.8 | 11.8 | 90.1 | 5.0 | 104.5 | 15.5 | 34.6 | 11.5-57.8 | 77.0 |
| Kinak |  |  |  |  |  | 136.2 | 280.4 | 179.6 | 105.8 | 111.1 |  | 42.7 | 142.6 | 76.1-209.2 | 161.7 |
| Kukak |  |  |  |  |  |  |  | 106.8 | 29.3 | 20.2 |  |  | 52.1 | 0-132.3 ${ }^{\text {a }}$ | 6.7 |
| Little |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 5.6 |
| Little wtfl. |  |  |  |  |  |  |  |  | 586.2 | 839.5 | 339.0 | 172.3 | 484.2 | 141.2-827.3 | 70.0 |
| Marka |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 10.3 |
| Miam |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 118.6 |
| Missak |  |  |  |  |  | 130.1 | 126.0 | 59.4 | 63.3 | 3.6 |  | 214.6 | 99.5 | 39.1-160.0 | 131.8 |

-Continued-

Appendix A.3. (page 2 of 2)

| STREAM | 1965 | 1967 | 1969 | 1971 | 1973 | 1975 | 1977 | 1979 | 1981 | 1983 | 1985 | 1987 | $\begin{gathered} \text { Average } \\ 1965-1987 \end{gathered}$ | $\begin{gathered} 90 \div \mathrm{CI} \\ 1965-1987 \end{gathered}$ | 1989 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Monashka |  |  |  |  |  |  | 3.6 | 21.0 | 50.6 | 119.8 | 3.3 |  | 39.6 | 22.9-56.4 | 2.7 |
| Narrows | 77.3 | 225.3 | 123.0 | 47.2 | 0.0 | 70.5 | 185.8 | 4.8 | 27.6 | 107.0 | 96.8 | 135.9 | 91.8 | 55.9-127.6 | 264.日 |
| Paramanof |  | 0.0 | 2.8 | 27.9 | 15.2 | 6.1 | 13.3 | 26.1 | 85.5 | 44.6 | 5.7 | 5.5 | 21.2 | 7.4-34.9 | 278.3 |
| Perenosa | 137.9 | 363.2 | 428.2 | 347.0 | 146.9 | 305.6 | 111.4 | 234.2 | 251.9 | 46.1 | 216.9 | 39.2 | 219.1 | 153.6-284.6 | 72.3 |
| Portage |  |  |  |  |  |  | 73.2 | 0.0 | 0.0 | 4.8 |  |  | 19.5 | 0-61.7 ${ }^{\text {a }}$ | 2.3 |
| Saltery | 35.2 | 76.3 | 33.9 | 34.2 | 72.4 | 126.6 | 194.8 | 81.5 | 163.2 | 46.7 | 44.4 | 227.5 | 94.7 | 59.7-129.7 | 521.5 |
| Seal Bay |  |  |  |  |  |  | 616.0 | 99.0 | 331.5 | 89.8 | 582.2 | 171.7 | 315.0 | 120.3-509.8 | 869.8 |
| 7-Rivers | 115.1 | 210.0 | 304.7 | 325.2 | 137.5 | 338.0 | 260.4 | 173.7 | 595.8 | 135.0 | 569.2 | 252.0 | 284.7 | 202.8-366.7 | 383.5 |
| Sheratin | 10.1 | 67.4 | 92.0 |  | 46.6 | 15.8 | 89.4 | 21.9 | 20.9 | 7.0 | 68.7 | 7.1 | 40.6 | 21.5-59.7 | 39.8 |
| Sid old's | 1.4 | 100.2 | 163.6 | 134.2 | 43.7 | 91.1 | 31.8 | 50.1 | 44.5 | 183.8 | 44.4 | 227.5 | 93.0 | 56.6-129.4 | 116.4 |
| Terror | 1.2 | 4.1 | 7.1 | 11.9 | 7.4 | 47.0 | 11.7 | 7.0 | 21.6 | 2.0 | 0.2 | 16.2 | 11.4 | 4.8-18.1 | 0.8 |
| Uganik | 4.2 | 96.5 | 9.9 | 20.9 | 166.7 | 92.1 | 67.1 | 154.2 | 172.9 | 7.1 | 0.4 | 1.4 | 66.1 | 30.5-101.7 | 9.4 |
| Uyak 202 | 59.5 | 175.2 | 313.8 | 187.7 | 70.9 | 377.8 | 258.5 | 199.1 | 605.5 | 233.4 | 162.5 | 447.7 | 257.6 | 175.9-339.4 | 1.070 .0 |
| Uyak 203 |  |  |  |  |  |  | 123.0 | 0.5 | 340.9 | 700.3 | 0.0 | 80.8 | 207.6 | 0-431.3 ${ }^{\text {a }}$ | 238.0 |
| zachar | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 3.6 | 34.7 | 0.0 | 45.7 | 0.0 | 0.0 | 38.0 | 10.2 | 0.9-19.4 | 12.1 |
| Total | 1,057.9 | 3,788.5 | 2,573.9 | $2,751.8$ | 1,401.4 | 3,707.8 | 6,133.8 | 2,584.6 | 7.513.5 | 5,042.8 | 4,506.6 | 6,308.5 | 5,303.2 | 175.9-339.4 | 8,321.7 |

${ }^{\text {a }}$ Zero was assigned to lower confidence intervals which were negative numbers.

Appendix A.4. Pink salmon escapement goals for odd and even year populations for Kodiak Management Area index streams employing peak counts and estimated total escapement.

| Index Stream | Even Year Indexed Escapement goals | Odd-Year Escapement goals |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Indexed ${ }^{\text {a }}$ |  | ted Tota |  |
| Name No. | Minimum Desired | Minimum Desired | Minimum | Desired | Mid Point |

## Afognak District


-Continued-

Appendix A.4. (page 2 of 4)

| Index Stream | Even Year Indexed Escapement qoals | Odd-Year Escapement goals |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Indexed ${ }^{\text {a }}$ |  | ted Tota |  |
| Name No. | Minimum Desired | Minimum Desired | Minimum | Desired | Mid Point |

## Southwest District

| Karluk ${ }^{\text {b }}$ | 255-101 | 800,000 | 1,600,000 | 20,000 | 60,000 | 20,000 | 60,000 | 40,000 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Sturgeon | 256-401 | 50,000 | 150,000 | 5,000 | 15,000 | 9,200 | 27,600 | 18,400 |
| Ayakulik ${ }^{\text {b }}$ | 256-201 | 400,000 | 800,000 | 5,000 | 15,000 | 5,000 | 15,000 | 10,000 |
| Subtotal |  | 250,000 | 2,550,000 | 30,000 | 90,000 | 34,200 | 103,600 | 68,000 |

Alitak Bay District

| Narrows | 257-401 | 2,000 | 6,000 | 2,000 | 6,000 | 3,700 | 11,000 | 7,400 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Dog Salmon | 257-403 | 50,000 | 150,000 | 60,000 | 180,000 | 60,000 | 180,000 | 120,000 |
| Deadman | 257-502 | 40,000 | 120,000 | 60,000 | 180,000 | 110,500 | 331,500 | 221,000 |
| Humpy | 257-701 | 70,000 | 210,000 | 90,000 | 270,000 | 165,800 | 497,300 | 331,500 |
| Subtotal |  | 162,000 | 486,000 | 212,000 | 636,000 | 340,000 | 1,019,500 | 679,900 |
| Northeast District |  |  |  |  |  |  |  |  |
| Sid Olds | 259-242 | 30,000 | 90,000 | 30,000 | 90,000 | 55,200 | 165,800 | 110,500 |
| American | 259-231 | 30,000 | 90,000 | 30,000 | 90,000 | 55,200 | 165,800 | 110,500 |
| Buskin ${ }^{\text {b }}$ | 259-211 | 60,000 | 180,000 | 50,000 | 150,000 | 50,000 | 150,000 | 100,000 |
| Subtotal |  | 120,000 | 360,000 | 110,000 | 330,000 | 160,400 | 481,600 | 321,000 |
| Eastside District |  |  |  |  |  |  |  |  |
| Seven Rivers | 258-701 | 40,000 | 120,000 | 40,000 | 120,000 | 73,700 | 221,000 | 147,300 |
| Kaiugnak | 258-542 | 10,000 | 30,000 | 10,000 | 30,000 | 18,400 | 55,200 | 36,800 |
| Barling | 258-522 | 30,000 | 90,000 | 30,000 | 90,000 | 55,200 | 165,800 | 110,500 |
| Kiliuda | 258-207 | 20,000 | 60,000 | 10,000 | 30,000 | 18,400 | 55,200 | 36,800 |

- Continued-

Appendix A.4. (page 3 of 4 )

| Index Stream |  | Even Year Indexed Escapement goals |  | Odd-Year Escapement qoals |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Indexed ${ }^{\text {a }}$ | Estimated Total |  |  |
| Name | No. |  |  | Minimum | Desired | Minimum | Desired | Minimum | Desired | Mid Point |
| Saltery ${ }^{\text {b }}$ | 259-415 | 20,000 | 60,000 | 30,000 | 90,000 | 30,000 | 90,000 | 60,000 |
| Miam | 259-412 | 20,000 | 60,000 | 10,000 | 30,000 | 18,400 | 55,200 | 36,800 |
| Hurst | 259-414 | 10,000 | 30,000 | 10,000 | 30,000 | 18,400 | 55,200 | 36,800 |
| Subtotal |  | 150,000 | 450,000 | 140,000 | 420,000 | 232,500 | 697,600 | 465,000 |
| Mainland District |  |  |  |  |  |  |  |  |
| Big River | 262-152 | 10,000 | 30,000 | 10,000 | 30,000 | 18,400 | 55,200 | $36.800^{\text {c }}$ |
| Village | 262-153 | 15,000 | 45,000 | 15,000 | 45,000 | 27,600 | 82,900 | 55,200 ${ }^{\text {C }}$ |
| Cape Chiniak | 262-205 | 5,000 | 15,000 | 3,000 | 9,000 | 5,500 | 16,600 | 11,000 ${ }^{\text {c }}$ |
| Big Hallo | 262-203 | 2,000 | 6,000 | 2,000 | 6,000 | 3,700 | 11,000 | $7,400^{\text {C }}$ |
| Kukak | 262-271 | 3,000 | 9,000 | 2,000 | 6,000 | 3,700 | 11,000 | 7,400 |
| Missak | 262-402 | 5,000 | 15,000 | 3,000 | 9,000 | 5,500 | 16,600 | 11,000 |
| Kinak | 262-451 | 20,000 | 60,000 | 20,000 | 60,000 | 36,800 | 110,500 | 73,700 |
| Geographic | 262-501 | 4,000 | 12,000 | 4,000 | 12,000 | 7,400 | 22,100 | 14,700 |
| Dakavak | 262-551 | 25,000 | 75,000 | 20,000 | 60,000 | 36,800 | 110,000 | 73,700 |
| Kashvik | 262-604 | 25,000 | 75,000 | 25,000 | 75,000 | 46,000 | 138,100 | 92,100 |
| Big Alinchak | 262-651 | 30,000 | 90,000 | 20,000 | 60,000 | 36,800 | 110,500 | 73,700 |
| Portage | 262-702 | 15,000 | 45,000 | 10,000 | 30,000 | 18,400 | 55,200 | 36,800 |
| Oil | 262-751 | 15,000 | 45,000 | 10,000 | 30,000 | 18,400 | 55,200 | 36,800 |
| Jute | 262-801 | 2,000 | 6,000 | 1,000 | 3,000 | 1,800 | 5,500 | 3,700 |
| Kanatak | 262-802 | 10,000 | 30,000 | 10,000 | 30,000 | 18,400 | 55,200 | 36,800 |
| Big Creek | 262-851 | 70,000 | 210,000 | 60,000 | 180,000 | 110,500 | 331,500 | 221,000 |
| Subtotal |  | 256,000 | 768,000 | 215,000 | 645,000 | 395,700 | 1,187,100 | 791,800 |
| Grand Total |  | 2,401,000 | 6,003,000 | 1,010,000 | 3,030,000 | 1,697,000 | 5,092,600 | 3,394,400 |
| Estimated total Escapement ${ }^{\text {d }}$ |  |  |  |  |  | 2,341,900 | 7,027,800 | $4,684,300^{e}$ |

-Continued-

Appendix A.4. (page 4 of 4 )
${ }^{\text {a }}$ Non weir systems indexed escapements are peak count derived.
${ }^{b}$ Systems with escapements enumerated through weirs.
c Index streams for which total available spawning habitat was not estimated; these streams escapement goals were not included in deriving overall spawner density for all index streams.
d Production from these 51 index streams account for $72.70 \%$ of the total odd-year pink salmon production founded upon indexed escapements during 1969-1987. Estimated total escapements were generated by expanding the minimum, desired, and mid-point total estimated escapements for index streams by a factor of 1.38 (Barrett et al. 1990).
e The mid-point estimated total escapement figure minus the escapement figures for those streams that did not have total available spawning habitat estimated divided by the total habitat estimate for all Kodiak Management Area index streams $(4,573,900 / 3,254,454)$ provides a spawning density of $1.4 \mathrm{fish} / \mathrm{m}^{2}$ of spawning habitat based upon the mid-point escapement goal for odd-year brood line

Appendix A.5. Pink salmon daily weir escapement counts for systems selected for stream life estimation, 1990.

| Date | Weired System Counts |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Afognak |  | Pink |  | Barling |  | Akalura |  | E. Paramanof |  | Saltery |  |
|  | Daily | Cum. | Daily | Cum. | Daily | Cum. | Daily | Cum. | Daily | Cum. | Daily | Cum. |
| 7/10 | 1. | 1 |  |  |  |  |  |  |  |  |  |  |
| 7/11 | 0 | 1 |  |  |  |  |  |  |  |  |  |  |
| 7/12 | 0 | 1 |  |  |  |  |  |  |  |  |  |  |
| 7/13 | 0 | 1 |  |  |  |  |  |  |  |  | 3 | 3 |
| 7/14 | 2 | 3 |  |  |  |  |  |  |  |  | 0 | 3 |
| 7/15 | 0 | 3 |  |  |  |  |  |  |  |  | 0 | 3 |
| 7/16 | 0 | 3 |  |  |  |  |  |  | 2 | 2 | 0 | 3 |
| 7/17 | 0 | 3 |  |  |  |  |  |  | 10 | 12 | 2 | 5 |
|  | 2 | 5 |  |  |  |  |  |  | 23 | 35 | 0 | 5 |
| 7/19 | 0 | 5 |  |  |  |  |  |  | 13 | 48 | 10 | 15 |
| 7/20 | 1 | 6 |  |  |  |  |  |  | 10 | 58 | 0 | 15 |
| $7 / 21$ | 0 | 6 |  |  |  |  |  |  | 0 | 58 | 4 | 19 |
| $7 / 22$ | 16 | 22 |  |  |  |  |  |  | 190 | 248 | 31 | 50 |
| 7/23 | 5 | 27 |  |  |  |  |  |  | 54 | 302 | 6 | 56 |
| 7/24 | 0 | 27 |  |  | 10 | 10 |  |  | 243 | 545 | 30 | 86 |
| 7/25 | 0 | 27 |  |  | 3 | 13 |  |  | 171 | 716 | 21 | 107 |
| 7/26 | 1 | 28 |  |  | 2 | 15 |  |  | 89 | 805 | 0 | 107 |
| 7/27 | 63 | 91 |  |  | 8 | 23 |  |  | 654 | 1,459 | 68 | 175 |
| 7/28 | 370 | 461 |  |  | 16 | 39 | 2 | 2 | 223 | 1,682 | 59 | 234 |
| 7/29 | 88 | 549 |  |  | 19 | 58 | 0 | 2 | 142 | 1, 824 | 110 | 344 |
| 7/30 | 13 | 562 |  |  | 18 | 76 | 0 | 2 | 66 | 1,890 | 90 | 434 |
| 7/31 | 93 | . 655 |  |  | 2 | 78 | 0 | 2 | 1,116 | 3,006 | 41 | 475 |
| 8/1 | 689 | 1,344 | 3 | 3 | 0 | 78 | 0 | 2 | 1,263a | 4,269 | 98 | 573 |
| 8/2 | 150 | 1,494 | 0 | 3 | 0 | 78 | 0 | 2 | 1,263 ${ }^{\text {a }}$ | 5,532 | 20 | 593 |
| 8/3 | 323 | 1,817 | 0 | 3 | 51 | 129 | 0 | 2 | 1,263 ${ }^{\text {a }}$ | 6,795 | $62^{\text {a }}$ | 655 |
| 8/4 | 313 | 2,130 | 14 | 17 | 24 | 153 | 0 | 2 | 1,263 ${ }^{\text {a }}$ | 8,058 | $62^{\text {a }}$ | 717 |
| 8/5 | 125 | 2,255 | 28 | 45 | 84 | 237 | 0 | 2 | 498 | 8,556 | $62^{\text {a }}$ | 779 |
| 8/6 | 123 | 2,378 | 31 | 76 | 98 | 335 | 0 | 2 | 511 | 9,067 | $62^{\text {a }}$ | 841 |
| 8/7 | 1, 124 | 3,502 | 94 | 170 | 64 | 399 | 0 | 2 | 1,188 | 10,255 | 117 | 958 |
| 8/8 | 1,995 | 5,497 | 283 | 453 | 302 | 701 | 1 | 3 | 95 | 10,350 | 239 | 1,197 |

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Appendix A.5. (page 2 of 3 )

Weired System Counts

| Date | Afognak |  | Pink |  | Barling |  | Akalura |  | E. Paramanof |  | Saltery |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Daily | Cum. | Daily | Cum. | Daily | Cum. | Daily | Cum. | Daily | Y Cum. | Daily | Cum. |
| 8/9 | 1,472 | 6,969 | 17 | 470 | 231 | 932 | 0 | 3 | 27 | 10,377 | 226 | 1,423 |
| $8 / 10$ | 295 | 7,264 | 4 | 474 | 335 | 1,267 | 0 | 3 | 59 | 10,436 | 100 | 1,523 |
| 8/11 | 1,635 | 8,899 | 995 ${ }^{\text {a }}$ | 1,465 | 250 | 1,517 | 6 | 9 | 0 | 10,436 | 100 | 1,623 |
| 8/12 | 654 | 9,553 | 156 | 1,621 | 260 | 1,777 | 6 | 15 | 0 | 10,436 | 55 | 1,678 |
| 8/13 | 770 | 10,323 | 113 | 1,734 | 323 | 2,100 | 0 | 15 | 0 | 10,436 | 117 | 1,795 |
| 8/14 | 935 | 11,258 | 30 | 1,764 | 370 | 2,470 | 6 | 21 | 117 | 10,553 | 258 | 2,053 |
| $8 / 15$ | 1,181 | 12,439 | 172 | 1,936 | 379 | 2,849 | 3 | 24 | 51 | 10,604 | 39 | 2,092 |
| 8/16 | 3,287 | 15,726 | 192 | 2,128 | 804 | 3,653 | 0 | 24 | 4 | 10,608 | 199 | 2,291 |
| ¢ 8/17 | 2,966 | 18,692 | 26 | 2,154 | 237 | 3,890 | 7 | 31 | 71 | 10,679 | 27 | 2,318 |
| ¢ 8/18 | 304 | 18,966 | 54 | 2,208 | 200 | 4,090 | 28 | 59 | 79 | 10,683 | 137 | 2,455 |
| 8/19 | 1,447 | 20,443 | 163 | 2,371 | 653 | 4,743 | 40 | 99 | 143 | 10,826 | 126 | 2,581 |
| $8 / 20$ | 1,216 | 21,659 | 131 | 2,502 | 913 | 5,656 | 7 | 106 | 908 | 11,734 | 1 | 2,582 |
| $8 / 21$ | 500 | 22,159 | 29 | 2,531 | 988 | 6,644 | 13 | 119 | 185 | 11,919 | 55 | 2,637 |
| 8/22 | 246 | 22,405 | 61 | 2,592 | 575 | 7,219 | 4 | 123 | 1,470 | 13,389 | 95 | 2,732 |
| 8/23 | 255 | 22,660 | 27 | 2,619 | 829 | 8,048 | 4 | 127 | 830 | 14,219 | 115 | 2,847 |
| 8/24 | 317 | 22,977 | 4 | 2,623 | 447 | 8,495 | 60 | 187 | 643 | 14,862 | 85 | 2,932 |
| $8 / 25$ | 182 | 23,159 | 90 | 2,713 | 839 | 9,334 | 112 | 299 | 369 | 15,231 | 62 | 2,994 |
| $8 / 26$ | 60 | 23,219 | 266 | 2,979 | 703 | 10,037 | 368 | 667 | 690 | 15,921 | 118 | 3,112 |
| $8 / 27$ | 102 | 23,321 | 44 | 3,023 | 652 | 10,689 | 463 | 1,130 | 653 | 16,574 | 37 | 3,149 |
| $8 / 28$ | 68 | 23,389 | 40 | 3,063 | 181 | 10,870 | 51 | 1. 181 | 244 | 16,818 | 76 | 3,225 |
| 8/29 | 10 | 23,399 | 1 | 3,064 | 492 | 11,362 | 183 | 1,364 | 212 | 17,030 | 64 | 3,289 |
| $8 / 30$ | 14 | 23,413 | 3 | 3,067 | 186 | 11,548 | 38 | 1,402 | 478 | 17,508 | 154 | 3,443 |
| 8/31 | 92 | 23,505 | 10 | 3,077 | 539 | 12,087 | 176 | 1,578 | 147 | 17,655 | 135 | 3,578 |
| $9 / 1$ | 155 | 23,660 | 1 | 3,078 | 705 | 12,792 | 18 | 1,596 | 868 | 18,523 | 141 | 3,719 |
| $9 / 2$ | 99 | 23,759 | 114 | 3,192 | 359 | 13,151 | 186 | 1,782 | 605 | 19,128 | 132 | 3,851 |
| 9/3 | 671 | 24,430 | 997 | 4,189 | 1,453 | 14, 604 | 198 | 1,980 | 1,549 | 20,677 | 91 | 3,942 |
| 9/4 | 267 | 24,697 | 13 | 4,202 | 417 | 15,021 | 23 | 2,003 | 400 | 21,077 | 96 | 4,038 |
| 9/5 | 141 | 24,838 | 2 | 4,204 | 584 | 15,605 | 126 | 2,129 | 212 | 21,289 | 117 | 4,155 |
| 9/6 | 96 | 24,934 | 13 | 4,217 | 1,005 | 16,610 | 38 | 2,167 | 159 | 21,448 | 8 | 4,163 |
| 9/7 | 374 | 25,308 | 27 | 4,244 | 570 | 17,180 | 38 | 2,205 | 85 | 21,533 | 247 | 4,410 |

-Continued-

Appendix A.5. (page 3 of 3 )

Weired System Counts

${ }^{\text {a }}$ Counts represent estimates while weir was inoperable due to high water conditions.

Appendix A.6. Kodiak Management Area odd-year pink salmon escapement, catch, and run numbers by fishing district, 1969-1991.

| District | Brood Year | Estimated Escapement | Commercial Catch | Run |
| :---: | :---: | :---: | :---: | :---: |
| Afognak |  |  |  |  |
|  | 1969 | 186,747 | 32,819 | 219,566 |
|  | 1971 | 165,335 | 3,389 | 168,724 |
|  | 1973 | 91,279 | 57,911 | 149,190 |
|  | 1975 | 90,474 | 119,150 | 209,624 |
|  | 1977 | 340,793 | 161,474 | 502,267 |
|  | 1979 | 580,287 | 782,546 | 1,362,833 |
|  | 1981 | 984,370 | 1,440,345 | 2,424,715 |
|  | 1983 | 422,129 | 192,784 | 614,913 |
|  | 1985 | 975,721 | 1,132,382 | 2,108,103 |
|  | 1987 | 597,819 | 505,497 | 1,103,316 |
|  | 1989 | 1,991,845 |  | 1,991,845 |
|  | 1991 | 472,029 | 1,112,192 | 1,584,221 |
| Average | (1969-1987) | 443,495 | 442,830 | 886,325 |
| Northwest Kodiak |  |  |  |  |
|  | 1969 | 685,383 | 653,549 | 1,338,932 |
|  | 1.971 | 535,828 | 133,210 | 669,038 |
|  | 1973 | 397,817 | 218,327 | 616,144 |
|  | 1975 | 690,221 | 1,155,439 | 1,845,660 |
|  | 1977 | 1,023,479 | 1,670,335 | 2,693,814 |
|  | 1979 | 1,565,862 | 3,040,943 | 4,606,805 |
|  | 1981 | 1,716,753 | 2,600,000 | 4,316,753 |
|  | 1983 | 1,371,842 | 1,364,787 | 2,736,629 |
|  | 1985 | 1,452,625 | 1,801,177 | 3,253,802 |
|  | 1987 | 1,334,032 | 1,329,254 | 2,663,286 |
|  | 1989 | 5,653,387 |  | 5,653,387 |
|  | 1991 | 792,347 | 3,787,647 | 4,579,994 |
| Average | (1969-1987) | 1,077,384 | 1,396,702 | 2,474,086 |
| Southwest Kodiak |  |  |  |  |
|  | 1969 | 2,100 | 6,669 | 8,769 |
|  | 1971 | 9,347 | 294 | 9,641 |
|  | 1973 | 896 | 4,261 | 5,157 |
|  | 1975 | 2,778 | 11,923 | 14,701 |
|  | 1977 | 80,571 | 8,188 | 88,759 |
|  | 1979 | 206,975 | 15,234 | 222,209 |
|  | 1981 | 86,081 | 8,090 | 94,171 |
|  | 1983 | 81,841 | 753 | 82,594 |
|  | 1985 | 97,832 | 88,270 | 186,102 |
|  | 1987 | 40,079 | 241,954 | 282,033 |
|  | 1989 | 168,972 |  | 168,972 |
|  | 1991 | 188,005 | 1,375,838 | 1,563,843 |
| Average | (1969-1987) | 60,850 | 38,564 | 99,414 |

-Continued-

Appendix A.6. (page 2 of 3 )

| District | Brood <br> Year | Estimated Escapement | Commerci Catch | Run |
| :---: | :---: | :---: | :---: | :---: |
| Alitak Bay |  |  |  |  |
|  | 1969 | 816,069 | 3,775,182 | 4,591,251 |
|  | 1971 | 780,024 | 100,896 | 880,920 |
|  | 1973 | 309,780 | 49,932 | 359,712 |
|  | 1975 | 550,886 | 235,711 | 786,592 |
|  | 1977 | 961,780 | 961,673 | 1,923,453 |
|  | 1979 | 1,330,071 | 1,664,410 | 2,994,481 |
|  | 1981 | 1,460,981 | 2,073,629 | 3,534,610 |
|  | 1983 | 1,029,028 | 1,428,526 | 2,457,554 |
|  | 1985 | 1,864,240 | 1,057,940 | 2,922,180 |
|  | 1987 | 877,906 | 916,883 | 1,794,789 |
|  | 1989 | 4,483,424 |  | 4,483,424 |
|  | 1991 | 662,294 | 2,373,521 | 3,035,815 |
| Average | (1969-1987) | 998,076 | 1,226,478 | 2,224,555 |
| Eastside Kodiak |  |  |  |  |
|  | 1969 | 707,874 | 6,704,242 | 7,412,116 |
|  | 1971 | 482,507 | 592,876 | 1,075,383 |
|  | 1973 | 299,784 | 91,799 | 391,583 |
|  | 1975 | 550,094 | 382,177 | 932,271 |
|  | 1977 | 885,092 | 2,215,285 | 3,100,377 |
|  | 1979 | 1,373,155 | 3,685,457 | 5,058,612 |
|  | 1981 | 1,310,255 | 2,456,641 | 3,766,896 |
|  | 1983 | 959,501 | 783,039 | 1,742,540 |
|  | 1985 | 916,113 | 81,673 | 997,786 |
|  | 1987 | 746,297 | 817,847 | 1,564,144 |
|  | 1989 | 3,088,875 |  | 3,088,875 |
|  | 1991 | 1,329,082 | 5,650,427 | 6,979,509 |
| Average | (1969-1987) | 823,067 | 1,781,104 | 2,604,171 |
| Northeast Kodiak |  |  |  |  |
|  | 1969 | 421,086 | 1,262,800 | 1,683,886 |
|  | 1971 | 194,143 | 163,216 | 357,359 |
|  | 1973 | 100,543 | 16,211 | 116,754 |
|  | 1975 | 136,900 | 218,793 | 355,693 |
|  | 1977 | 377,880 | 135,921 | 513,801 |
|  | 1979 | 510,039 | 458,116 | 968,155 |
|  | 1981 | 458,266 | 416,920 | 875,186 |
|  | 1983 | 330,144 | 193,880 | 524,024 |
|  | 1985 | 840,387 | 203,409 | 1,043,796 |
|  | 1987 | 600,770 | 276,657 | 877,427 |
|  | 1989 | 1,520,402 |  | 1,520,402 |
|  | 1991 | 524,455 | 296,438 | 820,893 |
| Average (1969-1987) |  | 397,016 | 334,592 | 731,608 |

-Continued-

Appendix A.6. (page 3 of 3 )

| District | Brood <br> Year | Estimated Escapement | Commercial Catch | Run |
| :---: | :---: | :---: | :---: | :---: |
| Mainland |  |  |  |  |
|  | 1969 | 709,324 | 65,562 | 774,886 |
|  | 1971 | 254,828 | 153,319 | 408,147 |
|  | 1973 | 168,016 | 24,356 | 192,372 |
|  | 1975 | 437,996 | 270,804 | 708,800 |
|  | 1977 | 1,267,433 | 343,295 | 1,610,728 |
|  | 1979 | 1, 284, 305 | 623,117 | 1,907,422 |
|  | 1981 | 1,222,730 | 271,758 | 1,492,488 |
|  | 1983 | 569,060 | 183,735 | 752,795 |
|  | 1985 | 1, 022,058 | 261,059 | 1,283,117 |
|  | 1.987 | 1,239,439 | 228,238 | 1,467,677 |
|  | 1989 | 7,614,276 |  | 7,614,276 |
|  | 1991 | 2,231,218 | 1,166,188 | 3,397,406 |
| Average | (1969-1987) | 817,519 | 242,524 | 1,060,006 |

Appendix B.1. Pink salmon peak counts and estimated total escapement for Chignik Management Area streams, $1989^{\text {a }}$.

| Stream No. | Numbers of Fish |  | ETE/PC ${ }^{\text {b }}$ |
| :---: | :---: | :---: | :---: |
|  | Peak Count (PC) | Estimated Total Escapement (ETE) |  |
| Eastern District |  |  |  |
| 272-701 | 3,200 | 4,636 | 1.44 |
| 272-702 | 53,000 | 53,000 | 1.00 |
| 272-703 | 17,000 | 26,983 | 1.59 |
| 272-720 | 0 | 0 |  |
| 272-721 | 10,900 | 14,000 | 1.28 |
| 272-722 | 0 | 0 |  |
| 272-723 | 0 | 0 |  |
| 272-801 | 10,900 | 14,080 | 1.29 |
| 272-802 | 9,000 | 16,018 | 1.78 |
| 272-802A | 200 | 226 | 1.13 |
| 272-802B | 750 | 848 | 1.13 |
| 272-803 | 36,000 | 36,000 | 1.00 |
| 272-804 | 10,600 | 10,600 | 1.00 |
| 272-805 | 19,000 | 19,000 | 1.00 |
| 272-821 | 0 | 0 |  |
| 272-822 | 0 | 0 |  |
| 272-823 | 0 | 0 |  |
| 272-84 | 0 | 0 |  |
| 272-843 | 0 | 0 |  |
| 272-844 | 0 | 0 |  |
| 272-845 | 0 | 0 |  |
| 272-900 | 2,300 | 2,300 | 1.00 |
| 272-901 | 10,000 | 10,000 | 1.00 |
| 272-902 | 22,000 | 49,270 | 2.24 |
| 272-903 | 11,400 | 15,480 | 1.36 |
| 272-903A | 18,400 | 21,200 | 1.15 |
| 272-903B | 52,000 | 52,000 | 1.00 |
| 272-904 | 32,000 | 32,000 | 1.00 |
| 272-905 | 89,000 | 89,000 | 1.00 |
| 272-906 | 18,000 | 20,340 | 1.13 |
| 272-921 | 3,200 | 3,616 | 1.13 |
| 272-922 | 8,000 | 8,000 | 1.00 |
| 272-923 | 9,000 | 9,000 | 1.00 |
| 272-941 | 0 | 0 |  |
| 272-961 | 53,000 | 82,592 | 1.56 |
| 272-961A | 8,500 | 8,500 | 1.00 |
| 272-961B | 135,000 | 202,667 | 1.50 |
| 272-961C | 25,000 | 58,333 | 2.33 |
| 272-962 | 4 | 5 | 1.25 |
| 272-962A | 350 | 817 | 2.33 |
| 272-962B | 0 | 0 |  |
| 272-963 | 10,500 | 10,500 | 1.00 |
| Subtotal | 678,204 | 871,011 | 1.28 |

-Continued-

Appendix B.1. (page 2 of 3 )

| Stream No. | Numbers of Fish |  | ETE/PC ${ }^{\text {b }}$ |
| :---: | :---: | :---: | :---: |
|  | Peak Count (PC) | Estimated Total Escapement (ETE) |  |
| Central District |  |  |  |
| 272-201 | 210 | 210 | 1.00 |
| 272-202 | 0 | 0 |  |
| 272-202A | 600 | 903 | 1.50 |
| 272-202B | 200 | 200 | 1.00 |
| 272-204 | 16,600 | 24,460 | 1.47 |
| 272-205 | 5 | 5 | 1.00 |
| 272-206 | 230 | 230 | 1.00 |
| 272-302 | 45,500 | 45,500 | 1.00 |
| 272-501 | 63,000 | 89,499 | 1.42 |
| 272-502 | 0 | 0 |  |
| 272-502A | 0 | 0 |  |
| 272-503 | 0 | 0 |  |
| 272-504 | 0 | 0 |  |
| 272-505 | 0 | 0 |  |
| 272-506 | 300 | 300 | 1.00 |
| 272-507 | 0 | 0 |  |
| 272-508 | 0 | 0 |  |
| 272-509 | 7,300 | 7,300 | 1.00 |
| 272-510 | 1,350 | 1,350 | 1.00 |
| 272-511 | 0 | 0 |  |
| 272-511A | 500 | 500 | 1.00 |
| 272-511B | 0 | 0 |  |
| 272-512 | 0 | 0 |  |
| 272-514 | 23,000 | 25,497 | 1.11 |
| 272-516 | 19,000 | 19,093 | 1.00 |
| 272-602 | 1,700 | 2,273 | 1.34 |
| 272-604 | 200 | 226 | 1.13 |
| 272-605 | 5,000 | 5,658 | 1.13 |
| 272-606 | 1,000 | 1,800 | 1.00 |
| Subtotal | 186,495 | 225,004 | 1.20 |

## Chignik Bay District

| 271-100 | 0 | 0 |  |
| :---: | :---: | :---: | :---: |
| 271-101A | 3,000 | 3,000 | 1.00 |
| 271-101B | 4,000 | 4,000 | 1.00 |
| 271-102A | 0 | 0 |  |
| 271-102B | 0 | 0 |  |
| 271-102C | 0 | 0 |  |
| 271-103 | 250 | 250 | 1.00 |
| 271-104 | 4,800 | 4,800 | 1.00 |
| 271-105 | 400 | 400 | 1.00 |
| 271-106 | 1,020 | 1,020 | 1.00 |
| Subtotal | 13,470 | 13,470 | 1.00 |

[^3]Appendix B.1. (page 3 of 3 )

| Stream No. | Numbers of Fish |  | ETE/PC ${ }^{\text {b }}$ |
| :---: | :---: | :---: | :---: |
|  | Peak Count (PC) | Estimated Total Escapement (ETE) |  |
| Western District |  |  |  |
| 273-702 | 2,900 | 2,900 | 1.00 |
| 273-720 | 0 | 0 |  |
| 273-722 | 32,000 | 32,000 | 1.00 |
| 273-723 | 1,500 | 1,500 | 1.00 |
| 273-802 | 10,800 | 10,800 | 1.00 |
| 273-821 | 0 | 0 |  |
| 273-822 | 2 | 2 | 1.00 |
| 273-823 | 1,700 | 1,700 | 1.00 |
| 273-842 | 1,200 | 1,200 | 1.00 |
| 273-843 | 1,700 | 1,700 | 1.00 |
| 273-844 | 0 | 0 |  |
| 273-845 | 110 | 110 | 1.00 |
| 273-941 | 4,500 | 5,982 | 1.33 |
| Subtotal | 56,412 | 57,894 | 1.03 |
| Perryville District |  |  |  |
| 275-401 | 4,200 | 4,760 | 1.13 |
| 275-402 | 9,400 | 9,400 | 1.00 |
| 275-403 | 0 | 0 |  |
| 275-404 | 3,800 | 5,813 | 1.53 |
| 275-405 | 0 | 0 |  |
| 275-406 | 161,000 | 168,403 | 1.04 |
| 275-408 | 200 | 207 | 1.03 |
| 275-502 | 51,000 | 51,000 | 1.00 |
| 275-503 | 40 | 45 | 1.12 |
| 275-504 | 3,700 | 3,700 | 1.00 |
| 275-505 | 19,000 | 19,000 | 1.00 |
| 275-506 | 2,380 | 2,618 | 1.10 |
| 275-600 | 100 | 103 | 1.03 |
| 275-601 | 1,600 | 2,370 | 1.48 |
| Subtotal | 256,420 | 267,419 | 1.04 |
| Grand Total | 1,191,001 | 1,434,798 | 1.20 |

a Table adapted from Barrett (1990).
$b$ Estimated total escapements derived by the area-under-the-curve (AUC) method included fractions of whole numbers which the ETE/PC ratio reflects in several instances.
${ }^{\mathrm{c}}$ Factor used to expand peak counts to Estimated total escapement for the years 1963-1983.

Appendix B.2. Pink salmon aerial survey odd-year peak counts, Chignik Management Area, 1969-1987.

| Stream |  | Peak Count |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Name | No. | 1969 | 1971 | 1973 | 1975 | 1977 | 1979 | 1981 | 1983 | 1985 | 1987 | Mean | 1989 |
| Eastern District |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Aniakchak | 272-605 | 100 | 0 | 1.200 | 1,450 | 1,500 | 0 | 2,700 | 1,000 | 0 | 2,500 | 1,045 | 5,000 |
| Cape Agutka | 272-606 | 2,000 | 2,000 | 1,000 | 1,200 | 900 | 13,000 | 5,800 | 100 | 0 | 4,200 | 3,020 | 1,800 |
| Main Creek | 272-702 | 2,000 | 1,000 | 1,000 | 3,700 | 2,300 | 13,500 | 5,800 | 9,000 | 13.600 | 11,100 | 6,300 | 53,400 |
| Northeast | 272-703 | 4,500 | 2,000 | 1,100 | 350 | 5,000 | 7,000 | 3,300 | 2,600 | 9,000 | 5,500 | 4,035 | 19,000 |
| Yantarni | 272-721 | 1,000 | 0 | 300 | 100 | 1,700 | 14,000 | 13,500 | 3,600 | 12,000 | 13,000 | 5,920 | 11,000 |
| Ocean Beach | 272-801 | 4,000 | 0 | 450 | 400 | 1, 000 | 1,000 | 10,500 | 2,600 | 4,500 | 9,350 | 3,380 | 11,400 |
| Nakalilok | 272-804 | 6,000 | 1,000 | 500 | 2,500 | 2,100 | 12,000 | 6,000 | 4,800 | 15,000 | 1,400 | 5,130 | 10,650 |
| Chiginagak | 272-902 | 1,000 | 2,500 | 300 | 400 | 2,100 | 11,000 | 5.000 | 400 | 2,500 | 3,200 | 2,840 | 22,550 |
| Chiginagak R . | 272-903 | , 0 | , 0 | 150 | 2,500 | 1,400 | - 400 | 6,000 | 1,300 | 1,000 | 32,900 | 4,565 | 18,800 |
| Chiginagak-904 | 272-904 | 5,500 | 1.000 | 500 | 650 | 3,100 | 7,000 | 6,950 | 1,700 | 5,000 | 11,000 | 4,240 | 32,100 |
| Chiginagak-905 | 272-905 | 2.400 | 4,000 | 800 | 1,300 | 20,800 | 61,000 | 25,000 | 3,100 | 15,000 | 20,700 | 15,410 | 99,000 |
| Agripina | 272-961 | 2,500 | 6,000 | 4,200 | 2,050 | 3,200 | 23,500 | 13,400 | 5,000 | 30,000 | 1,000 | 9,085 | 135,000 |
| Glacier | 272-962 | 0 | 0 | 0 | 150 | 0 | 0 | 50 | 1,300 |  | 5,500 | 700 | 350 |
| Kilokak | 272-963 | 200 |  | 50 | 250 | 200 | 1, 100 | 300 | 300 |  | 9,000 | 1,140 | 11,000 |
| Central District |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Thompson Valley | 272-204 | 19,000 | 6,500 | 2,300 | 8,700 | 24,500 | 35,000 | 6,500 | 1, 520 | 0 |  | 11,558 | 16,600 |
| Hook Bay | 272-302 | 17,000 | 10,000 | 3,700 | 1,900 | 15,300 | 42,700 | 12,000 | 1. 200 | 2,000 | 8,050 | 11,285 | 47,100 |
| Cape Kumliun | 272-501 |  | 51,000 | 23,300 | 10,000 | 113,000 | 153,000 | 35,000 | 0 | 118,100 |  | 55,933 | 63,245 |
| Bear | 272-505 | 0 | 0 | 0 | 0 | 500 | 0 | 100 | 2,000 | 0 | 12 | 261 | 0 |
| Rudy's | 272-509 | 200 | 0 | 0 | 0 | 5,800 | 12,000 | 700 | , 0 | 0 | 0 | 2,062 | 7,300 |
| North Fork | 272-514 | 4,000 | 0 | 1,900 | 350 | 4,400 | 12,700 | 14,000 | 3,500 | 5,000 | 6,650 | 5,250 | 23,180 |
| Western District |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Coal Cape | 273-702 | 40,000 | 8,000 | 1,065 | 13,550 | 78,500 | 50,000 | 84,900 | 11,300 | 0 | 11,700 | 29,901 | 2,900 |
| Ivan | 273-722 | 255,000 | 130,000 | 35,000 | 73,100 | 236,000 | 85,000 | 80,000 | 12,200 | 20,000 | 12,800 | 93,910 | 38,700 |
| Foot | 273-802 | 14,000 | 30,000 | 7,000 | 8,700 | 13,000 | 9,600 | 10,000 | 1,200 | 5,000 | 5,340 | 10,384 | 12,920 |
| Spoon | 273-823 | 6,500 | 7,000 | 700 | 3,500 | 4,100 | 7,000 | 6,700 | 400 | 200 | 30 | 3,613 | 1,800 |
| Portage | 273-842 | 21,000 | 10,000 | 14,000 | 400 | 3,500 | 17,500 | 6,500 | 300 | 0 | 0 | 7,320 | 1, 200 |
| Seal Bay | 273-843 | 7,500 | 5,000 | 1,500 | 5,000 | 3,000 | 200 | 9,000 | 1,000 | 5,000 | 500 | 3,770 | 1,700 |
| Perryville District |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Kupreanof | 275-401 | 6,500 | 3,500 | 200 | 650 | 3,000 | 28,000 | 14,000 | 3,500 |  |  | 7,419 | 4,200 |
| Smokey Hollow | 275-402 | 0 | 0 | 200 | 50 | 1,500 | 600 | 0 | 200 | 150 | 1,700 | 440 | 9,400 |
| Wasco's | 275-404 | 4,000 | 3,000 | 0 | 800 | 1,400 | 2,000 | 0 | 2,000 | 250 | 27,800 | 4,125 | 3,800 |

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Appendix B.2. (page 2 of 2 )

| Stream | Peak Count |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Name No. | 1969 | 1971 | 1973 | 1975 | 1977 | 1979 | 1981 | 1983 | 1985 | 1987 | Mean | 1989 |
| Ivanof $275-406$ <br> Humpback $275-502$ | $\begin{array}{r} 105,000 \\ 59,000 \end{array}$ | $\begin{array}{r} 20,000 \\ 8,000 \end{array}$ | $\begin{array}{r} 18,100 \\ 5,000 \end{array}$ | $\begin{array}{r} 20,800 \\ 8,100 \end{array}$ | $\begin{aligned} & 51,800 \\ & 48,200 \end{aligned}$ | $\begin{aligned} & 89,000 \\ & 59,000 \end{aligned}$ | $\begin{aligned} & 18,000 \\ & 39,000 \end{aligned}$ | $\begin{aligned} & 32,000 \\ & 11,000 \end{aligned}$ | $\begin{array}{r} 155,000 \\ 25,000 \end{array}$ | $\begin{array}{r} 123,400 \\ 19,800 \end{array}$ | $\begin{aligned} & 63,310 \\ & 28,510 \end{aligned}$ | $\begin{array}{r} 166,000 \\ 51,000 \end{array}$ |
| Total Peak Count ${ }^{\text {d }}$ (TPC) | 589,900 | 311,500 | 125,515 | 172,600 | 652,800 | 767,800 | 440,700 | 119,120 | 325,200 | 466,232 | 405,861 | 882,095 |
| ```Total Escapement }\mp@subsup{}{}{\textrm{b}}\mathrm{ (TE) Peak``` | 624,450 | 324,800 | 158,970 | 238,100 | 749,800 | 858,800 | 598.200 | 158,900 |  |  |  |  |
| TPC/TE ${ }^{\text {c }}$ | 94\% | 96\% | 79\% | 72\% | 87\% | 89\% | 74\% | 75\% |  |  |  |  |

${ }^{4}$ Total peak counts for 31 Chigink Management Area index streams.
b Total escapement for all aerial surveyed streams obtained from Thompson and Owen (1991).
${ }^{\text {c }}$ Average of TPC/TE for $1969-1983$ is $83.25 \%$.


[^0]:    ${ }^{\text {a }}$ Return is from brood year +2 .
    $\mathrm{b}^{\mathrm{b}}$ Excluding 1989.

[^1]:    -Continued-

[^2]:    -Continued-

[^3]:    -Continued-

